

Diel and seasonal vertical migrations of high-latitude zooplankton: knowledge gaps and
a high-resolution bridge

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dedicated to millions of zooplankton who sacrificed their lives for the advancement of science during the Billefjorden field campaign (2008-2009), the data of which fuels an integral part of this dissertation

Preface

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List of abbreviations

1D	-	One-dimensional
2D	-	Two-dimensional
ADCP	-	Acoustic Doppler Current Profiler
AZFP	-	Acoustic Zooplankton and Fish Profiler
CPU	-	Central Processing Unit
D	-	Key discovery about vertical migration
DVM	-	Diel vertical migration
ES	-	Echosounder
GA	-	Genetic Algorithm
HAS	-	Hypothesis about adaptive significance
HPC	-	Hypothesis about proximate control
LX	-	Laplace Crossover
MPTM	-	Makinen, Periaux and Toivanen mutation
R	-	Selected major review
RCGA	-	Real-Coded Genetic Algorithm
RAM	-	Random Access Memory
SVM	-	Seasonal vertical migration

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List of papers

- Paper I** **Bandara K**, Varpe Ø, Søreide JE, Wallenschus J, Berge J, Eiane K (2016) Seasonal vertical strategies in a high-Arctic coastal zooplankton community. *Marine Ecology Progress Series* 555:49-64
- Paper II** **Bandara K**, Varpe Ø, Ji R, Eiane K (2018) A high-resolution modeling study on diel and seasonal vertical migrations of high-latitude copepods. *Ecological Modelling* 368C:357-376
- Paper III** **Bandara K**, Varpe Ø, Ji R, Eiane K Artificial evolution of behavioral and life history strategies of Arctic *Calanus* spp. in response to bottom-up and top-down selection pressures. Manuscript

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Abstract

Vertical migration is a widespread behavior among zooplankton. Based on the periodicity, two types of vertical migrations are described for high-latitude marine zooplankton: the diel vertical migration (DVM, periodicity < 24 h) and seasonal vertical migration (SVM, periodicity ca. 1 year). Despite ca. 200 years of research, questions still remain about the influence of environmental variability on vertical migrations, and the integrative effect of DVM and SVM on life history strategies and fitness of high-latitude zooplankton. As a first step toward addressing these knowledge gaps, we used year-round zooplankton net data collected in a high-Arctic fjord to describe seasonal vertical distributions of the dominant herbivore and carnivore taxa in association with the environmental variability. Results suggested that SVM of most carnivorous zooplankton studied was likely reflecting a tracking of the seasonal pattern of depth distributions in their herbivore prey. However, SVM of herbivore zooplankton could not be explained by any of the environmental variables used, probably due to the mismatch between the coarser spatio-temporal resolution of the deployed nets and finer spatio-temporal scales at which their vertical migrations occur. To study vertical migrations of high-latitude zooplankton and their environmental relations in higher spatio-temporal resolution, we developed a new computationally efficient strategy-oriented modeling framework. This model could simulate DVM and SVM throughout the entire life cycle of herbivorous zooplankton. Testing of this model, by simulating a generalized herbivorous copepod suggested that DVM and SVM can have profound and largely different effects on fitness and phenology. Hence these migratory behaviors should be studied in concert. Further, by developing the above modeling framework to allow for more species-specific parameterization, we found that top-down selection pressures mediated by the environment (i.e. predation risk) can have an overwhelming influence on the vertical migratory and life history strategies of Arctic herbivorous copepods (i.e. *Calanus* spp.). Altogether, these findings suggest that studying vertical migrations in higher spatial, temporal and biological resolution may contribute toward bridging the gaps of our understanding of the subject matter.

1. Introduction

1.1 Migration

While the interactions between organisms and their environment forms the foundation of Ecology, the mechanisms by which these interactions occur are studied as animal behavior. One particular case of animal behavior clearly sets apart from the rest in such a way that it involves movement of animals typically from one habitat to another, and is termed migration ([Aidley 1981](#)).

In an ecological perspective, migration is a widespread behavior that involves persistent and active movement of individuals often caused by spatial and temporal variation in resources and risks ([Cresswell et al. 2011](#)). Animals can migrate varying distances in search of better feeding (e.g. [L'Abée-Lund & Vøllestad 1987](#), [Williamson et al. 1988](#), [Levey & Stiles 1992](#)), survival (e.g. [Werner et al. 1983](#), [Hebblewhite & Merrill 2007](#), [McKinnon et al. 2010](#)) and reproductive opportunities (e.g. [Hardy & Raymond 1980](#), [Smith & Moore 2005](#), [van Ginneken & Maes 2005](#)). For example, among birds, the annual migration of North American blue grouse (*Dendragapus obscurus*) ranges up to ca. 800–3000 body lengths (0.3–1 km, [Hoffmann 1956](#), [Cade & Hoffman 1993](#)), while the annual transpolar migration of the Arctic tern (*Sterna paradisaea*) can extend up to ca. 2.5×10^8 body lengths ([ca. 80000 km](#), [Egevang et al. 2010](#)). Irrespective of the distance travelled, migratory individuals will possess enhanced fitness compared to residents, given that the benefits of the migration overwhelm the costs ([Aidley 1981](#)). As a result, migration is seen as an adaptive strategy that enhances fitness.

1.2 Zooplankton migration: the concept of vertical migration

Migration is a common phenomenon in the realm of zooplankton. However, about a century ago, it was believed that zooplankton can only passively drift with water currents, but don't have a pronounced means of migration. Despite this early belief, it is well known today that they can actively migrate in the water column. Zooplankton

migrations are manifested through individual swimming movements brought about by rhythmical beating of flagella and cilia (e.g. [Keller & Rubinow 1976](#), [Matsumoto 1991](#)), antennae, maxillae and pleopods (e.g. [Strickler 1977](#), [Kils 1983](#)), or using fins and other modified muscular structures (e.g. [Satterlie et al. 1985](#), [Jordan 1992](#)).

On a timescale ranging from a few seconds to minutes, most zooplankton swimming trajectories span over a three-dimensional space and are often described using diffusive motility models (e.g. [Bundy et al. 1993](#), [Schmitt & Seuront 2001](#)). On a broader timescale (> 12 h), zooplankton vertical movements tend to be more pronounced compared to those in the horizontal dimension, and can span from a few meters in shallow freshwater bodies ([Cerbin et al. 2003](#)) to several kilometers in the open ocean ([van Haren & Compton 2013](#)). These vertical migrations are common among a wide range of zooplankton taxa, and probably represent the largest animal migration on earth in terms of biomass ([Hays 2003](#), [Williamson et al. 2011](#)).

Vertical migrations are viewed as a type of commuting between shallow foraging habitats and a deep-water refugia ([Dingle & Drake 2007](#)). Depending on the frequency of commuting (periodicity), vertical migrations can be broadly categorized into diel and seasonal components. The shorter-term diel vertical migration (DVM) refers to a daily commuting of zooplankton between the foraging and refugial habitats, where the longer-term seasonal vertical migration (SVM) refers to an annual commuting. Despite the generality of this view, it has been used as the foundation of numerous theoretical studies on zooplankton vertical migrations.

1.3 Diel vertical migration of zooplankton

1.3.1 Brief history and present understanding

Historical observations of DVM dates ca. two centuries back to the records of Georges Cuvier ([Cuvier 1817](#)). These early observations were centered on the daily appearance and disappearance of microscopic crustaceans (the term plankton was not defined by

then) from the upper water layers of freshwater lakes. However, phenomenon of a vertical migration remained doubted until late 1800s, where [Fuchs \(1882\)](#) used depth-stratified net samples and described the differential day and night vertical distributions of numerous lifeforms including crustaceans. Fuchs's pioneering work echoed in numerous subsequent DVM studies on a wide range of taxonomic groups in both freshwater and marine realms. Consequently, by early 1900s, the term diel vertical migration was firmly established in the literature ([see Russell 1927](#)), and is currently understood as a geographically and taxonomically widespread behavior of zooplankton.

At present, two patterns of DVM are well known: the nocturnal DVM and the reverse DVM ([reviewed in Bayly 1986](#)). The nocturnal DVM, first documented by [Cuvier \(1817\)](#), is the most common diel vertical migratory behavior among zooplankton. It is characterized by an ascent from depths in the dusk and occupation of shallower waters during the night time, and a descent in the dawn and occupation of depths during the daytime (Fig. 1B). Early observations of what is now known as the reverse DVM dates back to [Brook \(1886\)](#), but the term was coined and brought into prominence following the work of [Ohman et al. \(1983\)](#). As the name implies, reverse DVM is the opposite of classic DVM, where zooplankton tend to occupy shallower waters during the daytime and deeper waters during the night (Fig. 1C).

1.3.2 Hypotheses about proximate control

Up to the mid-1900s, most DVM research was centered on describing its proximate control (how migration occurs, Fig. 2). Most such hypotheses on the proximate control of classic DVM were inspired by the observation that diel variability of zooplankton vertical behavior appear to correlate with those of irradiance and temperature ([Russell 1927](#)). It was thus suggested that zooplankton remain at depths during the daytime and migrate to near-surface waters during the night to avoid intense levels of irradiance or temperature (e.g. [Cuvier 1817](#), [Parker 1902](#), [Michael 1911](#), [Clarke 1934](#)). The influence of irradiance on the proximate control of DVM was subjected to extensive field and laboratory testing especially during the first half of the 20th century (reviewed in [Russell](#)

[1927](#), [Cushing 1951](#)). Consequently, by mid 1920s the relative change of irradiance during the diel cycle was considered as the main cue for zooplankton DVM ([Rose 1925](#)).

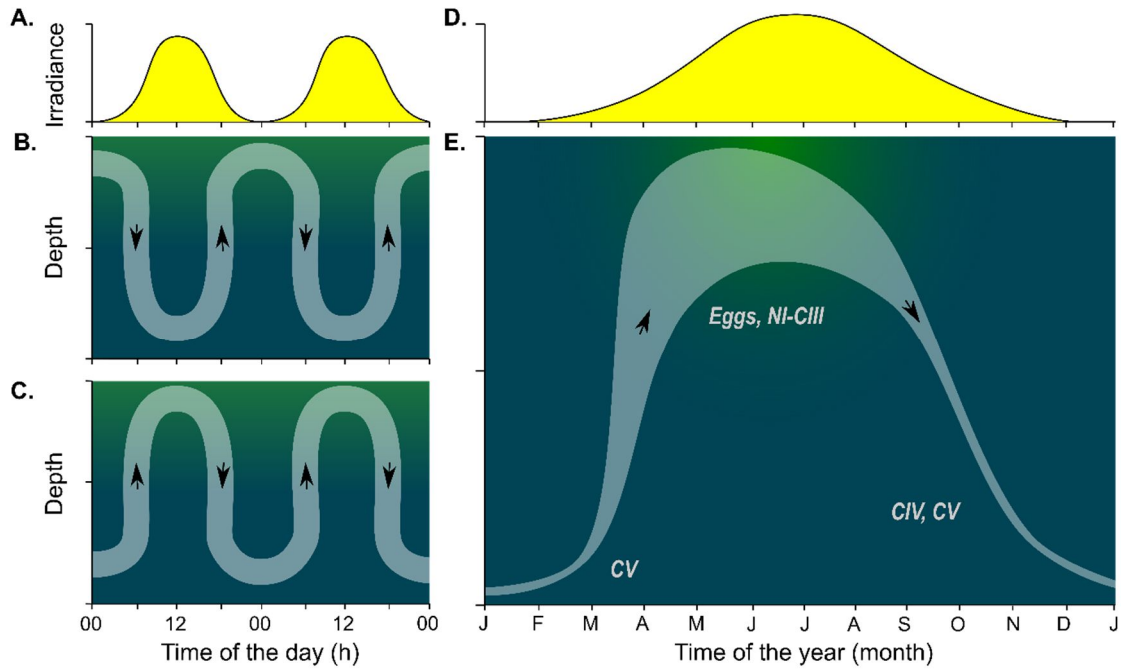


Fig. 1 Simplified conceptual models for DVM (A–C) and SVM (D, E). In nocturnal DVM, zooplankton tend to occupy shallower waters during the night and descends to depth during the daytime (B). The opposite happens in reverse DVM (C). Opaque white line represents the vertical trajectory of the population, and its thickness indicate the relative population size. Arrowheads point to the general direction of the vertical trajectory. Green color indicates productive vertical regions of the water column. Abscissae not to scale.

One main weakness of the irradiance intensity hypothesis is its inability to explain the reverse DVM. Although reverse DVM was iteratively reported following [Brook \(1886\)](#) (e.g. [Herdman 1907](#), [Tattersall 1911](#)), these field observations did not receive a prominent attention and were sometimes criticized as sampling artifacts. Hypotheses-

1817	D First written account on nocturnal DVM HPC Temperature	Cuvier (1817)
1874	HPC Irradiance (absolute light intensity)	Weismann (1874)
1878	HPC Visual predation	Forel (1878)
1882	D First comprehensive field evidence to support nocturnal DVM HPC Food	Fuchs (1882)
1886	D First written account on reverse DVM	Brook (1886)
1887	D Term 'plankton' is defined	Hensen (1887)
1893	D Comprehensive theory on irradiance influence on DVM	Loeb (1893)
1902	D Term 'geotrophism' is defined for plankton migrations HPC CO ₂ concentrations and viscosity	Parker (1902) Ostwald (1902)
1909	HPC Irradiance (relative light intensity)	Moore (1909)
1910	D Doubts DVM is a sampling artifact due to net-avoidance	Franz (1910)
1911	D First comprehensive experiments to support nocturnal DVM	Michael (1911)
1912	HPC UV radiation	Ewald (1912), Moore (1912)
1917	HPC Endogenous mechanisms	Esterly (1917)
1923	HPC Specific gravity	Eyden (1923)
1924	HPC Irradiance (optimum light intensity)	Rose (1925)
1927	R 1 st review (on vertical distributions)	Russell (1927)
1931	D Suggests DVM is an energy-wasting maladaptation	Worthington (1931)
1935	D Suggests carnivore DVM is a result of following herbivore prey	Hardy and Gunther (1935)
1936	HAS Resource-related hypothesis	Hardy (1936)
1951	R 2 nd review (dedicated review on DVM)	Cushing (1951)
1963	HAS Metabolic advantage hypothesis	McLaren (1963)
1964	R 3 rd review (on vertical distributions)	Basnse (1964)
1967	HPC Resurgence of visual predation hypothesis	Hutchinson (1967)
1974	HAS Demographic advantage hypothesis	McLaren (1974)
1976	HAS Predator-evasion hypothesis	Zaret and Suffern (1976)
1977	HAS Further development of metabolic advantage hypothesis	Enright (1977)
1983	D First comprehensive field evidence to support reverse DVM	Ohman <i>et al.</i> (1983)
1986	R 4 th review (dedicated review on DVM) D First field evidence to support the synchrony of DVM with lunar cycle	Bavly (1986) Gliwicz (1986)
1989	R 5 th review (on adaptive significance)	Lampert (1989)
2003	HAS Hunger-satiation hypothesis R 6 th & 7 th reviews (latter on adaptive significance)	Pearre (2003) Pearre (2003), Hays (2003)
2009	R 8 th review (on proximate control) D Detects DVM during the polar night in high-Arctic R 9 th review (dedicated review on DVM)	Cohen and Forward (2009) Berge <i>et al.</i> (2009) Ringelberg (2009)
2013	D Detects deep DVM down to ~1.6 km in open ocean HPC Irradiance (hypersensitivity to irradiance)	van Haren and Compton (2013) Båtnes <i>et al.</i> (2013)
2014	R 10 th review (on modern perspectives on DVM)	Brierly (2014)
2015	HPC Further support to zooplankton hypersensitivity to irradiance	Cohen <i>et al.</i> (2015)
2016	D Detects large scale DVM during the Arctic polar night	Last <i>et al.</i> (2016)

Fig. 2 Key discoveries (D) related to DVM of zooplankton that led to or has the potential of leading to hypotheses about proximate control (HPC) and hypotheses about adaptive significance (HAS), along with some selected major reviews (R). The list includes literature related to both freshwater and marine environments.

-that could provide a plausible explanation to both variants of DVM remained buried in the literature, perhaps overshadowed by extensive focus on irradiance intensity hypothesis. These include the early thoughts of [Forel \(1878\)](#) that avoidance of sun-lit waters by many pelagic crustaceans is due to the presence of pelagic fish, and [Fuchs \(1882\)](#) that DVM could be a feeding migration. These food- and predator-related control mechanisms of DVM did not receive a prominent scientific attention until the latter half of the 20th century ([see below, and also Hutchinson 1967](#)).

In the early 1900s, it was argued if DVM of zooplankton is behavior that minimizes exposure to harmful UV radiation (e.g. [Ewald 1912](#), [Moore 1912](#)). However, this perspective was quickly criticized based on the observations on deep-water DVM (where ascent migration doesn't extend up to the surface) and reverse DVM. Few years later, [Esterly \(1917\)](#) observed persistent DVM behavior in copepods kept under continuous darkness, and hypothesized that DVM is regulated by an endogenous mechanism. However, even to date, evidence supporting this view remains elusive. Apart from these, salinity and density (e.g. [Esterly 1919](#), [Eyden 1923](#)), and dissolved gases (e.g. [Ostwald 1902](#)) have also been viewed as proximate cues for DVM, but with little supporting evidence.

1.3.3 Hypotheses about adaptive significance

From mid-1960s, the focus on DVM research began to shift towards explaining its adaptive significance (also termed as ultimate causes, i.e. why migration occurs, Fig. 2). Pioneering work was done by Ian A. McLaren following his theoretical modeling work ([McLaren 1963](#)), possibly inspired by the earlier insights of George Evelyn Hutchinson ([Hutchinson 1959](#)). The field investigations, laboratory experiments, and modeling work conducted thenceforth have led to four well-established hypotheses regarding the adaptive significance of DVM (Table 1).

Table 1 Main hypotheses about the adaptive significance of DVM. Several additional hypotheses are reviewed in [Lampert \(1989\)](#) and [Pearre \(2003\)](#), but are not mentioned here due to its relatively less significance

Hypothesis	General description	Pioneering work
Metabolic advantage	Feeding in warm, food-rich surface waters at night and spending rest of the day in colder deeper waters possesses an energetic advantage.	McLaren (1963) Enright (1977)
Demographic advantage	Alike above, but posits that spending part of the day in colder deeper waters elevates the size at sexual maturity and ends up producing more offspring.	McLaren (1974)
Predator avoidance	Classic DVM: Zooplankton feed in food-rich surface waters during nighttime when visually orientating predators are least effective, and take refuge in deeper darker waters during the daytime.	Zaret and Suffern (1976)
	Reverse DVM: The above behavior is reversed in the case of non-visual (tactile) predators	Ohman et al. (1983)
Hunger/satiation	The daily ascent and descent migrations are driven by the nutritional state of the zooplankter. Where, hunger drives ascent migration and satiation causes descent (active downward swimming or passive sinking).	Pearre (2003)

Among hypotheses of adaptive significance, the predator avoidance hypothesis seems to be the most widely received: first, given the broad range of empirical evidence supporting it, and second, its ability to explain both the classic and the reverse variants of DVM ([Bayly 1986](#), [Lampert 1989](#), [Hays 2003](#)). Therefore, in a broader perspective, the diel vertical migration is generally seen as a strategy that trades off growth potential to

survival, by minimizing the time spent near food-rich surface waters during the time of the day that is most susceptible to visual ([Loose & Dawidowicz 1994](#)) or tactile predators ([Ohman et al. 1983](#), [Ohman 1990](#)).

1.4 Seasonal vertical migration of zooplankton

1.4.1 Brief history and present understanding

In late 1800s, the seasonal appearance and disappearance of pelagic animals puzzled the marine scientists working in northerly latitudes (e.g. [Schmidtlein 1879](#)). Following the methods of [Fuchs \(1882\)](#), [Chun \(1888\)](#) used depth-stratified net samples and showed that the summertime disappearance of jellyfish and crustaceans from surface waters was due to their migration to depths exceeding 1000 m in the Mediterranean Sea. Although not geographically or taxonomically widespread, this behavior is common to many high-latitude zooplankton, and termed seasonal vertical migration (SVM).

SVM is generally characterized by a descent to deeper waters during summer and autumn, and ascent to near-surface waters by the spring of the following year (Fig. 1E). This pattern is common among most high-latitude herbivore zooplankton (reviewed in [Conover 1988](#), [Falk-Petersen et al. 2009](#), [Varpe 2012](#)), and carnivores that seasonally follow their herbivore prey ([Hagen 1999](#)). However, among omnivores, SVM may either be absent ([Lischka & Hagen 2005](#)) or sometimes reversed ([Bandara 2014](#)). In most cases, as the seasonal vertical distribution changes with the developmental stage (Fig. 1E) this is also termed as the ontogenetic vertical migration ([Russell 1927](#), [Banse 1964](#), [Conover 1988](#), [Falk-Petersen et al. 2009](#)).

1.4.2 Hypotheses about proximate control

Hypotheses about proximate cues of SVM emerged since the time of its first introduction by [Chun \(1888\)](#), who suggested that warming of the surface waters in the

summer seasonally drives the pelagic species to deeper waters (Fig. 3). [Birge \(1904\)](#) considered both temperature and irradiance, and suggested the latter as a strong cue for the SVM of freshwater zooplankton and former is just a condition. In the marine realm however, the influence of irradiance on the SVM remained questionable, as the depth of the migration tend to extend well below the photic zone. However, by the end of the second decade of the 20th century, it was evident that irradiance could penetrate ocean depths greater than it was previously thought ([Klugh 1925](#), [Poole 1925](#), [Atkins 1926](#), [Poole & Atkins 1926](#)). This likely have motivated the bloom of literature supporting the hypothesis regarding irradiance as a proximate cue for the SVM of marine herbivorous copepods (e.g. [Bigelow 1926](#), [Russell 1926](#), [Nicholls 1933](#), [Sømme 1934](#), [Ussing 1938](#)).

Alternative hypotheses on the proximate control of SVM began to surface by mid-1900s (Fig. 3). Based on the deep SVM depths (> 1000 m) of copepods in the Norwegian Sea, [Østvedt \(1955\)](#) doubted if neither irradiance nor temperature could act as proximate cues. Instead, he was the first to suggest endogenous regulation (i.e. endocrinal changes related with gonadal development) as a proximate cue for SVM. Despite the detailed studies conducted in the mid-20th century on endocrinal regulation of SVM (e.g. [Carlisle & Pitman 1961](#), [Harris 1963](#)), conclusive evidence still remains elusive. A different endogenous regulation mechanism was suggested in In late 20th century, where it was hypothesized that the timing of SVM is regulated by an internal timer or a biological clock ([Grigg & Bardwell 1982](#), [Miller et al. 1991](#), [Hirche 1996a](#)). However, existence of such an endogenous timer is yet to be proven. Following a different line of reasoning, [Kaartvedt \(1996\)](#) suggested that the-timing and the depth dynamics of SVM may be regulated by the mortality risk imposed by planktivorous fish.

1878	D First written account on seasonal patterns of vertical distributions	Schmidtlie (1878)
1887	D Term 'plankton' is defined	Hensen (1887)
1888	D Discovery of SVM	Chun (1888)
1902	HPC Temperature	Gran (1902)
1904	HAS Food availability hypothesis	Birge (1904)
1926	HPC Irradiance	
1926	D First comprehensive field evidence to support irradiance influence on SVM	Russell (1926)
1927	R 1 st review (on vertical distributions)	Russell (1927)
1934	D First to describe overwintering among marine zooplankton	Sømme (1934)
1937	HAS Horizontal transport hypothesis	Mackintosh (1937)
1951	D Detects SVM down to a depth of ca. 2 km	Bertelsen (1951)
1955	R 2 nd review (on vertical distributions)	Bogorov and Vinogradov (1955)
1958	HPC Endogenous mechanism (endocrinal)	Østvedt (1955)
1958	D Suggests SVM of carnivores is a consequence of following herbivores	David (1958)
1959	D Reports SVM down to a depth of ca. 4 km	Vinogradov (1959)
1959	R 3 rd review (on vertical distributions)	
1961	HPC Experimental evidence for endocrinal control of SVM	Carstle and Pitman (1961)
1961	D Suggests SVM are uncoordinated ontogenetic vertical drifts	Bainbridge (1961)
1962	D First to emphasize the significance of SVM in vertical Carbon transport	Beyer (1962)
1962	D Experimental evidence for overwintering physiology of marine copepods	Conover (1962)
1962	D Suggests breeding modes may alter timing of SVM in herbivores	Heinrich (1962)
1964	R 4 th review (on vertical distributions)	Banse (1964)
1978	D Comparative study on diapause of copepods and insects	Elgmork and Nilsen (1978)
1983	HPC Food	Herman (1983)
1985	HPC Field evidence to support the influence of food on SVM	Head and Harris (1985)
1988	R 5 th review (on life history strategies)	Conover (1988)
1991	R 6 th review (on proximate control)	Miller <i>et al.</i> (1991)
1995	R 7 th review (on dormancy)	Dahms (1995)
1996	R 8 th & 9 th reviews (on diapause and reproductive strategies)	Hirche (1996a, b)
1996	HPC Visual predation	Kaartvedt (1996)
1999	HAS Lipids and buoyancy hypothesis	Visser and Joansdottir (1999)
1999	R 9 th review (on seasonal strategies)	Hagen (1999)
2000	HAS Predator avoidance hypothesis	Kaartvedt (2000)
2004	HAS Further development of lipids and buoyancy hypothesis	Irigoien (2004)
2009	R 10 th review (on overwintering, lipids and life history strategies)	Falk-Petersen <i>et al.</i> (2009)
2010	HAS Further development of predator avoidance hypothesis	Varpe and Fiksen (2010)
2012	R 11 th review (on seasonal strategies)	Varpe (2012)
2012	HAS Further development of horizontal transport hypothesis	Berge <i>et al.</i> (2012)

Fig. 3 Key discoveries (D) related to SVM of zooplankton that led to or has the potential of leading to hypotheses about proximate control (HPC) and hypotheses about adaptive significance (HAS), along with some selected major reviews (R).

Table 2 Main hypotheses that explain the adaptive significance of SVM.

Hypothesis	General description	Pioneering work
Food availability	SVM reflects adaptation to the seasonal changes of food availability, i.e. descent to depths for overwintering during the unproductive season and ascent to shallow productive waters at the onset of the productive season.	Gran (1902) Sømme (1934)
Predator avoidance	Generation length and hence the timing of the SVM is tuned to avoid seasonal peaks of visual predation risk, and depth of the SVM is determined by those with least visual predation risk.	Kartvedt (1996) Kartvedt (2000)
Lipids and buoyancy	Seasonal lipid stores play a central role in SVM by influencing the timing of the seasonal migration and the overwintering depth. This is controlled by the density of the stored lipids and the buoyancy effect it produces. The size of the lipid store is (locally) adapted to sink to a depth below the convective mixed layer during overwintering.	Visser and Jonasdottir (1999) Irigoien (2004)
Horizontal transport	Timing and depth of SVM allows zooplankton to be transported by water currents to regions with better feeding opportunities.	Mackintosh (1937)

1.4.3 Hypotheses about adaptive significance

Several key alternative views on adaptive significance of SVM are listed in Table 2. The food availability hypothesis seems to be more general and explains the seasonal migration patterns of zooplankton with different feeding strategies. Pioneering work was done by [Gran \(1902\)](#), who suggested that high-latitude herbivorous zooplankton descends to great depths to overwinter during the unproductive part of the year, and

ascend back to shallow waters as the primary production starts. However, [Heinrich \(1962\)](#) suggested that the timing of SVM in herbivorous zooplankton may be altered by different reproductive strategies (i.e. what is now known as capital and income breeding). [David \(1958\)](#) noted that the SVM of carnivorous zooplankton is a consequence of following their seasonally migrating herbivore-prey. SVM of omnivorous zooplankton seems to be less pronounced, and usually related to the year-round abundance of food (e.g. [Richter 1995](#), [Falkenhaus et al. 1997](#), [Lischka & Hagen 2005](#), [Darnis & Fortier 2014](#)). Compared to the food availability hypothesis, the rest are less well supported by literature. However, despite the generality of the food availability hypothesis, it is not clear if it is widely accepted among planktologists.

1.5 Knowledge gaps

Despite ca. 200 years of research, there are still gaps in our understanding about the causes and consequences of zooplankton vertical migrations. This knowledge gap appears to be wider regarding SVM compared to DVM. First, the proximate cues of SVM are poorly understood, mainly because there is often a spatial (i.e. vertical) mismatch between the depths of SVM and that at which the seasonal dynamics of physical environmental variables usually occur ([Østvedt 1955](#)). For example, it is commonly questioned that how a zooplankton overwintering several hundred meters would perceive spring–summer increase of irradiance or temperatures which are restricted to the shallow waters ([Banse 1964](#)). Alternative hypotheses, such as endogenous mechanisms, lipid level changes and predation risk appear to theoretically address these problems (Fig. 3), but are not satisfactorily supported by empirical evidence. Second, unlike DVM, there is no widely accepted hypothesis about the adaptive significance of SVM. This is because the timing and amplitude (the vertical extent) of SVM shows a profound variability between geographic locations (e.g. [Russell 1927](#), [Sømme 1934](#), [Wiborg 1954](#), [Daase et al. 2013](#), [Melle et al. 2014](#)), between different years within the same location (e.g. [Kosobokova 1999](#), [Pertsova & Kosobokova 2003](#)), between species (e.g. [Hirche 1991](#), [Unstad & Tande 1991](#), [Madsen et al. 2001](#), [Lischka & Hagen 2005](#),

[Hirche & Kosobokova 2011](#)) and between individuals of the same species ([Pearre 1979](#), [Pedersen et al. 1995](#)). However, none of the proposed hypotheses (Table 2) are capable providing a plausible explanation to this variability.

Despite extensive theoretical advancements on DVM, recent observations on high-amplitude DVM ([van Haren & Compton 2013](#)) and mid-winter DVM under the darkness of Arctic polar night ([Berge et al. 2009](#), [Hobbs 2016](#), [Last et al. 2016](#)) have challenged the current understanding of its proximate control. As with SVM, the question raised here is: how is it possible that animals residing at depths where irradiance-levels are extremely low, or in periods of the year with no sunlight use it as a proximate cue to perform DVM? A satisfactory explanation was provided to this problem in 2015, where detailed investigations of copepod and euphausiid zooplankton optical sensory mechanisms have revealed that zooplankton eyes are hypersensitive to irradiance, thus animals can show phototactic reactions to background irradiance sourced from moon, stars and aurora borealis ([Båtnes et al. 2015](#), [Cohen et al. 2015](#)). Despite these explanations, the adaptive significance of mid-winter vertical migrations remains questioned. This is because most polar night diel migrants are predominantly herbivorous copepods or euphausiids ([Berge et al. 2009](#), [Daase et al. 2014](#), [Błachowiak-Samołyk et al. 2015](#), [Last et al. 2016](#)), which are supposed to be in diapause in this time of the year. Although the reasons driving these migrations are not yet fully known, these motivate interactive research on DVM, SVM and overwintering strategies of high-latitude zooplankton.

2. Objectives

The main aim of this study is to improve our understanding of the influence of environmental variability on the diel and seasonal vertical migrations of high-latitude zooplankton, and to investigate the combined influence of DVM and SVM on their life history strategies and fitness. Here,

- (i). Paper-I is aimed at investigating the influence of seasonal environmental variability on SVM of high-latitude zooplankton using field data,
- (ii). Paper-II is aimed at addressing the key limitations of paper-I and developing a high-resolution modeling framework that allows studying the integrative effects DVM and SVM on the life history strategies and fitness of high-latitude zooplankton and,
- (iii). Paper-III is aimed at further developing the above model to allow a detailed investigation that analyzes the relative contributions of bottom-up and top-down environmental variables (selection pressures) in shaping up of behavioral and life history strategies of several selected high-latitude zooplankton taxa with different body sizes, reproductive strategies and generation times.

3. General discussion

3.1 Paper-I: background and main findings

In Paper-I ([Bandara et al. 2016](#)), we attempted to describe the influences of seasonal environmental variability on zooplankton seasonal vertical migrations (SVM), using depth-stratified zooplankton abundance data collected in a high-Arctic fjord (Billefjorden, Svalbard, 78.40°N). The selection of this location was motivated by the limited advective influence ([Nilsen et al. 2008](#)), so that zooplankton vertical distributions could be interpreted largely as reflecting local behavior. The data set covered almost an annual cycle, including the winter months. Data were collected on a bi-weekly to monthly basis (i.e. 14–30 d temporal resolution) using a net with a relatively large mesh width, which provided an year-round time series of zooplankton vertical distributions across a broad range of taxonomic groups, including larger developmental stages of copepods, chaetognaths, hydromedusae and ctenophores. We interpreted the seasonal vertical distributions of these taxa as patterns of SVM, and related those to local environmental variability, and with each other through identifiable predator-prey interactions, which were termed seasonal vertical strategies.

According to our findings, it appeared that seasonal vertical strategies of most zooplankton taxa were linked through trophic interactions. We found statistically significant correlations between the vertical distributions of predators and prey taxa, which could be traced up to three trophic levels. Further, estimated correlation coefficients were highly significant for specific predators and prey. Therefore, we concluded that vertical strategies of predatory zooplankton were driven by those of their prey. However, we could not find any meaningful statistical associations between the vertical distributions of herbivore zooplankton (i.e. *Calanus finmarchicus*, *C. glacialis* and *C. hyperboreus*) and biotic or abiotic environmental parameters. All *Calanus* species descended to deeper waters before the pelagic bloom had terminated, and ascended back to near-surface waters in mid-winter. Although the reason behind this strong decoupling of herbivore SVM and food availability could not be derived from our data, this finding aligned with recent observations of *Calanus* spp. occurring closer to surface waters during the Arctic polar night ([Daase et al. 2014](#), [Błachowiak-Samołyk et al. 2015](#)).

3.2 Paper-I: key limitations and motivation for Paper-II

Although Paper-I provided one of the best year-round evidence about how vertical migratory behavior can propagate through trophic levels, it was not an essentially novel observation. [David \(1958\)](#) was likely the first to suggest that Antarctic chaetognaths may seasonally follow their copepod prey to deeper waters. Similar seasonal prey-following strategies have been observed for numerous pelagic invertebrate (e.g. [Torres et al. 1994](#), [Kaartvedt et al. 2002](#)) and vertebrate predators (e.g. [Sims et al. 2003](#), [Born et al. 2004](#), [Laidre et al. 2007](#), [Geoffroy et al. 2011](#)).

Despite year-round coverage of the analyzed data, the means by which SVM of predominantly herbivorous *Calanus* species were controlled remained elusive. Two main factors could have affected this. First, our study did not focus on the seasonality of visual predation risk, which is a major influential factor for SVM ([Kaartvedt 1996](#), [Kaartvedt 2000](#), [Varpe & Fiksen 2010](#)). Second, the spatio-temporal scales at which the vertical migration had occurred in the fjord may have been finer than the coarse spatial (ca. 50 m vertical) and temporal (ca. 15–30 d) resolution employed in our study (e.g. [Pearre 1979](#), [Skjoldal et al. 2000](#)). However, given the logistic challenges of sampling at high-latitudes, the year-round bi-weekly depth-stratified zooplankton data used in our study represent some of the best available zooplankton net samples yet.

Limitations of Paper-I provided the motivation to look for a sampling gear that could allow studying zooplankton vertical migrations in superior resolution. Beside plankton nets, there are numerous alternative techniques that facilitate *in-situ* observation of zooplankton populations in a superior spatio-temporal resolution ([reviewed in Sameoto et al. 2000](#)). However, zooplankton nets possess the advantage of having superior biological resolution (i.e. zooplankton can be identified to species and their developmental stages), which many of these novel techniques do not (Table 3). This suggest that all three aspects of resolution (i.e. spatial, temporal and biological) cannot be simultaneously maximized without compromising each other. However, the lack of a ‘universal plankton sampler’ that can optimize these resolution trade-offs presents the opportunity for developing a methodological framework for studying zooplankton vertical migrations in higher resolution.

Table 3: Some popular plankton sampling methods and their associated spatial (i.e. vertical), temporal and biological (i.e. species or developmental stage) resolution. Long-term deployment refers to the capability to deploy these devices affixed to ship-hulls or moorings to make continuous measurements. See [Sameoto et al. \(2000\)](#) for a detailed account.

Sampling method	Spatial resolution	Temporal resolution	Biological resolution	Long-term deployment
Plankton nets	low	low	high	no
Acoustic devices (ES, ADCP, AZFP)	high	high	low	yes
Optical plankton counters	high	high	low	yes
Video plankton recorders	high	high	high ^a	no

^aDepends on the camera resolution, which largely doesn't allow for fine life-stage discrimination compared to that facilitated by plankton collecting devices.

ES: Echo-sounder, ADCP: Acoustic Doppler Current Profiler, AZFP: Acoustic Zooplankton and Fish Profiler

3.3 Paper-II: Background and main findings

3.3.1 Models; are they any good?

Mechanistic models offer a cost-effective means of studying zooplankton vertical migrations in higher spatial, temporal and biological resolution. Here, vertical migrations are modeled in a bottom-up fashion, i.e. based on the patterns described in the literature, relating associated traits that are manifested in the migration to internal states of the migrant and external environmental variables using mathematical functions (see [Enquist & Ghirlanda 2005](#), [Soetaert & Herman 2009](#), [Grimm & Railsback 2013](#)). Simulations are performed in artificial environments of varying spatial and temporal resolution to provide environment-specific predictions about migratory behavior, which are validated using field data or comparing with behavioral patterns described in the literature ([Rykiel 1996](#)).

Since the pioneering work of [McLaren \(1963\)](#), numerous models have been used to describe DVM, SVM and associated life history strategies of various marine and freshwater zooplankton taxa. These involve classic life history models (e.g. [Fiksen & Giske 1995](#), [Fiksen & Carlotti 1998](#), [Ji 2011](#)), individual based simulation models (e.g. [Miller et al. 1998](#), [Hjøllo et al. 2012](#), [Ji et al. 2012](#)), simulation models involving groups of individuals (e.g. [Carlotti & Wolf 1998](#), [Fiksen 2000](#)) and process-oriented models (e.g. [Enright 1977](#), [Eiane & Parisi 2001](#)). Models of DVM often possess the highest spatial (≤ 1 m) and temporal (≤ 1 h) resolution (e.g. [Fiksen & Giske 1995](#), [Eiane & Parisi 2001](#), [Liu et al. 2003](#), [Burrows & Tarling 2004](#), [Hansen & Visser 2016](#)). Models of SVM, diapause and associated seasonal strategies usually possess coarser spatio-temporal resolution, where vertical spatial elements range from 1 m depth bins to few vertical habitats, and time units ranging from 1 h pings to few seasons (e.g. [Carlotti & Wolf 1998](#), [Fiksen & Carlotti 1998](#), [Hind et al. 2000](#), [Ejsmond et al. 2015](#), [Sainmont et al. 2015](#), [Banas et al. 2016](#)). These differences indicate the contrasting spatial and temporal scales that DVM and SVM occur, and as a result lifetime dynamics of DVM cannot be simulated in models of SVM without significantly elevating computer time. This most likely be the reason why models of SVM and associated seasonal strategies of high-latitude zooplankton either fully or partly (i.e. of younger developmental stages) disregard DVM. However, since DVM is a geographically, taxonomically and ontogenetically widespread behavior ([Huntley & Brooks 1982](#), [Osgood & Frost 1994](#), [Hays 1995](#), [Bianchi & Mislan 2016](#), [Gjøsæter et al. 2017](#), [Knutsen et al. 2017](#)), prudence of such simplifications and its potential influence on model predictions remains questionable. Therefore, to test this further, and to see if the elevation of computer time by modeling lifetime dynamics of DVM and SVM in concert is a favorable trade-off for biological information ensued, we developed a novel, computationally efficient modeling framework.

3.3.2 The strategy-oriented modeling framework

The strategy-oriented modeling framework was built around three core arguments. First, DVM and SVM are adaptive strategies that enhance opportunities of feeding, growth, survival and reproduction, and hence fitness ([Aidley 1981](#), [Alerstam et al. 2003](#), [Cresswell et al. 2011](#), [Litchman et al. 2013](#)). Second, DVM and SVM predominantly occur as responses to spatio

temporal variability of resources and risks (e.g. [Huntley & Brooks 1982](#), [Bollens & Frost 1989](#), [Kaartvedt 1996](#), [Dingle & Drake 2007](#), [Bandara et al. 2016](#)). Third, vertical migration is not itself a trait that can be modelled, but is manifested by several underlying morphological, physiological and behavioral attributes, and their size-age-or stage-specific variability (reviewed in, [Zink 2002](#), [Cresswell et al. 2011](#)).

Based on the above arguments, in paper-II ([Bandara et al. 2018](#)), we defined an entity termed 'vertical strategy', which represents a unique timing and amplitude of DVM and SVM and its ontogenetic trajectory. Vertical strategies were manifested using multiple (six) evolvable proxies. Payoffs rendered by vertical strategies hardwired to copepods born in different times of the year were assessed by their growth, survival and reproductive performances, which we termed fitness. For a given model environment, fitness was heuristically maximized using a Genetic Algorithm ([GA: Holland 1975](#)) to derive environment-specific optimal estimates of the vertical strategy, time of birth and several other associated life history traits. A graphical summary of this modeling framework is given in the Fig. 2 of [Bandara et al. \(2018\)](#).

In the model, six evolvable proxies of vertical strategies together with the time of birth (see Table 2 in Paper-II) produced a complex, seven-dimensional optimization problem, which could be efficiently solved using heuristic techniques ([Zanakis & Evans 1981](#)). Further, to improve the accuracy and efficiency of the optimization process, we used a Real-Coded Genetic Algorithm (RCGA: [Davis 1989](#), [Lucasius & Kateman 1989](#), [Herrera et al. 1998](#)), with pre-benchmarked selection, recombination and mutation operators that are known to perform well with multi-dimensional optimization problems. However, as there is no guarantee that the GA would converge on the globally optimal solution ([Zanakis & Evans 1981](#), [Rardin & Uzsoy 2001](#), [Strand et al. 2002](#)), we performed 10 replicate simulations of each scenario modelled (cf. [Record et al. 2010](#), [Maps et al. 2011](#)).

We developed, executed and analyzed the model using R™ v.3.3.1 ([R Core Team 2016](#)) in the RStudio™ v. 1.0.136 ([RStudio Team 2016](#)) integrated development environment. As execution rates of even the most well-optimized code in R™ can be inherently slower compared to most low-level computing platforms, such as C++ and FORTRAN ([Wickham](#)

[2015](#)), we used the high performance computing package *Rcpp* ([Eddelbuettel et al. 2011](#)) to speed-up simulations.

In the trials, we estimated that a typical model run (i.e. from the seeding of 10^6 strategies to converging on an optimal strategy) takes on average ca. 18.3 hours to complete on sequential basis (single-threaded operation) using an Intel® Core™ i7-7700K central processing unit (CPU) running at an overclocked turbo speed at ca. 4.9 GHz. A parallelization trial using a loop-level construct (e.g. [Larus 1993](#), [Oplinger et al. 1999](#), [Yu 2002](#)) was shown to be ca. 3.6 times faster on the same CPU, which possesses four processor cores (workers). However, since we did not have access to a supercomputer with numerous processor cores, the efficiency of the above parallel execution could not be properly utilized. Further, given the high number of replicate model runs to be executed (410 model runs in total, Table 4) we decided to execute the model on a sequential basis, while running up to six instances (replicates) of the model on the same 4-core CPU without significantly reducing performance. This created an excessive memory (random access memory, RAM) demand, which was addressed using memory-efficient data objects manifested by the high-performance computing package *bigmemory* ([Kane et al. 2013](#)). We halved the total computational time by using a second CPU (with same configuration) to run the model (Table 4). Through this, we demonstrated that this model can be executed with reasonable efficiency even with the limited computational resources at hand.

3.3.3 Model predictions

Through the establishment of the strategy-oriented modeling framework, we could efficiently model both DVM and SVM in higher spatial (1 m vertical) and temporal (1 h) resolution covering the entire annual life cycle of the modelled copepods. Model runs along a latitudinal gradient under variable levels of visual predation risk have shown how differential patterns of DVM, SVM and other life history traits, such as birth times, sizes of overwintering stages and breeding modes would emerge in response to both bottom-up (i.e. temperature and food availability) and top-down (predation risk) environmental drivers. In the model, DVM emerged as a response to elevated visual predation risk, through which survival was enhanced at the cost of growth potential. Thus, it appeared that at higher levels

of visual predation risk, the modelled copepods cannot not perform SVM earlier in the year, as the DVM-induced loss of growth potential yields more time to develop to overwintering stages. However, by delaying their birth times, copepods could use higher summertime temperatures to attain higher growth, which was efficiently traded off for survival through pronounced DVM, while performing SVM more or less at the same time of the year predicted for lower visual predation risk. These findings indicated that new noteworthy information can be extracted from modelling DVM and SVM of high-latitude zooplankton in concert.

Table 4: Management of computer time (Paper-II). Executing up to six instances of the model (replicates) sequentially on two CPUs produced the lowest computer time. The two CPUs used here were identical (Intel® Core™-i7 7700K).

Attribute	Sequential single execution	Parallel execution	Sequential multiple execution
Unique model runs	41 ^a	41 ^a	41 ^a
Total model runs (x 10 replicates)	410	410	410
CPU clock speed (overclocked turbo)	ca. 4.9 GHz	ca. 4.6 GHz ^b	ca. 4.6 GHz ^b
No. of cores	4	4	4
Unit run time	18.3 h	5.1 h	18.8 h
No. of simultaneous executions possible	1	1	6
Total predicted run time (one CPU)	7503 h (ca. 312 d)	2091 h (ca. 87 d)	1285 h (ca. 53 d)
Total predicted run time (two CPUs)	3752 h (ca. 157 d)	1046 h (ca. 44 d)	643 h (ca. 27 d)

^aIncludes one basic run, 28 runs for sensitivity analysis and 12 runs along the modeled latitudinal gradient under variable levels of visual predation risk

^bCPU speed reduces due to increased thermal throttling at higher workloads

3.4. Paper-II: key limitations and motivation for Paper-III

The strategy-oriented modeling framework has three major limitations. First, since the strategies hold the primus over individuals (i.e. strategies are hardwired to individuals), the biological resolution tend to contrast the higher spatio-temporal resolution. In nature, individual personalities and motivation ([Kralj-Fišer & Schuett 2014](#)) can often override behavioral decisions, such as timing and amplitude of vertical migration ([Hays et al. 2001](#), [Pearre 2003](#)). Although this can be easily built into the model as state-dependent processes of decision reversal (e.g. [Houston & McNamara 1992](#), [McNamara & Houston 1996](#)), we did not implement these changes to the current iteration of the model (both in Paper-II and -III) for simplicity. The second limitation is lack of quantitative feedbacks between strategies and the model environment, such as the impact of grazing on the duration of the productive season. Consequently, competition for resources in the form of density-dependent games between strategies (e.g. [Fiksen 2000](#)) are not present in this model. Finally, the implemented 1 m spatial and 1 h temporal resolution can overlook zooplankton behavior occurring in finer spatio-temporal scales ([Visser 2001](#), [Seuront et al. 2003](#)). However, such finer resolutions were not adopted in this study to keep computer time under manageable limits.

In addition to the above, the most important limitation of the Paper-II ([Bandara et al. 2018](#)) in the context of this study was that the modelled copepod does not fully represent the life strategies of Arctic *Calanus* species. The fixed generation time (i.e. one year) and body mass (maximum $\approx 333 \mu\text{g C}$) used in [Bandara et al. \(2018\)](#) did not fully represent those observed for *C. finmarchicus*, *C. glacialis* and *C. hyperboreus* (e.g. [Conover 1988](#), [Unstad & Tande 1991](#), [Hirche et al. 1994](#), [Falk-Petersen et al. 2009](#)). Therefore, there was a general need for relaxing the generation time and body size assumptions, and modeling three different species rather than a generalized model copepod. Further, there was also a need of extending the modelled latitudinal gradient in [Bandara et al. \(2018\)](#) beyond 75°N, possibly toward ca. 80°N to represent seasonally ice-covered waters of the high-Arctic.

3.5 Paper-III: background and main findings

3.5.1 An improved model

In Paper-III (unpublished manuscript), we relaxed the generation time and body size restrictions implemented by Paper-II ([Bandara et al. 2018](#)) and adopted species-specific parameterization to develop three model-species representing the behavioral and life history strategies of *C. finmarchicus*, *C. glacialis* and *C. hyperboreus*. Here, apart from the vertical strategy, we introduced two evolvable parameters to represent the plasticity of body sizes and generation times observed for above species. The modelled latitudinal gradient was also extended toward ca. 80°N, but for simplicity, we did not model the seasonal sea-ice cover. By adopting these modifications, we attempted to answer the important question that remain unanswered in Paper-I ([Bandara et al. 2016](#)), i.e. how vertical migrations and associated life history strategies of Arctic *Calanus* spp. are influenced by environmental variability?

The model was based on the same strategy-oriented construct ([Bandara et al. 2018](#)), but with few modifications. First, to adopt the two additional evolvable parameters we removed two least influential evolvable parameters used in the first model (Paper-II), i.e. the size-specificity of light sensitivity parameter and the overwintering depth parameter (cf. Table 2 of Paper-II with that of Paper-III). Here, the former represented the ontogenetic DVM trajectory and the latter allowed overwintering depth to evolve in response to external environmental variability. In Paper III, we implemented general rules (based on literature) to describe these behavioral patterns. The second major modification was the adoption of a novel growth sub-model to allow species-specific patterns of growth. The growth sub-model presented in Paper-III was based on the parameters calculated from [Maps et al. \(2011\)](#), and was designed to address some practical concerns of the above (see Appendix S2 of Paper-III for a detailed account). The third major modification was the implementation of exponential size-dependent visual predation risk function to replace the more conservative linear function used in Paper-II (cf. Fig. 4a of Paper-II with Fig. 5a of Paper-III).

Apart from these improvements of parameterization, the RCGA was also reinforced with novel recombination and mutation operators, i.e. Laplace crossover operator ([LX: Deep & Thakur 2007](#)) for recombination, and Makinen, Periaux and Toivanen method ([MPTM:](#)

[Toivanen et al. 1999](#)) for mutation. Further, the number of seeds (strategies) were increased up to 2.5×10^6 .

Table 5: Management of computer time (Paper-III). Executing up to 16 instances of the model (replicates) sequentially on a single Intel® Core™-i9 7920X CPU produced the lowest computer time.

Attribute	Sequential single execution	Parallel execution	Sequential multiple execution
Unique model runs	96 ^a	96 ^a	96 ^a
Total model runs (x 10 replicates)	960	960	960
CPU clock speed (turbo)	ca. 4.3 GHz	ca. 4.3 GHz	ca. 4.3 GHz
No. of cores	12	12	12
Unit run time	29.4 h	2.8 h	29.4 h
No. of simultaneous executions possible	1	1	16
Total predicted run time	28224 h (ca. 1176 d)	2688 h (ca. 112 d)	1764 h (ca. 74 d)

^aIncludes three basic runs, 12 runs for sensitivity analysis and 81 runs along the modeled latitudinal gradient under variable levels of visual predation risk

The 2.5-fold increase in the number of strategies simulated in the model had a significant influence on the model performance. To reduce computer time, we made some performance tweaks to the code, and removed the semi-stochastic predictive algorithm (see Appendix S2 in Paper-II) for copepod vertical search behavior. With these changes, we could adopt the increased number of strategies without significantly elevating the computer time. In the new model, the estimated average computer time needed to complete a single sequential model run (i.e. from the seeding of 2.5×10^6 strategies to converging on an optimal strategy) was ca. 29.4 hours, for an Intel® Core™ i9 7920X CPU running at a turbo clock speed

at ca. 4.3 GHz. In this 12-core CPU, the loop-level parallel execution trial on *RMPI* ([Yu 2002](#)) was ca. 10.7 times faster than the sequential execution. However, given the higher number of replicate model runs that had to be performed (960 model runs in total), we executed the model on sequential basis, while running up to 16 instances of the model on the above 12-core CPU (Table 5).

3.5.2 Model predictions

In Paper-III, we found that when top-down selection pressures were insignificant (i.e. low visual predation risk), there were obvious inter-specific south to north (ca. 60°N–80°N) trends in the predicted behavioral and life history strategies of *Calanus* spp., driven predominantly by the both bottom-up environmental selection pressures (i.e. temperature and food availability). However, increasing visual predation risk caused the top-down selection pressures to become predominant, and the three model-species employed largely similar tactics to counter the threat of visual predation. As a result, the influence of bottom-up selection pressures became diminished, where the predicted latitudinal and species-specific differences of birth times, timing of SVM, timing of reproduction, size of overwintering stages, size at sexual maturity, breeding modes and fecundity became less pronounced. Here, modest elevations of visual predation risk were countered with the manifestation of DVM behavior. However, at the point where this growth potential to survival trade-off became unfavorable, further increase of visual predation risk was countered with the plasticity of body sizes, which affected numerous allometric processes of growth, reproduction and survival, and created a significant impact on the observed behavior and life history patterns. We used these findings to conclude that top-down selection pressures serve a more significant role in shaping up of behavioral and life history strategies of Arctic *Calanus* species.

3.6 Paper III: key limitations and motivation for a follow-up paper

Although several assumptions of the model were relaxed from Paper-II to Paper-III, we still feel the need to relax some key assumptions for the next iteration of this model. First is the assumption of semelparity, which holds well for *C. finmarchicus* and to some extent *C.*

glacialis ([but see Kosobokova 1999](#)). However, our findings showed that species-CH (i.e. the model-species representing *C. hyperboreus*) had the potential for an iteroparous breeding strategy, which is well-documented for this species ([Hirche 2013](#)). Second, to represent the full spectrum of generation times exhibited by Arctic *Calanus* species, the model should allow life cycles with < 1 year generation times. In both the Paper-II and -III, the model hinted the possibility of some copepods (species-CF in Paper-III) to produce more than one generations per year, which is a common observation made for *C. finmarchicus* in many lower-latitude locations (e.g. [Fish 1936](#), [Lie 1965](#), [Matthews et al. 1978](#), [McLaren et al. 2001](#), [Bagøien et al. 2012](#)). However, this modification cannot be harbored without modifying the fitness function, which at its current formulation, does not allow to estimate fitness of < 1 year generation time. The third limitation of this model is the lack of flexible feeding behavior among the modelled species. In nature, all *Calanus* spp. have shown the ability to switch to alternative food sources in the absence of phytoplankton ([Runge & Ingram 1991](#), [Plourde & Runge 1993](#), [Søreide et al. 2008](#), [Campbell et al. 2016](#)). We believe that investigating the adaptive significance of this omnivores feeding strategy may hold the key to answer the question why *Calanus* spp. in the high-Arctic are active during the polar night. Finally, none of the above modifications would make complete sense unless the model is run in a stochastic model environment. Although timing of pelagic bloom is more or less predictable at ice-free lower latitudes, if the model is to run at ice-covered waters of the High-Arctic, we believe that year-to-year unpredictability in bottom-up and top-down environmental drivers should be allowed. This will enable us to check how model predicted optimal strategies would be robust to spatio temporal heterogeneity of environment (e.g. [Fiksen 2000](#), [Eiane & Parisi 2001](#), [Ji 2011](#)).

4. Concluding remarks and outlook

The strategy-oriented modeling framework employed in this study could efficiently simulate both DVM and SVM in higher spatio-temporal resolution, and predicted how environmental variability influences vertical migrations, and the influence of the latter on fitness and phenology of high-latitude herbivorous zooplankton (Paper-II and -III). Such predictions could not be derived from the year-round field data used in this study (Paper-I), likely because its lower spatio-temporal resolution did not match the spatio-temporal scales at which the vertical migrations usually occur. Therefore, these findings highlight the importance of future field and modeling studies to investigate DVM and SVM in concert.

To model DVM and SVM in concert, prospective modeling studies should focus on innovative ways of improving spatial and temporal resolution without significantly elevating computer time (e.g. [Carlotti & Wolf 1998](#), [Huse et al. 1999](#), [Eiane & Parisi 2001](#)). Coupling such high-resolution 1D models with 2D ocean circulation models would allow accurate predictions about the robustness of behavioral and life history strategies of high-latitude zooplankton over space and time. Nonetheless, due to the scarcity of high-resolution field data, validating these models would remain challenging. To facilitate model validation with existing data, we encourage building of spatially explicit models to represent locations where long-term zooplankton data are available (e.g. White Sea and weather-ship M).

Mechanistic models are useful tools to develop understanding about complex biological processes or behavioral patterns where empirical data are scarce ([Kjørboe 2008](#), [Bauer & Klaassen 2013](#)). Nonetheless, all models are simplifications of the real world, and thus do not render a complete picture about the questions under investigation. Similarly, the strategy oriented high-resolution models that were employed in this study only provide a temporary bridge to span the current knowledge gaps of zooplankton diel and seasonal vertical migrations. A thorough build-up of our understanding on the subject matter relies on the development of empirical research toward year-round monitoring of individual zooplankters in higher spatial and temporal resolution ([Davis et al. 1996](#)). Until those glory-days arrive, findings of high-resolution models of zooplankton vertical behavior would resonate in the field of plankton ecology.

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Paper-I

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Seasonal vertical strategies in a high-Arctic coastal zooplankton community

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ABSTRACT: We studied the larger (>1000 µm) size fraction of zooplankton in an Arctic coastal water community in Billefjorden, Svalbard (78°40' N), Norway, in order to describe seasonal vertical distributions of the dominant taxa in relation to environmental variability. *Calanus* spp. numerically dominated the herbivores; *Aglantha digitale*, *Mertensia ovum*, *Beroë cucumis*, and *Parasagitta elegans* were the dominant carnivores. Omnivores and detritivores were numerically less important. Descent to deeper regions of the water column (>100 m) between August and October, and ascent to the shallower region (<100 m) between November and May was the overall seasonal pattern in this zooplankton community. In contrast to other groups, *P. elegans* did not exhibit pronounced vertical migrations. Seasonal vertical distributions of most species showed statistical associations with the availability of their main food source. The vertical distribution of later developmental stages of *Calanus* spp. was inversely associated with fluorescence, indicating that they descended from the shallower region while it was still relatively productive, and ascended before the primary production had started to increase. Strong associations between the vertical distributions of secondary consumer *M. ovum* and *Calanus* spp., and tertiary consumer *B. cucumis* and *M. ovum* indicated that these carnivores seasonally followed their prey through the water column. We conclude that seasonal vertical migrations are a widespread trait in the high Arctic community studied, and predator–prey interactions seem particularly central in shaping the associations between the seasonal vertical strategies of adjacent trophic levels.

KEY WORDS: Seasonal vertical migration · Food availability · Trophic interactions · Pelagic environments · Predator–prey interactions

INTRODUCTION

Pronounced seasonal oscillations in abiotic (e.g. solar radiation, temperature, sea ice) and biotic (e.g. food availability, predation pressure) environments offer challenges to zooplankton in high latitudes. In particular, seasonality in food availability is believed to be a significant challenge (Clarke & Peck 1991, Conover & Huntley 1991, Hagen 1999, Varpe 2012). Arctic zooplankton possess adaptations to counter a seasonally variable food supply, such as energy stor-

age (Lee et al. 2006, Varpe et al. 2009), diapause (Carlisle 1961, Hirche 1996), and seasonal vertical migrations (Conover 1988). Zooplankton seasonal vertical migrations are understood as an adaptive behavior that optimizes their position in the water column in response to seasonal variability in the environment (Werner & Gilliam 1984). We refer to this behavior as their 'seasonal vertical strategy'. Seasonal vertical strategies of some high-latitude herbivorous zooplankton are well-documented (e.g. Conover 1988, Falk-Petersen et al. 2009), and their adap-

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tive value has also been analyzed in modeling studies (e.g. Fiksen 2000, Varpe et al. 2007).

The underlying regulation of zooplankton seasonal vertical strategies has been a subject of interest since early 1900s (Russell 1927, Banse 1964). Seasonal variability in hydrography (Hirche 1991), photo-period (Sømme 1934, Miller et al. 1991), and visual predation (Kaarvedt 1996, Dale et al. 1999, Kaarvedt 2000) are some external environmental cues that are thought to regulate seasonal vertical strategies. Internal (endogenous) regulation through seasonal changes in gonad development (Østvedt 1955), lipids and buoyancy (Visser & Jónasdóttir 1999), and long-term endogenous timers (Miller et al. 1991, Hirche 1996) have also been suggested.

Many components of zooplankton life strategies are viewed as adaptations to seasonal variations in food supply (Ji et al. 2010, Varpe 2012), but the influence of food availability on seasonal vertical strategies remains poorly understood, particularly for carnivorous species. As the seasonal food supply is more pronounced for Arctic herbivorous zooplankton (Conover & Huntley 1991, Hagen 1999), whose energetic demands mainly depend on a short period of annual primary production (Falk-Petersen et al. 2009), it can be argued that their seasonal vertical strategies are tightly coupled with food availability. Although vertical distributions of *Calanus* spp. appear to be associated with food availability (chlorophyll *a* distributions) in spring (Herman 1983, Søreide et al. 2008, Basedow et al. 2010), it is less well-studied for the rest of the year. Compared to herbivores, Arctic carnivorous and omnivorous zooplankton rely to a greater extent on a year-round food supply (Hagen 1999). Therefore, it has been suggested that their seasonal adaptations are less pronounced compared to herbivores (Ji et al. 2010, Varpe 2012). As vertical distributions vary seasonally in a number of carnivorous hydromedusae (e.g. Pertsova et al. 2006), ctenophores (e.g. Siferd & Conover 1992), chaetognaths (e.g. Grigor et al. 2014), euphausiids (e.g. Lass et al. 2001), and copepods (e.g. Vestheim et al. 2005), it appears that seasonal vertical strategies of Arctic carnivorous zooplankton are more diverse than previously thought. Since many carnivores rely on herbivores as their main food source, the potential influence of the vertical strategies of herbivorous zooplankton on their predators may be ecologically significant. This is portrayed in the findings of Nelson et al. (1997) and Sims et al. (2005), where a close resemblance between the vertical behavior of planktivorous sharks and the diel vertical migration (DVM) of herbivorous zooplankton were reported. Whether such relation-

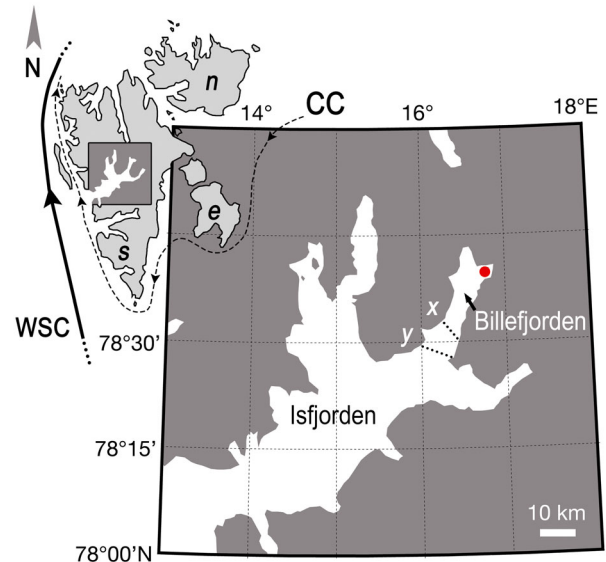


Fig. 1. Study area; sampling site is indicated by the red dot. *x* and *y* indicate coarse locations of the inner and outer sills of Billefjorden respectively. Positions of the west Spitsbergen current (WSC) and the coastal current (CC) were adopted from Svendsen et al. (2002). *s*: Spitsbergen, *n*: Nordaustlandet, *e*: Edgeøya

ships exist on seasonal timescales is not known, and open for investigation.

Investigating the seasonality of zooplankton strategies and interactions requires studying pelagic communities over the course of an annual cycle. Apart from a few studies (e.g. Hop et al. 2006), year-round zooplankton community investigations are rare in the Arctic. Here, we investigated seasonal vertical distributions of the dominant herbivore and carnivore zooplankton in a high-latitude coastal zooplankton community during a 10-month period in 2008 and 2009. We studied the extent to which the seasonal vertical distributions of the above zooplankton could be explained by the seasonal dynamics of their primary food source, or physical environmental variables such as temperature, salinity, and irradiance.

MATERIALS AND METHODS

Study site

Environmental variables and zooplankton samples were collected monthly between August 2008 and May 2009 at a 189 m deep station (78° 39.72' N, 16° 44.34' E) within the inner basin of Billefjorden, located at the west coast of Spitsbergen, the largest island in the Svalbard archipelago (Fig. 1). The inner basin of Billefjorden remains ice-covered from ca.

December to June (Arnkvaern et al. 2005). Two 50 to 70 m deep sills located near the mouth of the fjord (Fig. 1) act as a topographical barrier that hinders the advection of the Atlantic water masses into Billefjorden (Cottier et al. 2005, Nilsen et al. 2008). Because of this, Arnkvaern et al. (2005) argued that zooplankton population dynamics in Billefjorden are influenced more by internal processes than by advection.

Environmental variables

Temperature and salinity were profiled *in situ* using either a CTD/STD model DS 204 (SAIV) or a Seabird™ CTD (Sea-Bird Electronics). Since no CTDs were deployed on 27 August and 07 September 2008, and 23 March 2009 (Table 1), we obtained temperature and salinity data for these dates from a moored instrument series (www.sams.ac.uk/oceans-2025/arctic-mooring) deployed <0.5 nautical miles away (78° 39.76' N, 16° 11.24' E) from the sampling site (see Supplement 1 at www.int-res.com/articles/suppl/m555p049_supp.pdf). We measured photosynthetically active radiation (PAR) from a QSP 2300 log quantum scalar irradiance sensor (Biospherical Instruments), and fluorescence from a Seapoint™ chlorophyll fluorometer (Seapoint Sensors) affixed to the above mooring at 29 m. Fluorescence could not be accurately estimated due to the lack of fluorometer calibration coefficients for most

Table 1. Zooplankton samples and CTD casts collected during the study. A sample is a depth-stratified 0–180 m net haul. Day samples (D) were collected between 11:00 and 17:00 h; night samples (N), between 23:00 and 04:00 h local time (UTC + 1). Note that the lack of CTD data in August, September, and March (dashes in the rightmost column) were compensated by the data of the mooring (see Supplement 1 at www.int-res.com/articles/suppl/m555p049_supp.pdf)

Date (dd/mm/yyyy)	No. of Samples	Time of collection	CTD casts
27/08/2008	1	N	–
07/09/2008	2	D + N	–
23/09/2008	2	D + N	x
17/10/2008	3	D + N	x
04/11/2008	2	D + N	x
03/12/2008	2	D + N	x
14/01/2009	1	D + N	x
26/02/2009	3	D + N	x
23/03/2009	1	N	–
30/03/2009	1	D	x
20/04/2009	1	D	x
27/04/2009	1	D	x
04/05/2009	1	D	x

of the year. Therefore, raw voltage outputs of the fluorometer were presented as normalized values between 0 and 1 after removing some extreme readings (sensor noise). This provided an approximate variation of the fluorescence during the study, because according to the calibration equation (Seapoint Sensors; data not shown), fluorescence is estimated as a linear function of the voltage outputs.

Raw voltage outputs (O_i) of the irradiance sensor were converted to PAR by applying a wet calibration factor ($C = 5.05 \times 10^{12}$), and a dark voltage of 0.0130 V (Biospherical Instruments) as:

$$\text{PAR} = C(10^{O_i} - 10^{0.0130}) \quad (1)$$

Temperature and salinity measurements were visualized using the Spatial Analyst™ extension of ArcGIS™ version 9.3 (ESRI). Here, the data were interpolated temporally over the depth range using the natural neighbor method (Sibson 1981). Fluorescence and PAR data are presented as daily means. Sea ice charts developed by the Ice Information Portal of the Norwegian Meteorological Institute (<http://polarview.met.no/>) were used to describe the sea ice extent in Billefjorden during the study period.

Zooplankton

Zooplankton were sampled by vertical hauls using a WP-3 net (area of the opening: 1 m²; mesh size: 1 mm) fitted with a Nansen-type messenger-operated closing device. Samples were taken from the vessel, or with a tetrapod-mounted cable towed by a snowmobile at ca. 1 m s⁻¹ when sampling from sea ice. Three depth strata were sampled (0–50, 50–100, and 100–180 m), excluding the bottommost 10 m. Larger (>10 mm) gelatinous zooplankton that could dissolve upon formaldehyde preservation were identified, and their body lengths were measured immediately after collection. The rest of the samples were preserved in a borax-buffered 4 % formaldehyde-in-sea-water solution.

In the laboratory, the larger specimens were counted from the entire samples. The smaller and more numerous individuals (predominantly copepods) were counted in subsamples obtained using a box splitter (Motoda 1985) until a minimum of 100 individuals were counted per sample. On average, ~24 % (range: 0.15 to 100 %) of the total sample volume was used. Zooplankton were identified to the lowest possible taxonomic level, and classified into trophic groups according to the literature (see Table 2).

Prosoma lengths (PL) of copepods were measured to the nearest 0.1 mm using a stereomicroscope (Leica Microsystems). We measured bell heights (BH) of hydromedusae and total lengths (TL) of other zooplankton. The copepodite stage 4 (CIV) and older developmental stages of *Calanus hyperboreus* (which were the only stages captured in this species) were identified by the presence of an acute spine on their fifth thoracic segment (e.g. Parent et al. 2011). The rest of the *Calanus* spp. were identified by a length frequency analysis following Arnkværn et al. (2005) using the R (R Core Team 2013) package 'mixdist' v.0.5-4 (Macdonald & Du 2012). We used the PLs of 3908 CVs, 1409 adult females, and 387 adult males of *Calanus* spp. pooled over the study period for the analysis. PL boundaries derived by the length frequency analysis were evaluated against those published in relevant literature to distinguish species. We also used this method on monthly pooled length measurements (BH or TL) of other taxa to identify any size groups.

Zooplankton abundances (ind. m⁻³) were estimated assuming 100% filtration efficiency of the WP-3 net. Monthly mean abundances were used in data presentation and analyses. This was estimated by averaging the total abundance of a given taxon in a given month over the number of samples (i.e. net hauls) collected in that month (Table 1).

Seasonal vertical distributions of the dominant zooplankton species (i.e. those that contributed >0.1% of the total numerical abundance [corresponding to 5 ind. m⁻³], and were captured more or less throughout the investigation) were presented as monthly mean abundances in each depth stratum. Since the relative abundance of dominant taxa in each depth stratum in day and night replicate samples (Table 1) varied <9%, the mean abundances of the replicates were used in the presentation and analyses of seasonal vertical distributions.

Seasonal vertical strategies

In order to describe zooplankton seasonal vertical distributions as seasonal vertical strategies, we described the water column in 2 regions: a shallower region (0 to 100 m), and a deeper region (100 to 180 m). We considered the maximum sill depth of the fjord (~70 m), maximum thermohaline stratification depth (~80 m) recorded in the study, and the vertical resolution of our sampling design (minimum 50 m) in making the above discrimination. We estimated a vertical distribution index (V) for each spe-

cies by taking the difference between the population proportions of the 2 vertical regions in each month as:

$$V = \frac{(N_{0-100} - N_{100-180})}{(N_{0-100} + N_{100-180})} \quad (2)$$

where N_{0-100} and $N_{100-180}$ represent the monthly mean abundance of the shallow and deeper regions of the water columns, respectively. V ranges between -1 and 1, in which the upper limit represents the entire population distributed in the shallower region of the water column, and the lower limit represents the opposite scenario. Here we assumed the influences of zooplankton advection in and out of this community to be negligible (see Supplement 2 at www.int-res.com/articles/suppl/m555p049_supp.pdf), and therefore, the dynamics of V over the time series is primarily due to the vertical migration of zooplankton across the 2 vertical regions.

We used correlation analyses to describe the association between the monthly vertical distribution indices of the dominant taxa and physical (i.e. mean temperature, salinity, and PAR) and biological (availability of the main food source) environmental variables, assuming a linear association between the above. We tested the above variables for normality (Shapiro-Wilk test; Shapiro & Wilk 1965), and homoscedasticity (2-sample Levene's test; Levene 1960), and found that most variables violated the assumptions of parametric correlation tests. Therefore, we used the nonparametric Kendall's rank correlation test with adjustment to tied ranks (coefficient = τ_b) (Kendall 1938, 1945) in the analyses.

RESULTS

Environmental variables

The inner basin of Billefjorden was covered with land-fast sea ice from late December 2008 until the end of the investigation in May 2009 (Fig. 2a). Maximum PAR and fluorescence values were recorded between August and September, and decreased to 0.2 $\mu\text{mol m}^{-2} \text{s}^{-1}$ and 0.10 units respectively after November (Fig. 2a,b). Pronounced thermo-haline stratifications observed in the early part of the study broke down between November and January, and resulted in a well-mixed, cold (<-1.0°C), and relatively high saline (>34 PSU) water column (Fig. 2c,d). This lack of stratification persisted until the end of sampling.

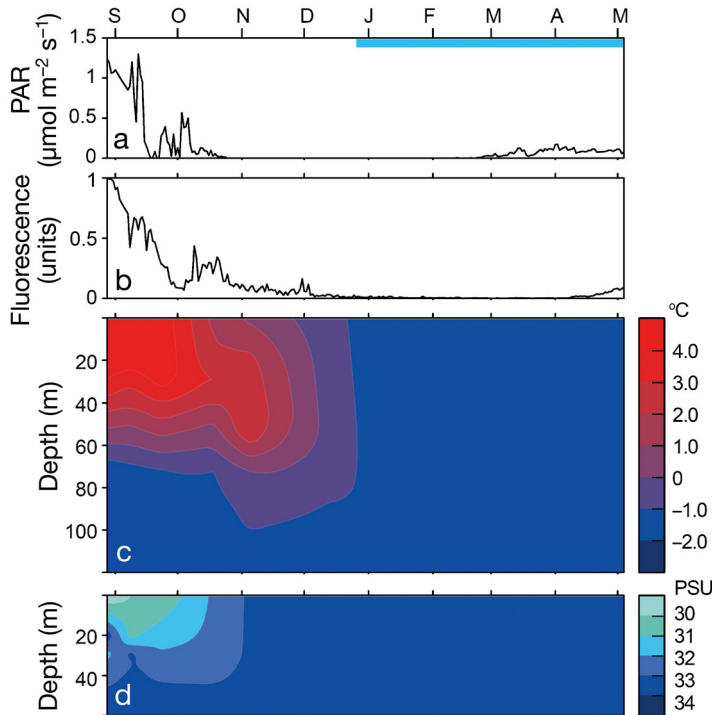


Fig. 2. Seasonal variability in (a) photosynthetically active radiation (PAR), (b) fluorescence (normalized between 0 and 1 unit), (c) temperature, and (d) salinity during the study. The blue bar in (a) indicates the period of land-fast sea ice cover. The ordinates of (c) and (d) are cropped at 120 and 60 m respectively due to the prevailing homogeneity of those parameters. Note that the abscissa extends from 27 August 2008 to 04 May 2009

Zooplankton community composition and trophic relationships

A total of 8 herbivores, 8 omnivores, 4 detritivores, and 17 carnivores comprised the 37 zooplankton taxa captured in this study (Table 2). The PL boundaries derived from the length–frequency analysis of *Calanus* spp. (Table 3) were in accordance with those published for *C. finmarchicus* and *C. glacialis* (see Supplement 3 at www.int-res.com/articles/suppl/m555p049_supp.pdf). Numerically, *C. glacialis* dominated the herbivore community (relative abundance ~77.6%; Table 2) alongside *C. finmarchicus* (~17%) and *C. hyperboreus* (~2%). Carnivores accounted for ~2.5% of the total numerical abundance (Table 2), and were dominated by the chaetognath *Parasagitta elegans* (~1.2%), the ctenophores *Mertensia ovum* (~0.5%) and *Beroë cucumis* (~0.4%), and the hydromedusa *Aglantha digitale* (~0.2%). Omnivorous and detritivorous zooplankton only contributed to ~1% of the total numerical abundance (Table 2).

Table 2. Zooplankton taxa captured in this study, their relative abundances, and feeding modes (references given as numbers in superscript). Indet.: indeterminate

Taxon	Feeding mode	Relative abundance (%)
<i>Bougainvillia</i> spp.	Carnivore ^{02, 42}	0.02
<i>Halitholus</i> spp.	Carnivore ¹⁸	0.01
<i>Sarsia</i> spp.	Carnivore ⁴²	<0.01
<i>Aglantha digitale</i>	Carnivore ^{05, 27, 42}	0.21
<i>Mertensia ovum</i>	Carnivore ^{35, 39}	0.45
<i>Beroë cucumis</i>	Carnivore ^{22, 31}	0.36
<i>Clione limacina</i>	Carnivore ^{04, 23}	0.02
<i>Limacina helicina</i>	Omnivore ^{17, 21}	0.03
<i>L. retroversa</i>	Herbivore ⁴⁰ , omnivore ¹⁷	<0.01
Gastropoda indet.	–	0.01
<i>Parasagitta elegans</i>	Carnivore ^{16, 24}	1.18
<i>Eukrohnia hamata</i>	Carnivore ^{07, 20}	0.08
<i>Anonyx nugax</i>	Scavenger ^{14, 19}	<0.01
<i>Themisto abyssorum</i>	Carnivore ¹⁶	<0.01
<i>T. libellula</i>	Carnivore ³²	<0.01
Amphipoda indet.	–	<0.01
<i>Munnopsis</i> spp.	Herbivore/detritivore ²⁸	<0.01
Isopoda indet.	–	<0.01
Mysidae indet.	–	0.03
<i>Meganyciophanes norvegica</i>	Carnivore ^{03, 10}	<0.01
<i>Thysanoessa inermis</i>	Herbivore ^{03, 19}	0.11
<i>T. longicaudata</i>	Omnivore ^{11, 36}	<0.01
<i>T. raschii</i>	Omnivore ^{09, 10}	<0.01
<i>Eualus gaimardii</i>	Carnivores ^{29, 37}	<0.01
<i>Pandalus borealis</i>	Omnivore ¹⁵	<0.01
<i>Necora puber</i>	Carnivore ^{26, 41}	<0.01
<i>Hyas</i> spp.	Carnivore/scavenger ⁴³	<0.01
<i>Calanus</i> sp.	–	0.11
<i>Calanus finmarchicus</i>	Herbivore ^{25, 38}	16.92
<i>C. glacialis</i>	Herbivore ^{25, 38}	77.56
<i>C. hyperboreus</i>	Herbivore ^{25, 38}	2.02
<i>Microcalanus</i> spp.	Herbivore/detritivore ¹³	<0.01
<i>Pseudocalanus</i> spp.	Herbivore ³⁴	<0.01
<i>Paraeuchaeta norvegica</i>	Carnivore ^{30, 33}	<0.01
<i>Metridia longa</i>	Omnivore ⁰¹	0.86
<i>Oikopleura</i> spp.	Particle feeder/omnivore ⁰⁶	<0.01
<i>Leptoclinus</i> spp. (larvae)	Carnivore ¹²	<0.01

References (in chronological order): ⁰¹Haq (1967); ⁰²Fraser (1969); ⁰³Ackman et al. (1970); ⁰⁴Conover & Lalli (1972); ⁰⁵Smedstad (1972); ⁰⁶Alldredge (1976); ⁰⁷Sullivan (1980); ⁰⁸Falk-Petersen et al. (1981); ⁰⁹Sargent & Falk-Petersen (1981); ¹⁰Falk-Petersen et al. (1982); ¹¹Williams & Lindley (1982); ¹²Eschmeyer et al. (1983); ¹³Hopkins (1985); ¹⁴Sainte-Marie & Lamarche (1985); ¹⁵Shumway et al. (1985); ¹⁶Falk-Petersen et al. (1987); ¹⁷Lalli & Gilmer (1989); ¹⁸Larson & Harbison (1989); ¹⁹Sainte-Marie et al. (1989); ²⁰Øresland (1990); ²¹Gilmer & Harbison (1991); ²²Purcell (1991); ²³Hermans & Satterlie (1992); ²⁴Alvarez-Cadena (1993); ²⁵Graeve et al. (1994); ²⁶Freire & Gonzalez-Gurriaran (1995); ²⁷Pagès et al. (1996); ²⁸Brusca (1997); ²⁹Graeve et al. (1997); ³⁰Olsen et al. (2000); ³¹Falk-Petersen et al. (2002); ³²Auel & Werner (2003); ³³Skarra & Kaartvedt (2003); ³⁴Lischka & Hagen (2005); ³⁵Lundberg et al. (2006); ³⁶Blachowiak-Samolyk et al. (2007); ³⁷Nygård et al. (2007); ³⁸Falk-Petersen et al. (2009); ³⁹Graeve et al. (2008); ⁴⁰Bernard & Froneman (2009); ⁴¹Silva et al. (2010); ⁴²Prudkovsky (2013); ⁴³Boxshall et al. (2015)

Table 3. Prosome length boundaries (mm) used to separate the 2 *Calanus* taxa, with their % composition within each developmental stage in parentheses. The rightmost column presents chi-squared statistic of the fitted model with the degrees of freedom in parentheses.
* $p < 0.05$; ** $p < 0.01$; *** $p < 0.001$

Developmental stage	Prosome length (% composition)		χ^2 (df)
	<i>C. finmarchicus</i>	<i>C. glacialis</i>	
CV	2.45–2.98 (38.96)	≥ 2.98 (59.74)	187.97** (11)
Adult females	2.38–2.92 (16.64)	≥ 2.92 (83.24)	54.47** (13)
Adult males	≤ 3.04 (7.20)	> 3.04 (92.08)	19.55* (10)

Based on the literature, we considered fluorescence as an indicator of the primary food source for herbivorous zooplankton, and identified *Calanus* spp. as the main prey of the secondary consumers *A. digitale*, *M. ovum*, and *P. elegans*, and *M. ovum* as that of the tertiary consumer *B. cucumis* (see references in Table 2).

Seasonal variability in abundance of the dominant zooplankton

Herbivores

The highest mean abundances of *C. finmarchicus* (~100 ind. m^{-3}), *C. glacialis* (~430 ind. m^{-3}), and *C. hyperboreus* (~13 ind. m^{-3}) were recorded between

August and November (Fig. 3a–c). During this period, CV was the dominant developmental stage of *C. finmarchicus* and *C. glacialis* (>95%: Fig. 3d,e). After November, relative abundance of CV decreased, and adult male and female copepodites increased. In *C. hyperboreus*, CIV was the dominant developmental stage throughout the study (Fig. 3f).

Carnivores

The mean abundances of *A. digitale* and *M. ovum* peaked at ~4 ind. m^{-3} in October (Fig. 4a,b). *B. cucumis* was captured in relatively large numbers (mean abundance: ~2.5 ind. m^{-3}) in October and May (Fig. 4c). We could not identify any size groups of the 3 above species from length–frequency analyses. However, their abundance peaks were dominated by relatively small individuals (mean \pm SD body length: 6.6 ± 1.5 mm for *A. digitale*, 6.7 ± 5.3 mm for *M. ovum*, and 2.9 ± 1.6 mm for *B. cucumis*; Fig. 4 e–g). The mean body lengths of *A. digitale* and *M. ovum* increased throughout the study period, while that of *B. cucumis* decreased after reaching a maximum (9.31 ± 6.4 mm) in November. *P. elegans* was captured in higher numbers in September (~5.5 ind. m^{-3}), December (~4.5 ind. m^{-3}), and between April and May

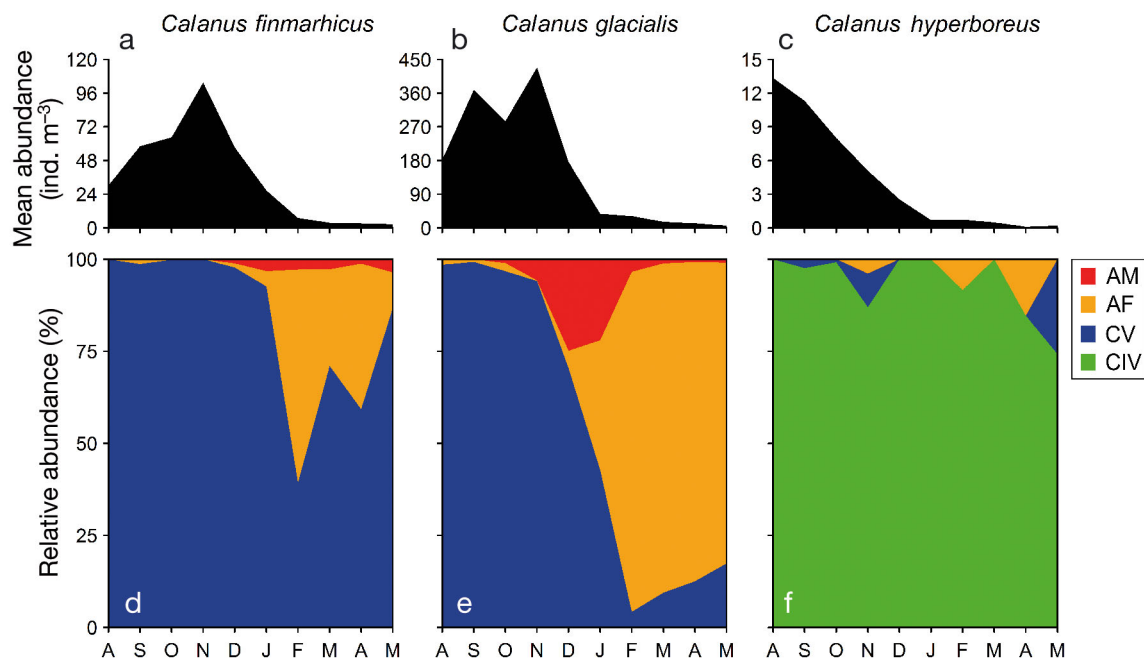


Fig. 3. Seasonal variability in (a–c) mean abundance and (d–f) relative developmental stage composition of dominant herbivores during the study. AM: adult males; AF: adult females; CV and CIV: copepodite stage 5 and 4, respectively

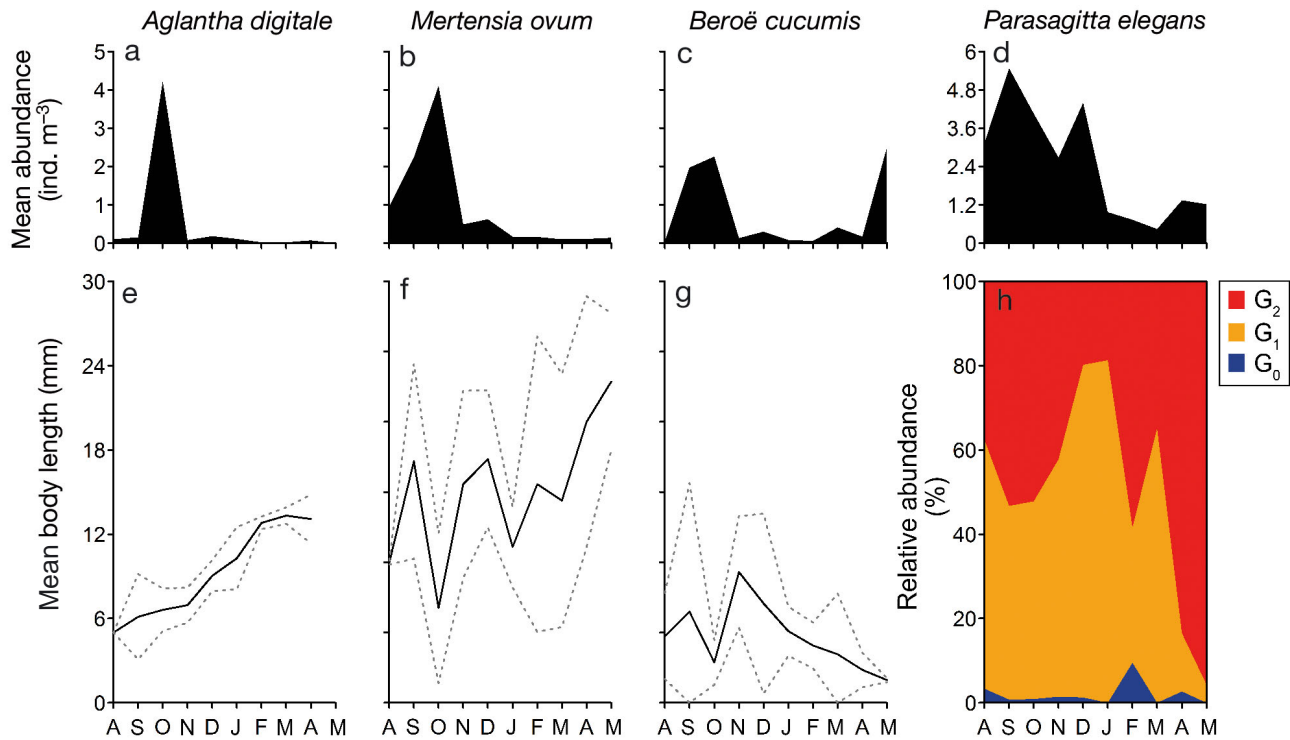


Fig. 4. Seasonal variability in (a–d) mean abundance and (e–h) mean body length of dominant carnivores during the study. Body lengths are presented as bell height for *A. digitale* and total length (TL) for other species. The TL dynamics of *P. elegans* in (h) is presented as variation in relative abundance of the 3 size groups. Dashed lines in (e–g) denote standard deviation of body length (mm)

(~ 1.5 ind. m^{-3}) (Fig. 4d). We derived 3 size groups for *P. elegans* from the length–frequency analysis (G_0 , G_1 , and G_2 ; see Supplement 4 at www.int-res.com/articles/suppl/m555p049_supp.pdf). The first abundance peak was composed of more or less equal proportions of the 2 relatively large size groups (G_1 : mean \pm SD TL: 23.4 ± 1.8 mm; G_2 : 34.2 ± 1.4 mm), with G_1 dominating $\sim 80\%$ of the second abundance peak (Fig. 4h). The relative abundance of G_2 increased from January to $>80\%$ in April and May, while the smallest size group (G_0 : 14.7 ± 1.2 mm) remained less prominent ($<10\%$) throughout the investigation.

Seasonal variability in vertical distribution of the dominant zooplankton

Herbivores

Between August and November, the mean abundance of *C. finmarchicus* (CV) and *C. glacialis* (CV and adult females) in the lower 80 m of the water column gradually increased (Fig. 5a,b). Conversely, the mean abundance of these 2 species in the upper 100 m decreased from August, and reached a mini-

mum in October, during which their vertical distribution indices were at the lowest ($V \sim -0.9$; Fig. 6a,b). From November onwards, *C. finmarchicus* and *C. glacialis* CVs had relocated to the upper 100 m along with adult copepodites. By February, the vertical distribution indices of these 2 species reached the maximum ($V \sim 0.6$ for *C. finmarchicus* and $V \sim 0.8$ for *C. glacialis*). Thereafter, the mean abundance of CV and adult copepodites of the above species in the upper 100 m decreased, and by the end of the investigation in May, their vertical distribution indices remained around zero. The mean abundance of CIV *C. hyperboreus* in the lower 80 m progressively decreased from August, and was only distributed in the upper 100 m between November and January (Figs. 5c & 6c). From February onwards, a few *C. hyperboreus* CIV, CV and adult female copepodites (mean abundance <1 ind. $m^{-3} mo^{-1}$) were relocated in the lower 80 m.

Carnivores

The mean abundance of *A. digitale*, *M. ovum*, and *B. cucumis* in the upper 100 m gradually

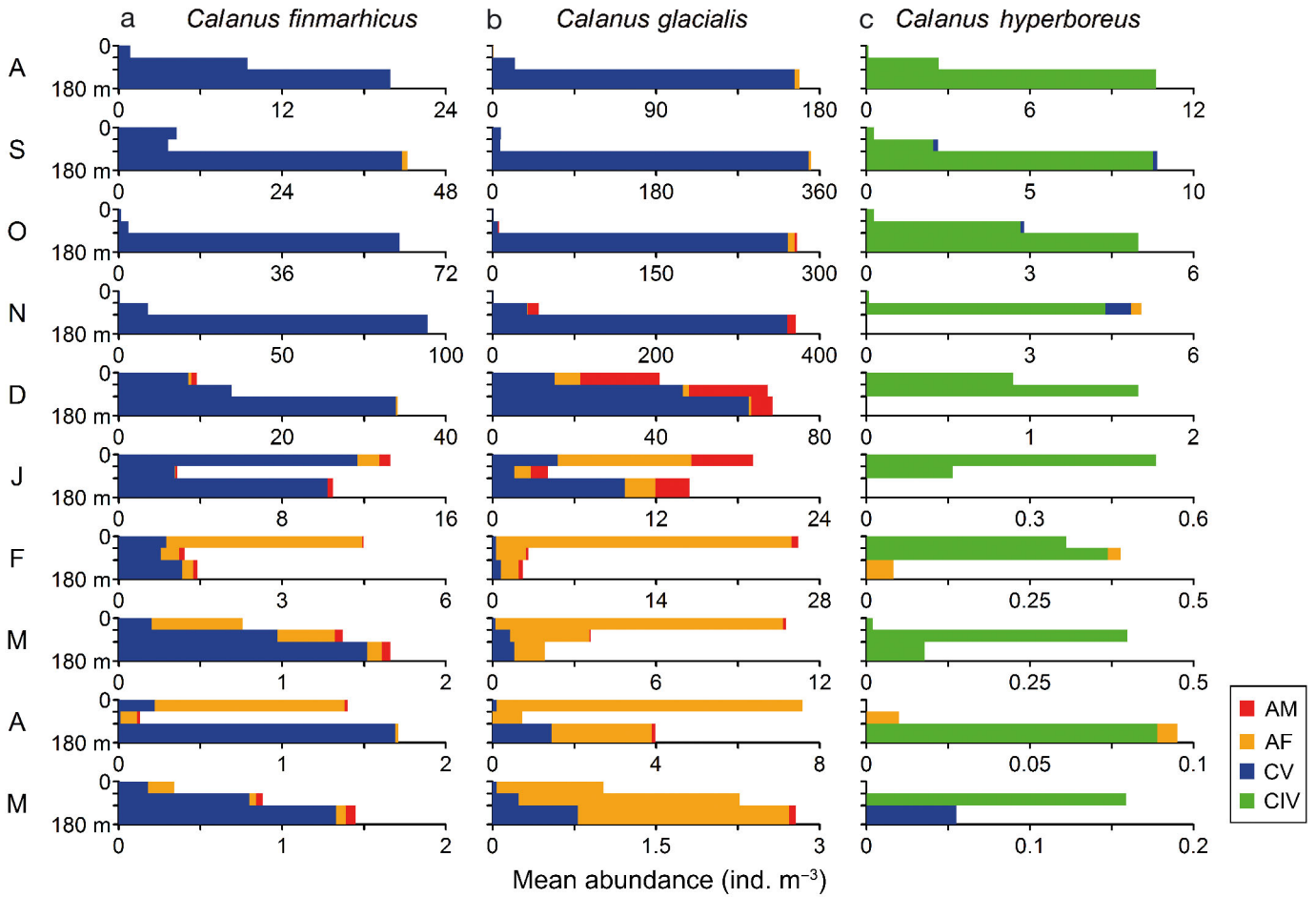


Fig. 5. Seasonal vertical distributions of dominant herbivores during the study. Ordinates represent depth (0–50, 50–100, and 100–180 m). AM: adult males, AF: adult females; CV and CIV: copepodite stage 5 and 4, respectively

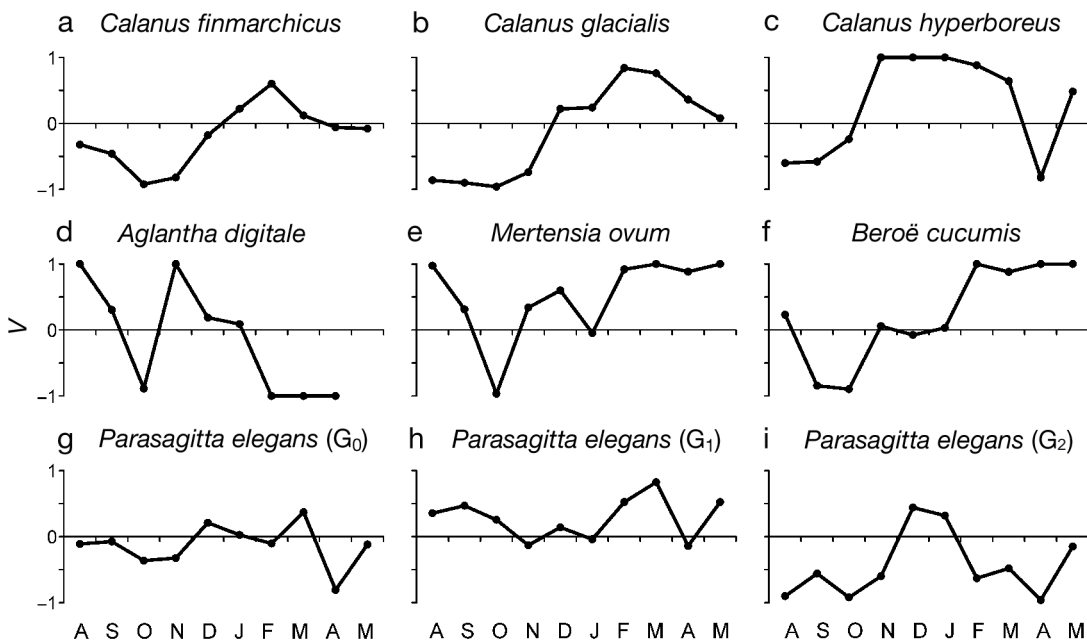


Fig. 6. Seasonal variability in the vertical distribution indices (V) of the dominant zooplankton taxa during the study. V ranges from -1 to 1 , in which the latter represents the entire population distributed in the shallower region, and the latter represents the opposite scenario. *A. digitale* was not captured to compute its V in May. See Supplement 5 for more information

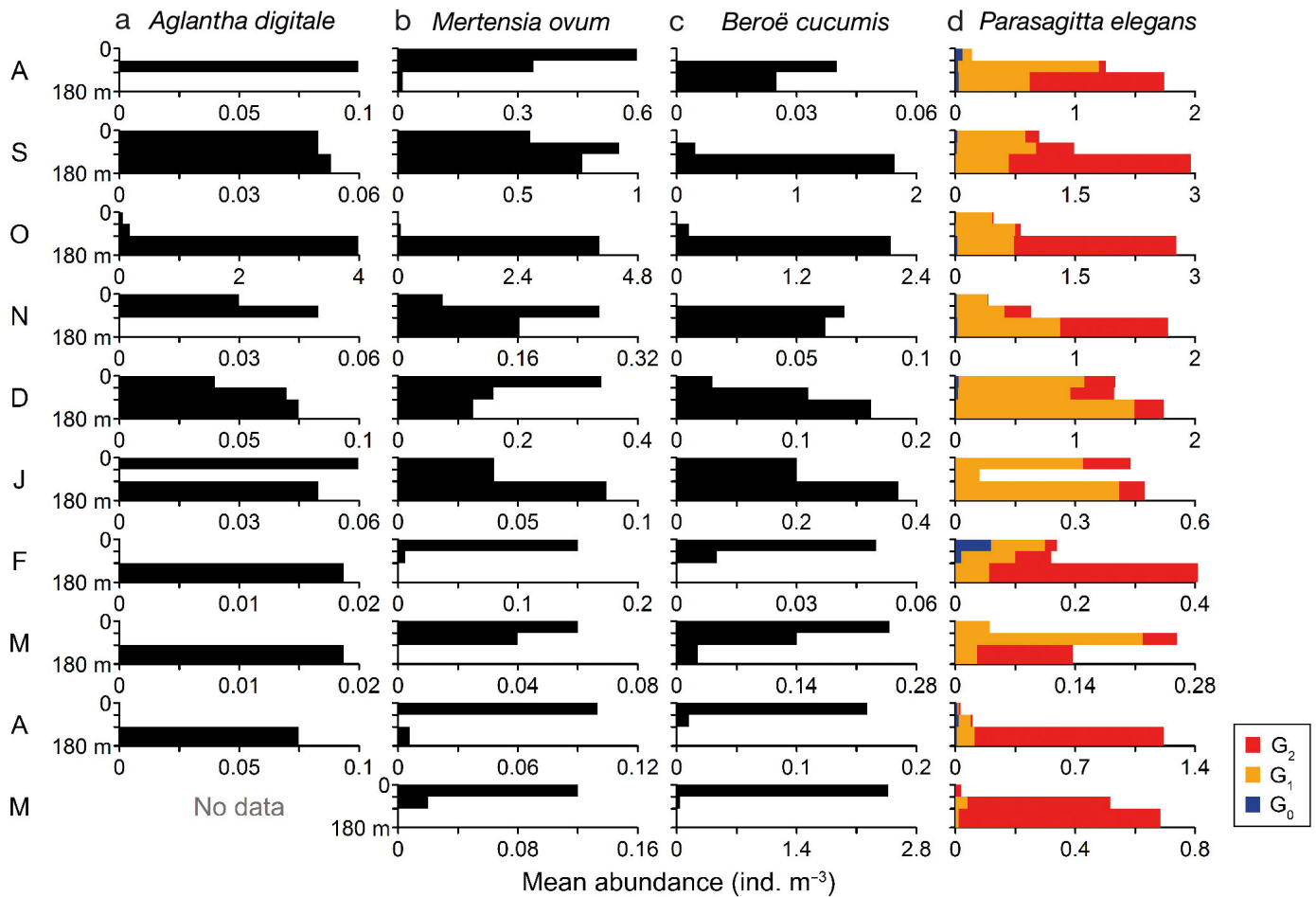


Fig. 7. Seasonal vertical distributions of dominant carnivores during the study. Ordinates represent depth (0–50, 50–100, and 100–180 m). Note that *Aglantha digitale* was not captured in May 2009

decreased from August (Fig. 7a–c), and their vertical distribution indices gradually decreased to ~ -0.9 in October (Fig. 6d–f). From November onwards, the mean abundance of *M. ovum* and *B. cucumis* in the upper 100 m, and their vertical distribution indices gradually increased, and the latter remained ~ 1 from February until the end of sampling in May (Fig. 6e, f). Although *A. digitale* had relocated to the upper 100 m between November and January, it was captured in the lower 80 m after February (Fig. 7a). Throughout this study, *P. elegans* was captured in all 3 depth strata (Fig. 7d). The vertical distribution index of the G_2 size group of *P. elegans* remained < -0.5 for most of the time series (Fig. 6i), indicating that $>75\%$ of its population was distributed in the lower 80 m throughout the study. Conversely, the G_0 and G_1 size groups were distributed across the entire depth range (Fig. 6g,h).

Seasonal vertical distributions and environmental variables

The vertical distribution index (V) of *Calanus* spp. (all species and developmental stages combined; see Table S5 in Supplement 5 at www.int-res.com/articles/suppl/m555p049_supp.pdf) showed a strong negative association with mean fluorescence ($\tau_b = -0.72$, $p < 0.01$, $n = 10$), and a weak negative association with mean temperature ($\tau_b = -0.49$, $p = 0.05$, $n = 10$) (Table 4, Fig. 8a). While the vertical distribution index of *M. ovum* showed a moderate positive association with that of *Calanus* spp. ($\tau_b = 0.51$, $p = 0.04$, $n = 10$), we found a strong positive association between the vertical distribution indices of *B. cucumis* and *M. ovum* ($\tau_b = 0.71$, $p < 0.01$, $n = 10$) (Table 4, Fig. 8b,c). The vertical distribution index of *A. digitale* showed a moderate negative association with mean temperature ($\tau_b = -0.57$, $p = 0.04$, $n = 9$). Vertical distribution indices of *P. elegans* were not significantly associ-

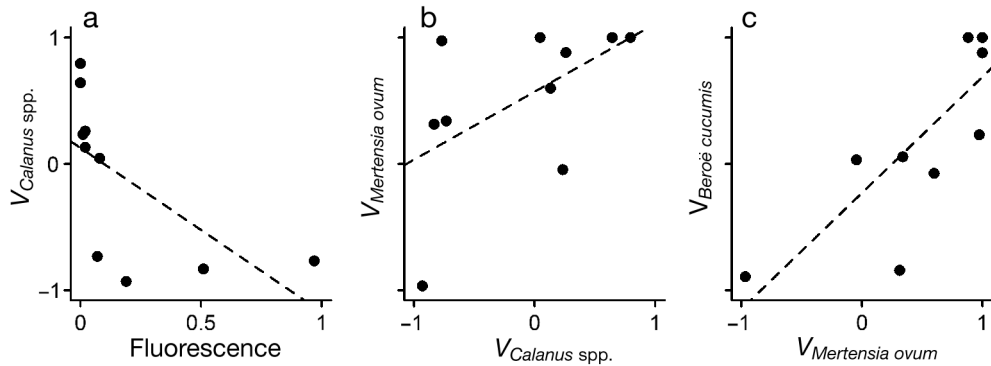


Fig. 8. Representation of statistically significant relationships between the vertical distribution indices (V) of dominant taxa and the availability of their main food source (cf. Table 4). Trend lines (dashed) were estimated by linear regression, and are solely for visualization of patterns in the data

Table 4. Associations between the vertical distribution indices (V) of dominant taxa and environmental variables presented as Kendall's rank correlation coefficients (τ_b). See Fig. 8 for additional information. PAR: photosynthetically active radiation; G_0 , G_1 , G_2 : size groups based on length–frequency analysis; * $p < 0.05$; ** $p < 0.01$

Species	Temperature	Salinity	PAR	Availability of the main food source			n
				Fluorescence	$V_{Calanus\ spp.}$	$V_{M.\ ovum}$	
<i>Calanus spp.</i>	-0.49*	0.31	-0.30	-0.72**	–	–	10
<i>Aglantha digitale</i>	-0.57*	-0.21	0.06	–	-0.53	–	9
<i>Mertensia ovum</i>	-0.30	0.12	0.14	–	0.51*	–	10
<i>Beroë cucumis</i>	-0.44	0.21	0.05	–	–	0.71**	10
<i>Parasagitta elegans</i> (G_0)	-0.13	0.09	-0.26	–	0.24	–	10
<i>P. elegans</i> (G_1)	-0.02	-0.11	0.14	–	0.04	–	10
<i>P. elegans</i> (G_2)	-0.13	0.27	-0.44	–	0.16	–	10

ated with any physical or biological environmental parameters that were used in our analyses (Table 4).

DISCUSSION

Seasonal patterns in vertical distributions and their relation to environmental variability

A gradual decrease in the vertical distribution index from August to October, and an increase from November to May were common to most of the investigated herbivorous (*Calanus spp.*) and carnivorous (*Aglantha digitale*, *Mertensia ovum* and *Beroë cucumis*) zooplankton taxa (Fig. 6a–f). Descent to the deeper region (>100 m) of the water column in early autumn, and ascent to the shallower region (<100 m) from late autumn to early spring was hence the overall seasonal pattern in this high Arctic zooplankton community. During their descent in the autumn, large numbers of zooplankton appeared to migrate from the warmer, sunlit, and productive shallow waters of this fjord (Fig. 2). Zooplankton abundances sharply declined during the winter (Figs. 3 & 4), and during the spring, most of the remaining individuals had ascended to a colder, darker, and unproductive

water mass. As an exception, *Parasagitta elegans* did not show seasonal migrations (Fig. 6g–i).

Seasonal vertical distributions of most zooplankton taxa showed statistical associations with the availability of their main food source (Table 4, Fig. 8). The inverse association between the vertical distribution index of *Calanus spp.* and mean fluorescence indicates that they descended from the shallower region while it was relatively productive, and ascended before the primary production had started to increase (Figs. 2b & 6a–c). Therefore, it seems that the seasonal vertical strategies of the dominant herbivorous zooplankton in this study were not regulated by food (phytoplankton) availability. As vertical distribution indices of the secondary consumer *M. ovum* and *Calanus spp.*, and the tertiary consumer *B. cucumis* and *M. ovum* were positively associated, we argue that these predatory zooplankton seasonally followed their prey (e.g. Fraser & David 1959, Torres et al. 1994, Hagen 1999). The seasonal vertical strategies of the above carnivores were likely regulated by seasonality in food availability (i.e. seasonal vertical strategies of their main prey), and further indicates that seasonal vertical strategies of zooplankton in lower trophic levels influence those in higher levels through trophic interactions. Still, we observed con-

siderable variability, and a lack of seasonal migrations in *P. elegans*. Consequently, numerous other factors, such as the timing and trade-offs between feeding and other life cycle events (Heath 1999, Varpe 2012), differences in prey selection (Greene 1986), feeding on alternative food sources (Hirche & Kwasniewski 1997, Søreide et al. 2006, Casanova et al. 2012), and predation risk (Kaartvedt 1996, Dale et al. 1999, Varpe & Fiksen 2010) may also have contributed to the regulation of the observed seasonal vertical strategies.

Seasonal vertical strategies of the dominant zooplankton

Herbivores

The CVs of *Calanus finmarchicus*, CVs and adult females of *C. glacialis*, and CIVs of *C. hyperboreus* likely resided in the deeper region until November (e.g. Conover 1988, Falk-Petersen et al. 2009, our Fig. 5). The gradually decreasing vertical distribution indices of *C. finmarchicus* and *C. glacialis* until October (Fig. 6a,b), and simultaneous increase in their mean abundances in the deeper region (Fig. 5a,b) indicate that a considerable fraction of the CVs of these 2 species descended and recruited to their deep water populations in the autumn. Conversely, the vertical distribution data of *C. hyperboreus* indicate neither a descent (which may have occurred prior to the commencement of sampling), nor recruitment to its deep water population (Figs. 5c & 6c). In order to build up energy reserves, a fraction of the *C. finmarchicus* and *C. glacialis* CVs may have grazed in the shallower region relatively late into the productive season prior to their descent (Fig. 2b). These CVs may have been the *Calanus* spp. reported by Berge et al. (2014) that contributed to the acoustic backscattering detected near a chlorophyll maximum in this fjord in late September. Østvedt (1955), Pedersen et al. (1995) and Hirche (1996) also observed a part of the summer–autumn *C. finmarchicus* population feeding in surface waters, while the rest resided in deep waters.

The gradually increasing vertical distribution indices indicate an ascent of *Calanus* spp. between November and February (Fig. 6a–c). By February, a maximum of ~80% of the *Calanus* community had ascended to the shallower region (Fig. 5). Similar to our findings, Daase et al. (2014) and Blachowiak-Samolyk et al. (2015) reported shallow vertical distributions (<100 m) of *Calanus* spp. in January from

~81°N in Rijpfjorden, Svalbard. However, the timing of the ascent we report here is earlier than the March to June period reported in most high-latitude investigations (e.g. Heath 1999, Gislason & Astthorsson 2000, Hirche & Kosobokova 2011, Melle et al. 2014).

As the vertical strategy of *Calanus* spp. was inversely related to fluorescence, it is unlikely that food availability served as a primary cue for their descent and ascent. However, a definitive conclusion on this matter cannot be made since vertical fluorescence profiles were not used in our study. We suggest that these herbivores, dominated by *C. glacialis*, ascended early as a part of a capital breeding strategy or to feed on ice algae, which were not detected by our fluorescence measurements (e.g. Varpe et al. 2009, Søreide et al. 2010). *Calanus* spp. use ice algae as an alternative food source to spawn prior to the phytoplankton bloom (Runge & Ingram 1991, Hirche & Kwasniewski 1997, Søreide et al. 2010). A summer–autumn descent while there is still food available near the surface, and ascent to shallow waters during the dark, unproductive winter (Fig. 2a,b) suggest a migration driven by processes other than the availability of food. The negative association between the *Calanus* vertical distribution index and mean temperature (Table 4) reflects the tendency of the seasonal descent and subsequent induction of diapause in *C. glacialis* to occur in relation to the summer–autumn warming of the surface waters (Niehoff & Hirche 2005, Pertsova & Kosobokova 2010). The overwintering depth and timing of the seasonal migration of *Calanus* spp. can also be influenced by planktivorous fish (Kaartvedt 1996, Dale et al. 1999, Kaartvedt 2000, Varpe & Fiksen 2010). Although we did not sample fish populations in this investigation, trawl samples collected in Billefjorden in August 2008 showed that ~60% of the stomach contents of polar cod *Boreogadus saida* consisted of *Calanus* spp. (Renaud et al. 2012). Therefore, the potential influence of visual predation on the seasonal vertical strategies of *Calanus* spp. in this fjord should not be ruled out.

Carnivores

Vertical distribution data of *A. digitale*, *M. ovum*, and *B. cucumis* indicate that these carnivores descended to the deeper region between August and October (Fig. 6d–f). From November onwards, *M. ovum* and *B. cucumis* gradually ascended and remained in the shallower region from February to the end of this investigation in May. Unlike the 2 cteno-

phore species, the ascent of *A. digitale* is not clearly evident (Fig. 6d). It should be noted that the vertical distribution data of this species after November may not be accurate due to its low numerical abundance (Fig. 4a). Descent to deeper waters in autumn, and ascent to shallower waters between spring and summer has been reported for *A. digitale*, *M. ovum*, and *B. cucumis* from ~59°N in the Northeast Atlantic (Williams & Conway 1981), ~62.5°N in Frobisher Bay (Percy 1989), ~67°N in White Sea (Pertsova et al. 2006), and ~74°N in Resolute Passage (Siferd & Conover 1992). In addition, shallow spring–summer vertical distributions of *M. ovum* and *B. cucumis* have been reported from ~55.5°N in the Bornholm basin of the Baltic Sea (Lehtiniemi et al. 2013), between 68 and 80°N in the Barents Sea and Fram Strait, (Swanberg & Båmstedt 1991a), and between 72 and 75°N in the western Arctic Ocean (Purcell et al. 2010).

Based on the positive association between the vertical distribution indices (Table 4, Fig. 8b), we argue that *M. ovum* seasonally followed *Calanus* spp. *M. ovum* is a secondary consumer that feeds on *Calanus* spp., and specifically on their older developmental stages (Greene 1986, Purcell 1991, Swanberg & Båmstedt 1991b). In the winter, *M. ovum* feeds on overwintering *Calanus* populations (Larson & Harbison 1989, Siferd & Conover 1992) and accumulates lipids (Percy 1989, Lundberg et al. 2006). Therefore, the older developmental stages (CIV, CV, and adult copepodites) of *Calanus* spp. sampled in this study may have served as a main prey source for *M. ovum*, and this predator–prey relationship is reflected by their similar vertical strategies. However, it should be noted that younger developmental stages of *Calanus* spp. which occupy shallower waters between March and May in this fjord (e.g. Arnkværn et al. 2005, Bailey 2010) may also have been a potential source of prey for *M. ovum*. Although *A. digitale* is a secondary consumer that primarily feed on copepods (see references in Table 2), its vertical distribution was not significantly associated with that of *Calanus* spp. (Table 4). Despite the similarities in the vertical strategies of *A. digitale* and *M. ovum* until October (Fig. 6d,e), the low numerical abundances of the former may have inaccurately represented its vertical distribution thereafter, and probably affected the results of the correlation analyses.

The positively associated vertical distribution indices suggest that the predatory ctenophore *B. cucumis* seasonally followed *M. ovum* (Table 4, Fig. 8c). *B. cucumis* is a tertiary consumer that specifically feeds on *M. ovum* (see references in Table 2). Therefore, it

is likely that the strong predator–prey relationship between these 2 ctenophores were reflected in their markedly similar vertical strategies (Fig. 6e,f). Similar spatial associations between these 2 species have been reported from ~74°N in Resolute Passage (Siferd & Conover 1992), and between 75 and 79°N in the Barents Sea (Swanberg & Båmstedt 1991a, Søreide et al. 2003). Although the mean TL of *M. ovum* became substantially larger than that of *B. cucumis* after November (Fig. 4f,g), it may not have affected their predator–prey relationship as *Beroë* can feed on prey larger than itself (Tamm & Tamm 1991), or on body parts of the prey (Swanberg 1974).

The accuracy of interpreting statistical associations between predator and prey zooplankton, as their trophic relationships can be hampered by the coarse vertical resolution of our samples (e.g. Pearre 1979). It is possible for predator and prey zooplankton to coexist in a depth stratum of 50 m (the vertical sampling resolution of this study) without encountering each other. As this bias tends to be pronounced in periods with low predator and/or prey abundances (e.g. Greene 1986), we did not interpret the vertical strategies of *A. digitale* (after November), or the G₀ size group of *P. elegans* in detail (Fig. 4a,d,h). Therefore, further analyses (e.g. gut content analyses and dietary lipid analyses) would be required in order to verify whether the associations between the vertical strategies of predators and prey zooplankton observed in this study truly reflect their trophic interactions.

The 3 size groups of *P. elegans* did not show pronounced seasonal migrations (Fig. 6g–i) irrespective of the seasonal oscillations of the environmental parameters observed in this study (Table 4). However, the largest size group (G₂) occupied the deeper region for most of the study, while the smaller G₀ and G₁ size groups were distributed throughout the water column. Deep water residence of larger individuals of *P. elegans* has been documented from 50°N at ‘Ocean Station P’ (Sullivan 1980, Terazaki & Miller 1986), ~75°N in Baffin Bay (Samemoto 1987), and ~78°N from our study location in Billefjorden (Grigor et al. 2014). The vertical strategy of the smallest size group (G₀) may not be accurate because our samples did not capture sufficient numbers of those sizes Grigor et al. (2014) reported from this fjord (sampled by nets with finer mesh size and documented as cohort–0 of their study: cf. length data in Table S4 in Supplement 4).

As larger chaetognaths prefer larger prey, such as the older development stages of *Calanus* (Greene 1986, Falkenhaug 1991, Saito & Kiørboe 2001), it is

likely that the G₂ size group of *P. elegans* fed on *Calanus* copepodites occupying the deeper regions of the water column. Despite the ascent of *Calanus* spp. between November and February, a fraction of the *C. finmarchicus* and *C. glacialis* population was observed in the deeper region throughout this study (Fig. 5a,b), and may have served as a year-round prey source for the largest *P. elegans* size group. In support of this view, gut content and lipid analyses of *P. elegans* collected from Billefjorden and other adjacent fjords by Grigor et al. (2015) suggests that *P. elegans* primarily feed on *Calanus* spp. It remains unclear why the relatively small size fraction of *P. elegans* population remained in the shallower region throughout this investigation (Fig. 7d). One possibility is that they may have preyed on smaller developmental stages of *Calanus* spp. and smaller copepod species, such as *Oithona similis*, *Microcalanus* spp., *Pseudocalanus* spp., and *Metridia longa* (Falkenhaus 1991, Walkusz et al. 2003, Grigor et al. 2015), prey categories which were undersampled by the large mesh width of the WP-3 net used in our investigation.

CONCLUSIONS

This study is one of few that have investigated seasonal vertical distributions of multiple members of a zooplankton community in the Arctic over a near-annual, high-resolution time series. Our findings suggest that seasonal vertical migrations are a widespread trait in the community, and that seasonality in food availability relates to seasonal vertical strategies of zooplankton in different trophic levels. This relationship was positive and strongest for the associations between herbivores and secondary consumers, and between secondary consumers and tertiary consumers. Further year-round field investigations that can combine high-resolution sampling methods with high spatial resolution (e.g. Norrbin et al. 2009), information on individual variability in size and energy reserves (e.g. Vogedes et al. 2010), accurate species determination (e.g. Parent et al. 2011, Gabrielsen et al. 2012), and year-round, mooring-based monitoring of the environment would be necessary to test the generality of our findings.

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The following supplements accompany the article

Seasonal vertical strategies in a high-Arctic coastal zooplankton community

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Supplement 1. Sources of additional hydrographic data

Table S1. Sensors of the mooring from which temperature and salinity data were obtained for August 27 and September 07, 2008, and March 23, 2009

Parameter	Sensor	Moored depth(s) (m)
Temperature	Seabird 16plus SeaCAT recorder	30
	VEMCO minilog-II-T thermal logger	43, 53, 63, 73, 111, 126, 152
	SBE 37-SM MicroCAT recorder	20, 88.5, 186
Salinity	Seabird 16plus SeaCAT recorder	30
	SBE 37-SM MicroCAT recorder	20, 88.5, 186

Supplement 2. Potential influences of Atlantic Water (AW) advection during the study

We used the temperature and salinity data measured in the study and identified different water masses that prevailed in Billefjorden. However, we didn't find any signatures of AW in during the study (see Fig. S1 below). Other investigations conducted in Billefjorden in the same period (e.g. Bailey 2010, Grigor et al. 2014) have also suggested that the influence of AW advection was negligible.

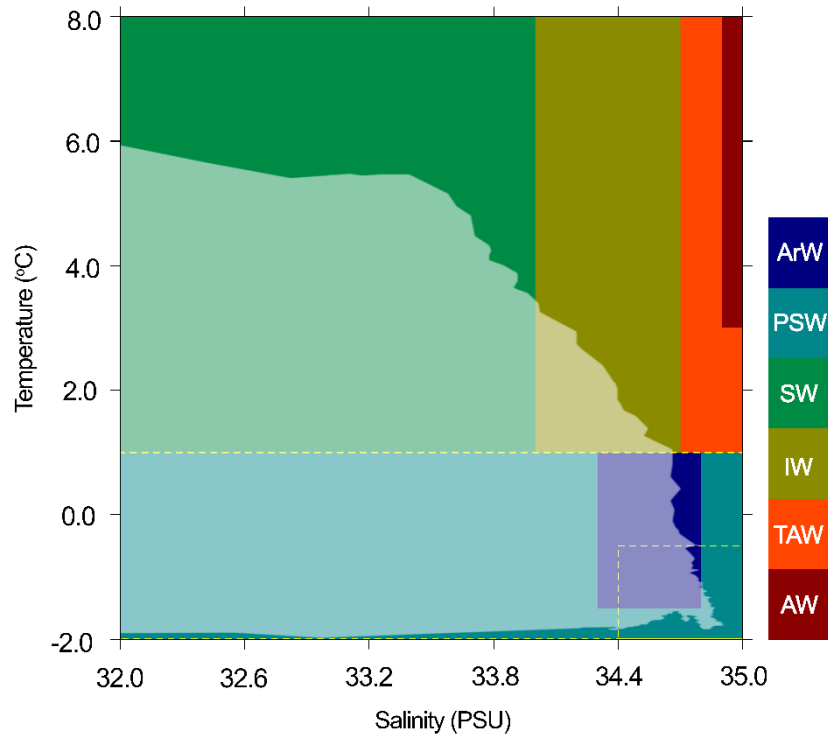


Fig. S1. The range of temperature and salinity measurements recorded in the study (opaque white polygon), and their water mass associations (colored polygons). The abscissa is cropped at 32 PSU. ArW: Arctic water, PSW: polar surface water, SW: surface water, IW: intermediate water, (T)AW: (transformed) Atlantic water. Dashed line: local water (LW, above), and winter cooled water (WCW, below). Water mass associations were adopted from Swift (1986), Hopkins (1991), Svendsen et al. (2002), and Nilsen et al. (2008)

Supplement 3. The PL based separation of *Calanus* taxa

We observed three components in CV and adult female *Calanus* PL distributions (Fig. S2a, b), and two components in that of adult males (Fig. S2c). Although we fitted normal distribution models for all the components, the fraction of the *Calanus* community represented by the smallest component (*Calanus* sp. in Table 2 in the text) was not used in the analyses because the PL boundaries separating it from the other two did not match any literature in Table S3, and its abundance was extremely low. These may be mis-staged smaller copepodites.

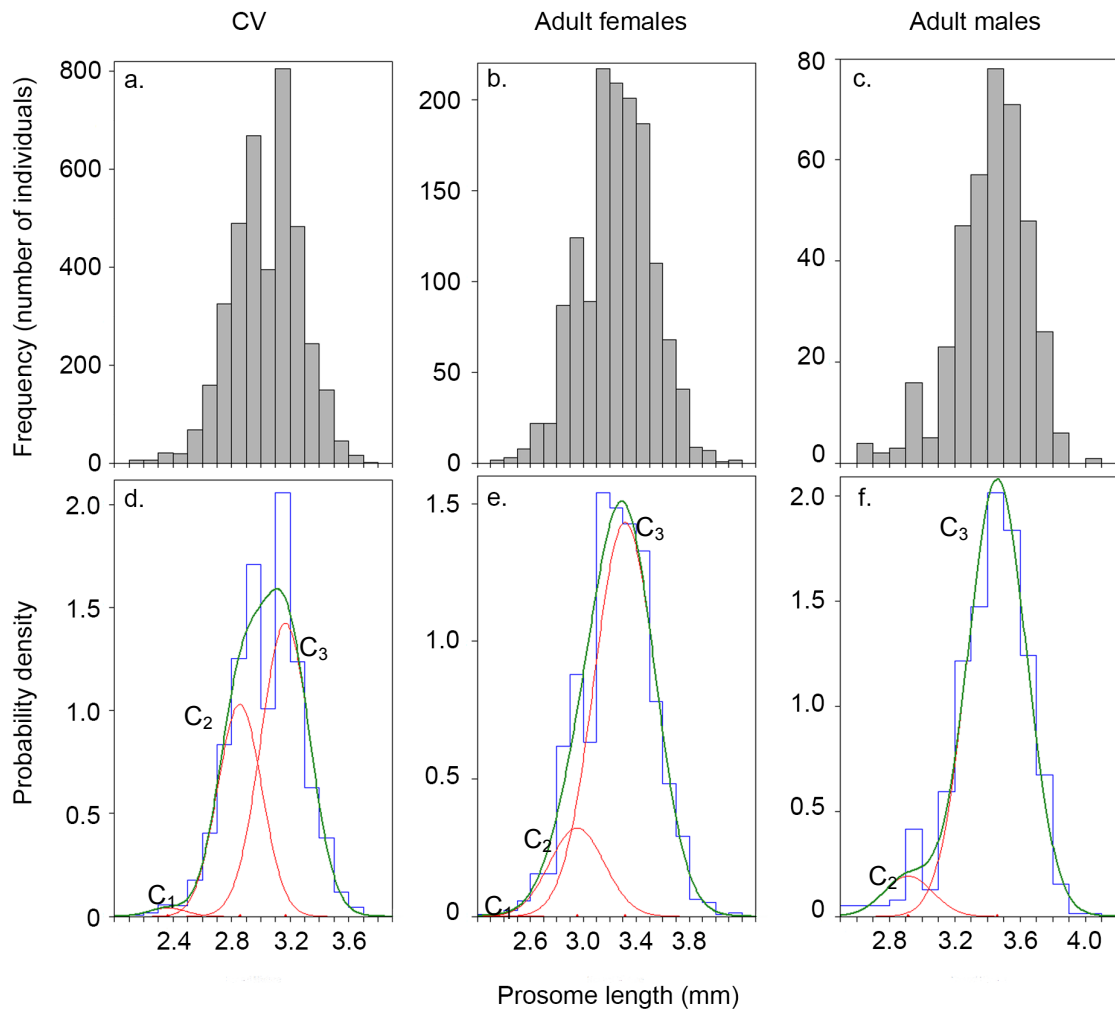


Fig. S2. Overlapped components within the PL distributions of *Calanus* spp. (a–c), and normal distribution models fitted to represent each component (d–f). The means of each fitted normal component model are presented by red ticks on the abscissae of the bottom panels

Table S2. The % overlap between the fitted normal components (C_i) in the *Calanus* spp. PL distributions (see Fig. S2 for reference) estimated by numerical integration

Component	C_1 & C_2	C_1 & C_3	C_2 & C_3
CV	6.36	1.26	33.17
Adult females	12.20	0.00	39.97
Adult males	–	–	10.12

Table S3. PL boundaries (mm) used in some high latitude investigations for separation of coexisting *C. finmarchicus* (CF) and *C. glacialis* (CG) populations. These PL boundaries are comparable with those derived in the present investigation (cf. Table 3 in the text). However, note that the mesh widths of the nets/filters used in these investigations are lesser (50–300 μm) than that of our investigation (1000 μm)

Authors	CV		Adult females		Adult males	
	CF	CG	CF	CG	CF	CG
Jaschnov (1972)	–	–	2.20–3.00	3.60–4.50	–	–
Hirche (1991)	< 3.10	> 3.10	< 3.10	> 3.20	–	–
Unstad and Tande (1991)	< 3.00	3.00–3.40	< 3.20	3.20–4.50	–	–
Koszteyn and Kwasniewski (1992)	< 3.05	3.05–3.95	2.85–3.00	3.50–4.40	–	–
Hirche et al. (1994)	1.95–3.05	2.95–3.90	2.35–3.20	3.20–4.60	–	–
Madsen et al. (2001)	1.75–2.70	2.73–3.90	< 3.00	> 3.00	–	–
Kwasniewski et al. (2003)	< 2.90	\geq 2.90	< 3.20	\geq 3.20	–	–
Daase and Eiane (2007)	< 2.94	> 2.94	< 3.24	> 3.24	–	–
Hirche and Kosobokova (2011)	1.70–2.85	2.90–3.50	2.90–3.15	3.20–4.60	1.85–2.90	2.95–3.60

Supplement 4. The TL based separation of *P. elegans* size groups

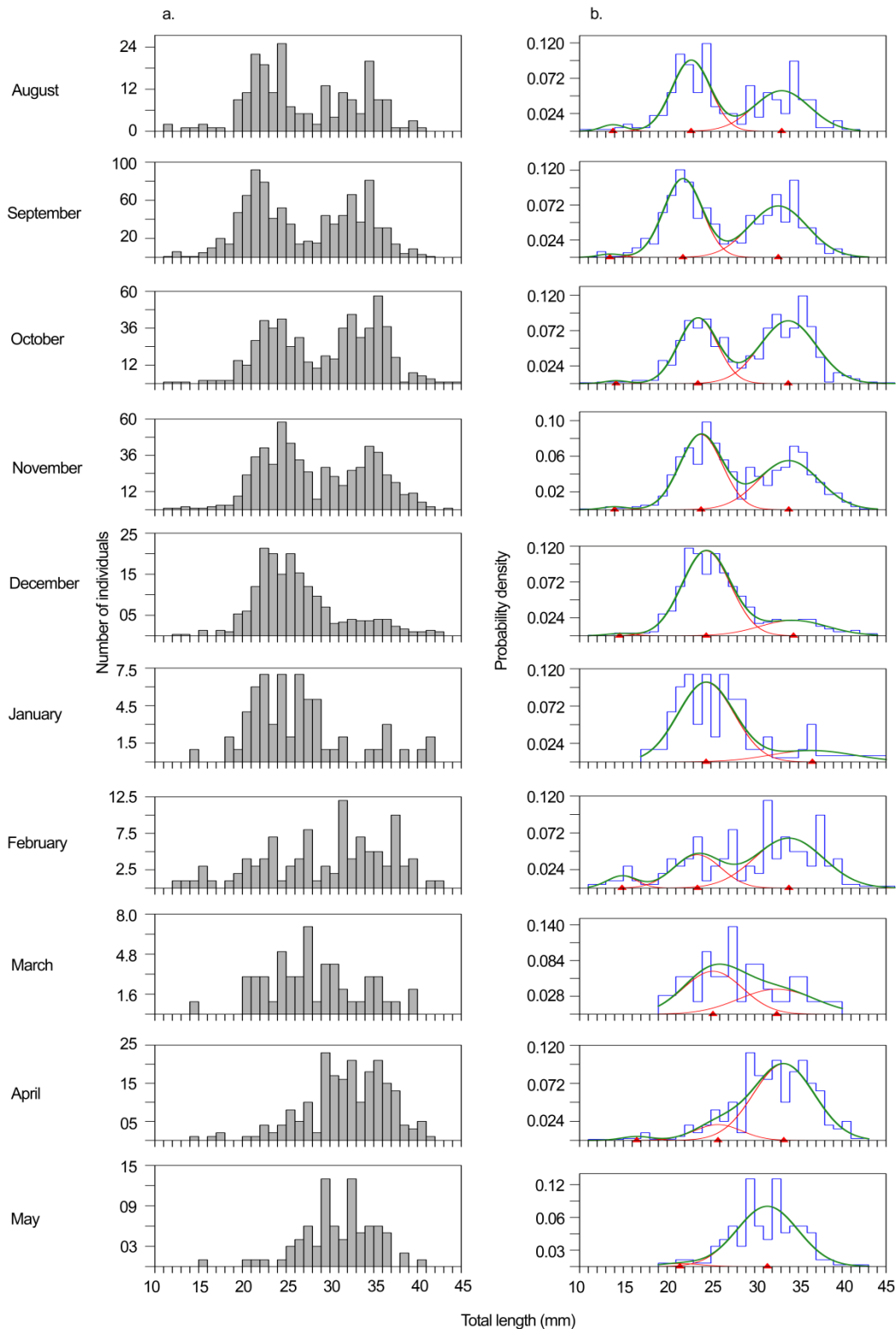


Fig. S3. Overlapped components within monthly TL distributions of *P. elegans* (a), and normal distribution models fitted to represent each component (b). The mean PLs of each fitted component are presented as red ticks on abscissae of the right panels

Table S4. The number individuals (n) used in the length frequency analysis, and the chi-square statistics (χ^2) with the degrees of freedom (df) of the fitted component distribution models for monthly TL distributions of *P. elegans* (see Fig. S3 for reference). The monthly mean TLs (mm) of each size group is given in the three right columns

Month	n	df	χ^2	Mean TL \pm SD		
				G ₀	G ₁	G ₂
August	210	21	47.55***	14.16 \pm 1.83	23.18 \pm 2.04	33.62 \pm 3.04
September	915	24	91.37***	13.22 \pm 0.83	22.21 \pm 2.27	33.30 \pm 3.01
October	574	26	63.52***	14.20 \pm 1.79	23.79 \pm 2.19	34.28 \pm 3.11
November	589	25	42.89**	14.00 \pm 1.41	24.27 \pm 2.38	34.60 \pm 3.08
December	546	22	20.54	15.17 \pm 1.33	24.97 \pm 2.70	35.60 \pm 3.19
January	63	16	17.95	–	24.69 \pm 3.31	37.27 \pm 3.52
February	105	22	33.72*	15.28 \pm 1.38	23.45 \pm 2.31	34.30 \pm 3.86
March	52	15	14.25	–	24.87 \pm 3.11	33.77 \pm 3.24
April	208	18	36.26**	17.00 \pm 1.41	25.04 \pm 1.72	33.84 \pm 3.33
May	82	14	26.07*	–	19.67 \pm 2.38	31.92 \pm 3.59

* $p < 0.05$, ** $p < 0.01$, *** $p < 0.001$

A note on the *P. elegans* size groups

The mean TLs of each size group we derived for *P. elegans* from the length frequency analysis matched those described by Grigor et al. (2014) for this species in this fjord following the same time series. We termed the three size groups as G₀, G₁, and G₂ in comparison with the cohorts 0, 1, and 2 of their investigation. As a WP-2 net was used in their study, they captured higher numbers of G₀ (cohort 0) individuals. See the above work for a detailed account of the population dynamics of this species.

Supplement 5. The vertical distribution index: additional data

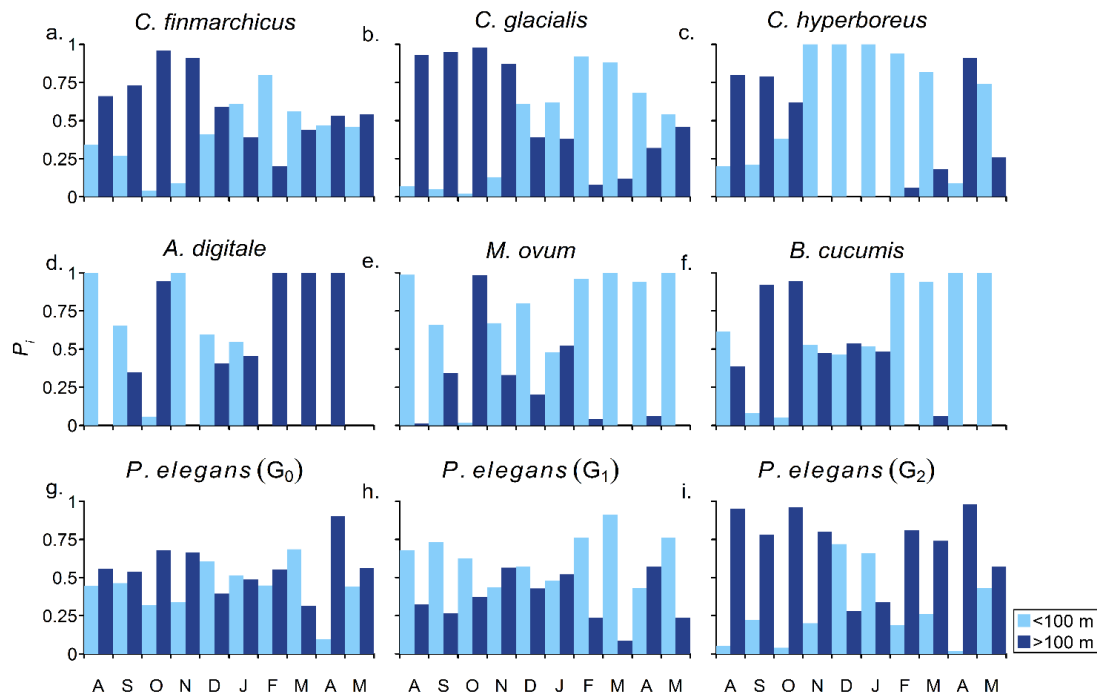


Fig. S4. Proportion of the population (P_i of dominant taxa) distributed in the two vertical regions of the water column during the study. For a given species in a given month, the difference between its population proportions of the shallower region and the deeper region was calculated as the vertical distribution index (V ; See Table S5 below).

Table S5. Seasonal variability in vertical distribution index (V) of the dominant zooplankton taxa during the study. V ranges from -1 to 1, in which the former represents the entire population distributed in the shallower region, and latter represents the opposite scenario. *A. digitale* was not captured to compute its V in May

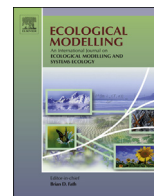
Species	Aug	Sep	Oct	Nov	Dec	Jan	Feb	Mar	Apr	May
<i>C. finmarchicus</i>	-0.32	-0.46	-0.91	-0.82	-0.19	0.21	0.59	0.12	-0.06	-0.08
<i>C. glacialis</i>	-0.86	-0.90	-0.95	-0.73	0.22	0.23	0.84	0.77	0.35	0.09
<i>C. hyperboreus</i>	-0.59	-0.57	-0.24	1.00	1.00	1.00	0.89	0.64	-0.81	0.48
<i>Calanus</i> spp.	-0.77	-0.83	-0.93	-0.73	0.13	0.23	0.79	0.64	0.26	0.04
<i>A. digitale</i>	1.00	0.31	-0.89	1.00	0.19	0.09	-1.00	-1.00	-1.00	–
<i>M. ovum</i>	0.97	0.31	-0.97	0.34	0.60	-0.04	1.00	1.00	0.88	1.00
<i>B. cucumis</i>	0.23	-0.84	-0.89	0.06	-0.07	0.03	1.00	0.88	1.00	1.00
<i>P. elegans</i> (G_0)	-0.11	-0.07	-0.36	-0.32	0.21	0.03	-0.10	0.37	-0.81	-0.12
<i>P. elegans</i> (G_1)	0.35	0.47	0.25	-0.13	0.14	-0.04	0.52	0.82	-0.14	0.52
<i>P. elegans</i> (G_2)	-0.90	-0.56	-0.92	-0.60	0.44	0.32	-0.63	-0.48	-0.96	-0.15

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Paper-II

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A high-resolution modeling study on diel and seasonal vertical migrations of high-latitude copepods



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ABSTRACT

Despite diel and seasonal vertical migrations (DVM and SVM) of high-latitude zooplankton have been studied since the late-19th century, questions still remain about the influence of environmental seasonality on vertical migration, and the combined influence of DVM and SVM on zooplankton fitness. Toward addressing these, we developed a model for simulating DVM and SVM of high-latitude herbivorous copepods in high spatio-temporal resolution. In the model, a unique timing and amplitude of DVM and SVM and its ontogenetic trajectory were defined as a vertical strategy. Growth, survival and reproductive performances of numerous vertical strategies hardwired to copepods spawned in different times of the year were assessed by a fitness estimate, which was heuristically maximized by a Genetic Algorithm to derive the optimal vertical strategy for a given model environment. The modelled food concentration, temperature and visual predation risk had a significant influence on the observed vertical strategies. Under low visual predation risk, DVM was less pronounced, and SVM and reproduction occurred earlier in the season, where capital breeding played a significant role. Reproduction was delayed by higher visual predation risk, and copepods that spawned later in the season used the higher food concentrations and temperatures to attain higher growth, which was efficiently traded off for survival through DVM. Consequently, the timing of SVM did not change much from that predicted under lower visual predation risk, but the body and reserve sizes of overwintering stages and the importance of capital breeding diminished. Altogether, these findings emphasize the significance of DVM in environments with elevated visual predation risk and shows its contrasting influence on the phenology of reproduction and SVM, and moreover highlights the importance of conducting field and modeling work to study these migratory strategies in concert.

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1. Introduction

Vertical migration is a common behavior of many zooplankton taxa. Based on the periodicity, vertical migrations of high-latitude zooplankton are classified into diel and seasonal components, which have been studied since the late-19th century (reviewed in Russell, 1927; Cushing, 1951; Banse, 1964). The short-term diel vertical migration (DVM) has a periodicity of up to 24 h, and is understood as a strategy that trades off growth potential to reduce the mortality risk imposed by visual predators (Lampert, 1989;

Ohman, 1990; Loose and Dawidowicz, 1994). The long-term seasonal vertical migration (SVM) has a periodicity of up to one year, and reflects adaptations to seasonal extremities of food availability (Head and Harris, 1985; Hind et al., 2000; Bandara et al., 2016), temperature (Hirche, 1991; Astthorsson and Gislason, 2003) and predation risk (Kaartvedt, 1996; Bagøien et al., 2000; Varpe and Fiksen, 2010). In either case, since both DVM and SVM can alter feeding, growth, survival and reproduction, and ultimately affect fitness (Aidley, 1981; Alerstam et al., 2003; Cresswell et al., 2011; Litchman et al., 2013), these migratory strategies are termed vertical strategies (Bandara et al., 2016).

Empirical knowledge on zooplankton vertical strategies largely comes from studying the dynamic vertical positioning of populations in a water column, and are often rather coarse in spatial (vertical) and temporal resolution (Pearre, 1979). This can undermine the key concept that such migrations are individual responses

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Table 1
Some endogenous and exogenous cues that are believed to proximately or ultimately regulate diel and seasonal vertical migrations of marine and freshwater zooplankton. Literature do not come from an exhaustive review and only serve as examples.

Cue	DVM	SVM
Temperature	McLaren (1963), Enright (1977)	Hirche (1991), Heath and Jónasdóttir (1999), Astthorsson and Gislason (2003)
Light (absolute or relative irradiance from sun, moon, stars, or aurora borealis, photoperiod, spectral quality, polarization etc.)	Clarke (1933), Gliwicz (1986), Frank and Widder (1997), Berge et al. (2009), Båtnes et al. (2015), Cohen et al. (2015), Bianchi and Mislan (2016), Bozman et al. (2017)	Sømme (1934), Ussing (1938), Miller et al. (1991)
Dissolved oxygen	Devol (1981), Bianchi et al. (2013)	–
Water depth, transparency and UV radiation	Rhode et al. (2001), Williamson et al. (2011), Ekvall et al. (2015)	Dupont and Aksnes (2012)
Tides, currents and advective transport	Hardy (1935), Wroblewski (1982), Kimmerer and McKinnon (1987)	Berge et al. (2012), Irigoien (2004)
Food availability	Hardy and Gunther (1935), Huntley and Brooks (1982), George (1983), Johnsen and Jakobsen (1987)	Herman (1983), Hind et al. (2000), Head and Harris (1985), Bandara et al. (2016)
Visual and tactile predation	Zaret and Suffern (1976), Iwasa (1982), Ohman (1990), Bollens et al. (1992), Loose and Dawidowicz (1994)	Kaartvedt (1996), Kaartvedt (2000), Dale et al. (1999), Bagoien et al. (2000), Varpe and Fiksen (2010)
Body size, ontogeny and pigmentation	Zaret and Kerfoot (1975), Uye et al. (1990), Hays et al. (1994), Dale and Kaartvedt (2000)	Østvedt (1955), Hind et al. (2000)
Nutritional state and lipid reserves	Fiksen and Carlotti (1998), Sekino and Yamamura (1999)	Visser and Jónasdóttir (1999), Thorisson (2006)
Endogenous rhythms and internal biological clocks	Cohen and Forward (2009), van Haren and Compton (2013)	Carlisle and Pitman (1961), Miller et al. (1991), Hirche (1996)

to certain cues or stimuli and not a property of the population (Zink, 2002), and may complicate the understanding of the relationships between vertical strategies and environmental variables (see Table 1 for examples). Moreover, since diel and seasonal vertical migrations occur on different spatial and temporal scales, studying these migrations together in the field in adequate resolution remains a major challenge. Although novel optical and acoustic methods of *in-situ* observation offer a solution to some of these problems (e.g. Basedow et al., 2010; Sainmont et al., 2014b; Bozman et al., 2017; Darnis et al., 2017), long-term deployment and accurately resolving the identity of the migrants remain as key challenges.

Mechanistic models offer an alternative means of studying zooplankton vertical strategies in higher resolution. Models related to DVM usually encompass the highest spatial (≤ 1 m), temporal (≤ 1 h) and biological (=individual) resolution (e.g. Fiksen and Giske, 1995; Eiane and Parisi, 2001; Liu et al., 2003; Burrows and Tarling, 2004; Hansen and Visser, 2016). Models related to SVM and diapause (i.e. hibernation in deeper waters, e.g. Hirche, 1996) encompass the same biological resolution, but are usually coarse in spatio-temporal resolution. Here, the time intervals range from 1 h to 1d and vertical spatial elements are usually resolved to either absolute depth units (e.g. 1 m bins) or segregated habitats (e.g. Fiksen and Carlotti, 1998; Miller et al., 1998; Hind et al., 2000; Ji, 2011; Ji et al., 2012; Sainmont et al., 2015; Banas et al., 2016). The choice of a coarser spatio-temporal resolution of these models reflects the broader space and time scales at which the SVM and diapause occurs. This contrasting spatio-temporal scale makes it difficult to harbor lifetime dynamics of DVM to be simulated in SVM models without significantly increasing computer time. Consequently, most models that simulate SVM tend to either fully (e.g. Hind et al., 2000) or partly (i.e. of younger developmental stages, e.g. Fiksen and Carlotti, 1998) disregard DVM. However, the validity of such simplifications are questionable, given the geographically and taxonomically widespread nature of zooplankton DVM behavior and its ontogenetic patterns (Huntley and Brooks, 1982; Huang et al., 1993; Osgood and Frost, 1994; Hays, 1995). It is thus interesting to investigate whether the extra biological information resulting from modeling DVM and SVM in concert is a worthy trade-off for the elevated computer time. If so, such models may lead to improvements of the current understanding about how environmental seasonal-

Table 2

Evolvable (soft) parameters optimized in the model. The first six are proxies that define the vertical strategy. Vertical strategies of copepods spawned in different times of the year (t_B) are optimized using the GA.

Term	Definition	Range	Interval	Unit
α	Light sensitivity parameter	$0-I_{\max}^a$	1	$\mu\text{mol m}^{-2} \text{s}^{-1}$
β	Size-specificity of light sensitivity parameter	0–10	1	dim.less
γ	Growth allocation parameter	0–1	0.01	dim.less
δ	Seasonal descent parameter	0–1	0.01	dim.less
ζ	Overwintering depth	1–500	10	m
ε	Seasonal ascent parameter	0–1	0.01	dim.less
t_B	Time of birth ^b	1–8760	1	h

^a The upper limit of α changes with the maximum surface irradiance of the model environment, i.e. $I_{\max} = 1500 \mu\text{mol m}^{-2} \text{s}^{-1}$ for Environment-L, $1300 \mu\text{mol m}^{-2} \text{s}^{-1}$ for Environment-M and $1100 \mu\text{mol m}^{-2} \text{s}^{-1}$ for Environment-H (cf. Fig. 1).

^b Time of being spawned.

ity shapes up vertical strategies, and the means of which the latter influences life histories of high latitude zooplankton.

In this study, we present a model of zooplankton vertical strategies. The model operates in a high-latitude setting and simulates both DVM and SVM of a herbivorous copepod with an annual life cycle in high spatial (vertical) and temporal resolution. Using this model, we aim to investigate the influence of environmental variables on vertical strategies, and how vertical strategies affect fitness and phenology in seasonal environments. We further discuss how short-term behavior (DVM) influences and interacts in the longer-term and shape-up different life history components of copepod strategies.

2. Materials and methods

Although the model is not strictly individual-based, it is described following the Overview, Design concepts and Details (ODD) protocol (Grimm et al., 2006, 2010) to improve reproducibility.

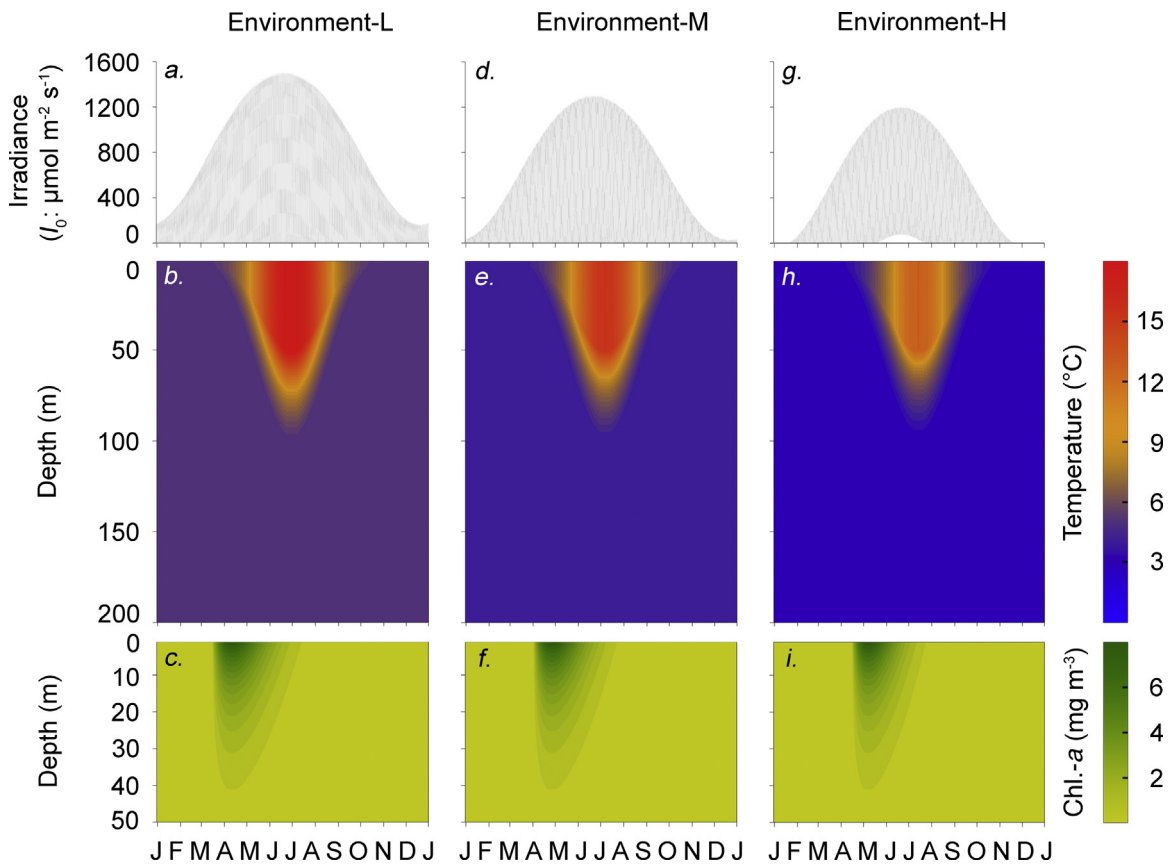


Fig. 1. The modelled dynamics of irradiance incident on the sea surface (hourly estimates; a, d, g), temperature (b, e, h) and food availability (c, f, i, expressed as Chlorophyll-a biomass) in the three model environments. See Appendix A1 in Supplementary material for a detailed comparison.

2.1. Purpose

The purpose of the model is to investigate the bottom-up and top-down influences of environmental variability (i.e. irradiance, temperature, food-availability and predation risk) on vertical strategies of a high-latitude herbivorous copepod, and to understand the influences of vertical strategies on its fitness and phenology.

2.2. Entities, state variables and scales

The model consists of three entities: vertical strategies, model organism and the model environment. Vertical strategies define the timing, amplitude and the ontogenetic trajectories of DVM and SVM, and are described using six evolvable (soft) parameters (Table 2). These are hardwired to the model organism, i.e. copepods spawned in different times of the year.

The model organism is a hypothetical herbivorous semelparous female copepod (hereafter, the copepod) with an annual life cycle that resembles *Calanus finmarchicus* and *C. glacialis* in terms of body size, behavior and life history strategies (Conover, 1988). These two species often dominate the copepod biomass in the North Atlantic and most Eurasian sub-Arctic and Arctic seas and shelves (Falk-Petersen et al., 2009). Their life cycle consists of an embryonic stage (egg), six naupliar stages (N1–NVI), five copepodite stages (C1–CV) and an adult. Eggs that are released in near-surface waters in the spring usually develop into CIV or CV stages toward the end of the productive season. As further development is typically constrained by the duration of the productive season and seasonal peaks of visual predation risk, CIVs and CVs descend into deeper waters and remain in a state of diapause/dormancy with minimal phys-

iological activity (Hirche, 1996). Overwintering stages ascend to near-surface waters as the primary production commences in the following year, molt into adults and start to reproduce (Conover, 1988; Varpe, 2012). The life cycle of the two species is usually completed within one year in most sub-Arctic and Arctic locations (Falk-Petersen et al., 2009; Daase et al., 2013), within which reside the model environments of this study.

The model runs in three 500-m deep artificial seasonal environments that represent three high-latitude locations along the southern and southeastern coast of Norway (60–70°N). These environments do not point to specific geographic locations, but the modelled environmental dynamics were adopted from field measurements from the above region (Appendix A1 in Supplementary material). The baseline model simulation (hereafter, the basic run) runs in Environment-L, representing the lower end (ca. 60°N) of the selected geographical range. Here, the modelled irradiance, temperature and food availability are highly seasonal and vertically structured (Fig. 1a–c), but are assumed constant between years. The irradiance incident on the sea surface follows the global clear-sky horizontal irradiance model of Robledo and Soler (2000), and peaks at ca. $1500 \mu\text{mol m}^{-2} \text{s}^{-1}$ (Fig. 1a, Appendix A1 in Supplementary material). The sea surface temperature reaches a maximum of 18°C in the summer (e.g. Bagøien et al., 2000), and distributes evenly in the surface mixed layer (Fig. 1b). Below this, the temperature decreases with depth and converges to a minimum of 4°C at ca. 100 m (e.g. Ingvaldsen and Loeng, 2009). The pelagic productive season extends ca. 180 days, with a chlorophyll-a peak at 8 mg m^{-3} in mid-April (Fig. 1c; Sakshaug et al., 2009; Daase et al., 2013). We manipulated the environmental parameters of Environment-L to formulate two additional artificial environments: Environment-M (ca. 65°N, Fig. 1d–f) and Environment-H (ca. 70°N, Fig. 1g–i),

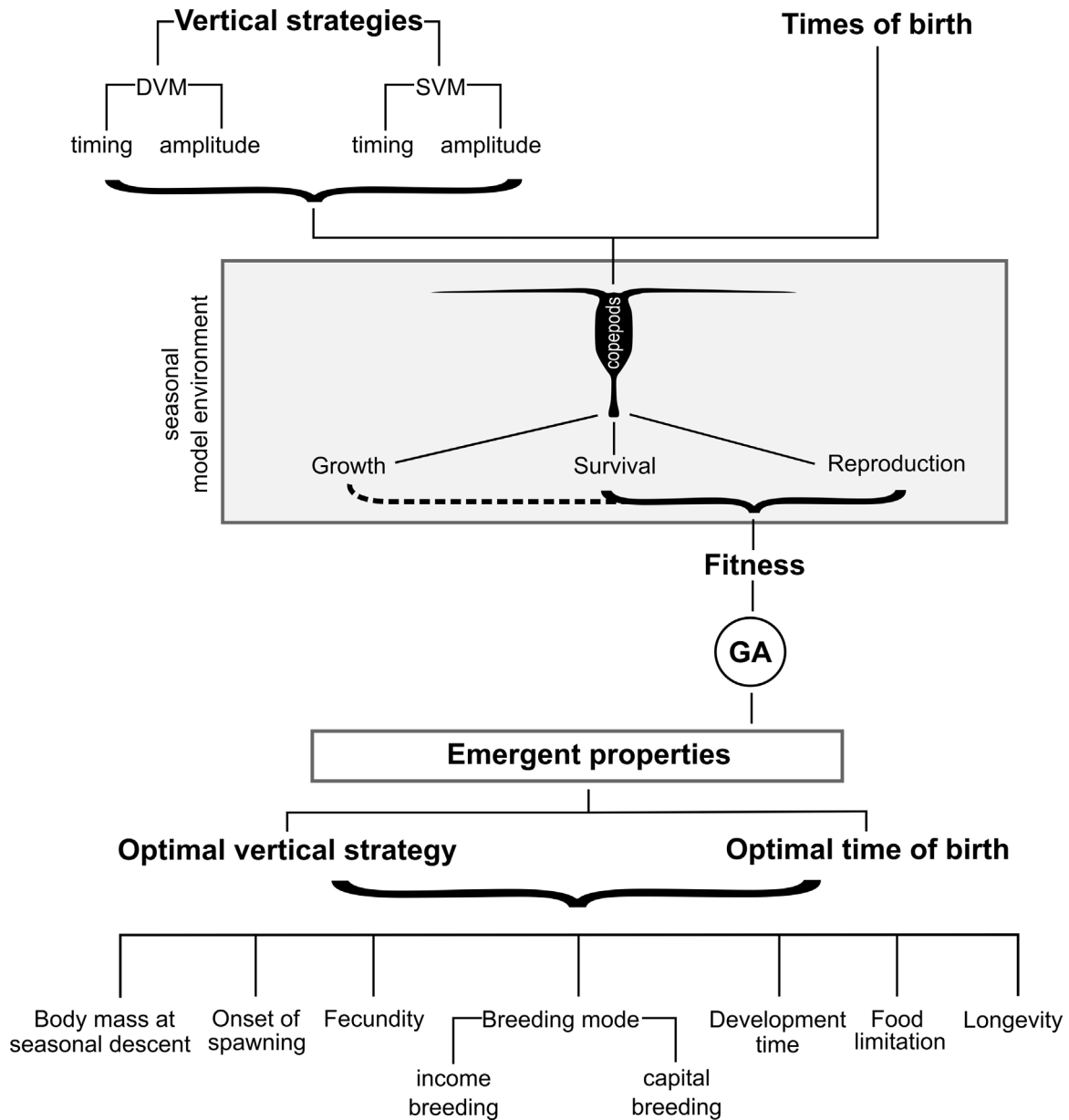


Fig. 2. The model overview. Vertical strategies that define the timing and amplitude of DVM and SVM are hardwired to copepods born in different times of the year. Growth, survival and reproduction of these copepods are simulated in a seasonal environment to derive a fitness estimate that is heuristically maximized by the GA to derive the optimal vertical strategy, time of birth and several associated life history traits emerging from the model. Dashed line represents the indirect dependency of the fitness estimate on growth (Section 2.6.4).

representing the mid-point and the higher end of the selected geographical range (Appendix A1 in Supplementary material).

Copepods are characterized by six states: vertical location (depth), structural body mass, energetic reserve, reproductive output (fecundity), survivorship and developmental stage. The model has a temporal coverage of an annual cycle and a unidimensional (vertical) spatial coverage of 500 m. The time and space consist of 1 h and 1 m discreet intervals.

2.3. Process overview and scheduling

At each timestep, the model follows vertical strategies hardwired to copepods born in different times of the year and simulates their growth, survival and reproduction. State variables are updated simultaneously. Vertical strategies are evaluated using a fitness function based on the expected survival and reproductive per-

formances. The fitness is heuristically maximized using a Genetic Algorithm (GA, Holland, 1975) to estimate the optimal vertical strategy and optimal time of birth for a given set of environmental conditions (Fig. 2).

2.4. Design concepts

2.4.1. Basic principles

The high spatial and temporal resolution implemented in the model allow both DVM and SVM to be simulated over the entire annual life cycle of the copepod. Carlotti and Wolf (1998) have implemented a similar construct, but the SVM of their model was constrained by fixing the timing of ascent and descent to match the field observations of the region of interest. In contrast, the timing and the amplitude of DVM and SVM of our model are flexible and allowed to evolve according to the environmental conditions. To

Table 3

Emergent properties of the model. The timing and amplitude and of DVM and SVM altogether forms the vertical strategy of a copepod.

Trait/attribute	Units	Description
Time of birth	Day of the year	Time of being spawned
Surface time	h	Unified estimate representing the timing of DVM, i.e. the stage-specific mean no. of hours per day occupied in waters with highest growth potential (usually the surface waters)
Amplitude of diel vertical migration	m	The vertical range corresponding to the above
Time of seasonal descent and ascent	Day of the year	Separate estimates representing the timing of SVM (ascent and descent)
Amplitude of seasonal vertical migration	m	Overwintering depth
Body mass at seasonal descent	μg C	Structural and energetic reserve mass at the onset of diapause
Onset of egg production	Day of the year	–
Fecundity	No. of eggs	No. of eggs produced during the lifetime
Breeding mode index	dim.less	Proportion of capital breeding eggs (0 = pure income breeding, 1 = pure capital breeding)
Food limitation index	dim.less	Stage-specific total no. hours with food-limited growth (Eq. (3)) as a fraction of stage duration (0 = no food limitation, 1 = total food limitation)
Development time	d	From egg to a given stage
Longevity	d	Duration of the life cycle, from birth to death

achieve this level of flexibility, we used multiple evolvable proxies to represent vertical migration (Table 2). This resulted in a complex seven-dimensional optimization problem that can be efficiently solved using heuristic techniques (Zanakis and Evans, 1981). As evolutionary algorithms provide an efficient means of solving multi-dimensional optimization problems (Deb, 2001; Eiben and Smith, 2003), we used a GA as the optimization platform of this model. Further, to increase the precision of the evolvable parameters and that of the behavioral strategies and life history traits ensued (Fig. 2), we used a GA variant with floating point representation (i.e. a Real-Coded Genetic Algorithm, Davis, 1989; Lucasius and Kateman, 1989; Herrera et al., 1998).

The strategy-oriented construct of this model contrasts classic individual-based models of zooplankton life history and behavior in two main ways: first, trading off of biological resolution (strategies vs. individuals) to accommodate higher spatio-temporal resolution, and second, the lack of population-level responses such as density dependence. As a result, modelled vertical strategies do not interact with each other and show no quantitative feedbacks with the model environment (e.g. impact of grazing on food concentration and duration of the productive season).

2.4.2. Emergence

The behavioral strategies and life history traits emerging from the model are presented in Fig. 2 and described in Table 3.

2.4.3. Adaptation and sensing

Copepods are sensitive to their internal states (i.e. structural body mass, mass of the energetic reserve and developmental stage) and external stimuli (i.e. irradiance, temperature, food concentration and depth). Altogether, these determine the size- or stage-specific patterns of growth, metabolism, reproduction and vertical behavior (Section 2.6).

2.4.4. Objectives

The model uses a fitness estimate that evaluates the expected reproduction and survival performances rendered by different vertical strategies (Section 2.6.4).

2.4.5. Prediction and stochasticity

The vertical search pattern of copepod behavior is based on a semi-stochastic predictive algorithm (Section 2.6.2.2 and Appendix A2 in Supplementary material). Stochasticity plays a central role in the model initialization (Section 2.5) and selection, recombination and mutation operators of the GA (Section 2.6.4).

2.4.6. Observations

For a given model environment, the model produces heuristic estimates of the optimal vertical strategy and optimal time of birth, along with a range of associated life history traits (Fig. 2, Table 3).

2.5. Initialization

The model initializes with seeding of $N (=10^6)$ eggs at random times of the year to random depths (<50 m) of the water column. Each seed represents an embryonic stage of a copepod with a specific vertical strategy, which is determined by randomly assigning values to the evolvable proxies. The ranges (bounds) and resolutions of these proxies are listed in Table 2.

2.6. Submodels

2.6.1. Growth and development

We modelled somatic growth in Carbon units (μg C) according to the growth model of Huntley and Boyd (1984) (Eqs. (1)–(8) below), using a Chlorophyll-a/C ratio of 0.030 (Båmstedt et al., 1991; Sakshaug et al., 2009). This growth sub-model was used due to its simplicity and general applicability, which are shown by its utility to model several different copepod taxa with varying body sizes representing a wide range of geographical locations (e.g. Robinson, 1994; Fiksen and Giske, 1995; Roman et al., 2000). Definitions and units of all the terms described henceforth are listed in Table 4.

For ambient food concentrations (F : μg C ml⁻¹) above a specific saturation concentration (f), growth is food-independent, and occurs at a maximum rate (G_T : μg C ind⁻¹ h⁻¹) dependent only on the ambient temperature (T) as;

$$(G_T)_{i,t,z} = (G'_{max})_{t,z} \cdot W_{i,t} \quad (1)$$

Here, i represents individual, t time and z is depth, where G'_{max} (μg C mg dry mass h⁻¹) is the maximum temperature-dependent mass-specific growth rate, assuming a Carbon: dry body mass (W , mg) ratio of 0.40 (Huntley and Boyd, 1984), defined as;

$$(G'_{max})_{t,z} = 0.903 \cdot \exp(0.110 \cdot T_{t,z}) \quad (2)$$

If the ambient food concentration drops below the saturation concentration, the growth occurs at a rate limited by food availability (G_F) as;

$$(G_F)_{i,t,z} = a \cdot b_{t,z} \cdot W_{i,t}^{n_{t,z}} \cdot F_{t,z} - k \cdot W_{i,t}^{m_{t,z}} \quad (3)$$

Table 4
Definitions, values and units of the terms used in the model.

Term	Definition	Value/formula	Units
a	Assimilation coefficient	0.70 ^b	–
$b_{t,z}$	Clearance coefficient	Eq. (4) ^a	ml mg dry mass h ⁻¹
E	Egg development parameter	717 ^{e,f}	–
$f_{i,t,z}$	Saturation food concentration	Eq. (8) ^a	μg C ml ⁻¹
$F_{t,z}$	Ambient food concentration	Section 2.2	μg C ml ⁻¹
$(G'_{max})_{t,z}$	Maximum mass-specific growth rate	Eq. (2) ^a	μg C mg dry mass h ⁻¹
$(G_F)_{i,t,z}$	Food-limited growth rate	Eq. (3) ^a	μg C ind ⁻¹ h ⁻¹
$(G_T)_{i,t,z}$	Non food-limited growth rate	Eq. (1) ^a	μg C ind ⁻¹ h ⁻¹
$H_{i,t,z}$	Survivorship	Eq. (15)	–
i	Individual	–	–
$I'_{t,z}$	Remapped $I_{t,z}$	$0.9 \geq I'_{t,z} \geq 0.1$	–
$I_{t,0}$	Irradiance incident on sea surface	Appendix A1 in Supplementary material ^c	μmol m ⁻² s ⁻¹
$I_{t,z}$	Downwelling irradiance at depth z	Eq. (9)	μmol m ⁻² s ⁻¹
j	Developmental stage	0–12 (Egg–Adult)	–
$K_{i,t}$	Scalar for visual predation risk	$1 > K > 0$	–
$k_{t,z}$	Respiratory coefficient	Eq. (5) ^a	μg C mg dry mass h ⁻¹
$(M_n)_{t,z}$	Non-visual predation risk	Section 2.6.2.1	–
$(M_s)_{i,t,z}$	Starvation risk	Eq. (12)	–
$m_{t,z}$	Exponent (respiration)	Eq. (6) ^a	–
$(M_v)_{i,t,z}$	Visual predation risk	Eq. (10)	–
N	No. of initial seeds	1,000,000	–
$n_{t,z}$	Exponent (clearance)	Eq. (7) ^a	–
R_i	Fecundity	Eq. (13)	no. of eggs
t	Time	1–8760	h
$T_{t,z}$	Ambient temperature	Section 2.2	°C
$U_{i,t}$	Cruising velocity	Eq. (11)	m h ⁻¹
$(W_c)_{i,t}$	Structural mass	–	μg C
W_E	Unit egg mass	0.55 ^d	μg C
$W_{i,t}$	Dry body mass (assuming 40% C)	–	mg
$(W_q)_{i,t}$	Catabolized structural mass (proportion to the maximum lifetime structural mass)	$0 \geq W_q \geq 0.5$	–
$(W_R)_{i,t,z}$	Matter allocated for egg production	–	μg C
$(W_s)_{i,t}$	Storage (energetic reserve) mass	–	μg C
W_x	Stage-specific critical molting mass	Table 5	μg C
z	Depth	0–500	m
Φ	Termination condition of the RCGA	Section 2.6.4	–
ψ	Light attenuation coefficient	0.06 ^g	m ⁻¹
ω	Parameter for weighing fitness	0 or 1	–
Ω_i	Fitness	Eq. (14)	–

^a Huntley and Boyd (1984).

^b Fiksen and Giske (1995).

^c Robledo and Soler (2000).

^d Calculated from Salzen (1956).

^e Campbell et al. (2001).

^f Ji et al. (2012).

^g Eiane and Parisi (2001).

where two terms of the right-hand side of the equation refer to the assimilation and respiratory rates respectively. The assimilation coefficient (a) is assumed to be constant (Table 4), but Huntley and Boyd (1984) found that the coefficients of clearance (b) and respiration (k), and the exponents (n and m) vary with ambient temperature as;

$$b_{t,z} = 1.777 \cdot \exp(0.234 \cdot T_{t,z}) \quad (4)$$

$$k_{t,z} = 0.375 \cdot \exp(0.0546 \cdot T_{t,z}) \quad (5)$$

$$n_{t,z} = 0.671 \cdot \exp(0.0199 \cdot T_{t,z}) \quad (6)$$

$$m_{t,z} = 0.858 \cdot \exp(-0.008 \cdot T_{t,z}) \quad (7)$$

At the point where F reaches f , Eqs. (1) and (3) balance out, and the f becomes;

$$f_{i,t,z} = \frac{\left[(G'_{max})_{t,z} \cdot W_{i,t} + k_{t,z} \cdot W_{i,t}^{m_{t,z}} \right]}{a \cdot b_{t,z} \cdot W_{i,t}^{n_{t,z}}} \quad (8)$$

This growth sub-model is not applicable to the first two nauplii stages, which do not feed (Fig. 3a, Marshall and Orr, 1972; Mauchline, 1998). For simplicity, we assumed the growth of NI and NII stages to occur at a temperature-dependent rate (Eqs. (1) and

(2)). The growth of early developmental stages (NI–CIII) is solely allocated to the building up of structural mass (W_c , μg C, Fig. 3a, b and Table 5).

The embryonic development follows a Bělehrádek temperature function (Campbell et al., 2001; Ji et al., 2012). The post-embryonic development (from stage j to $j+1$) occurs only if W_c exceeds a stage-specific critical molting mass (W_x , μg C, Table 5). However, for simplicity, we did not model the dependence of molting process on the physiological state (Nival et al., 1988) and the limitation of growth by the exoskeleton (Mauchline, 1998).

2.6.2. Survival

2.6.2.1. *Predation risk.* Visual (v) and non-visual (n) predators induce mortality, which is estimated as a probability following Eiane and Parisi (2001) as;

$$I_{t,z} = I_{t,0} \cdot \exp(-\psi \cdot z) \quad (9)$$

where I_z and I_0 are irradiance at depth z and surface at a given time, and ψ ($=0.06 \text{ m}^{-1}$) is the attenuation coefficient for downwelling directed irradiance in the water column. We remapped irradiance (I) between 0.1–0.9 (I') so that visual predation risk is not nullified even at the lowest levels of irradiance, and the copepod has some chance of survival even at highest levels of irradiance.

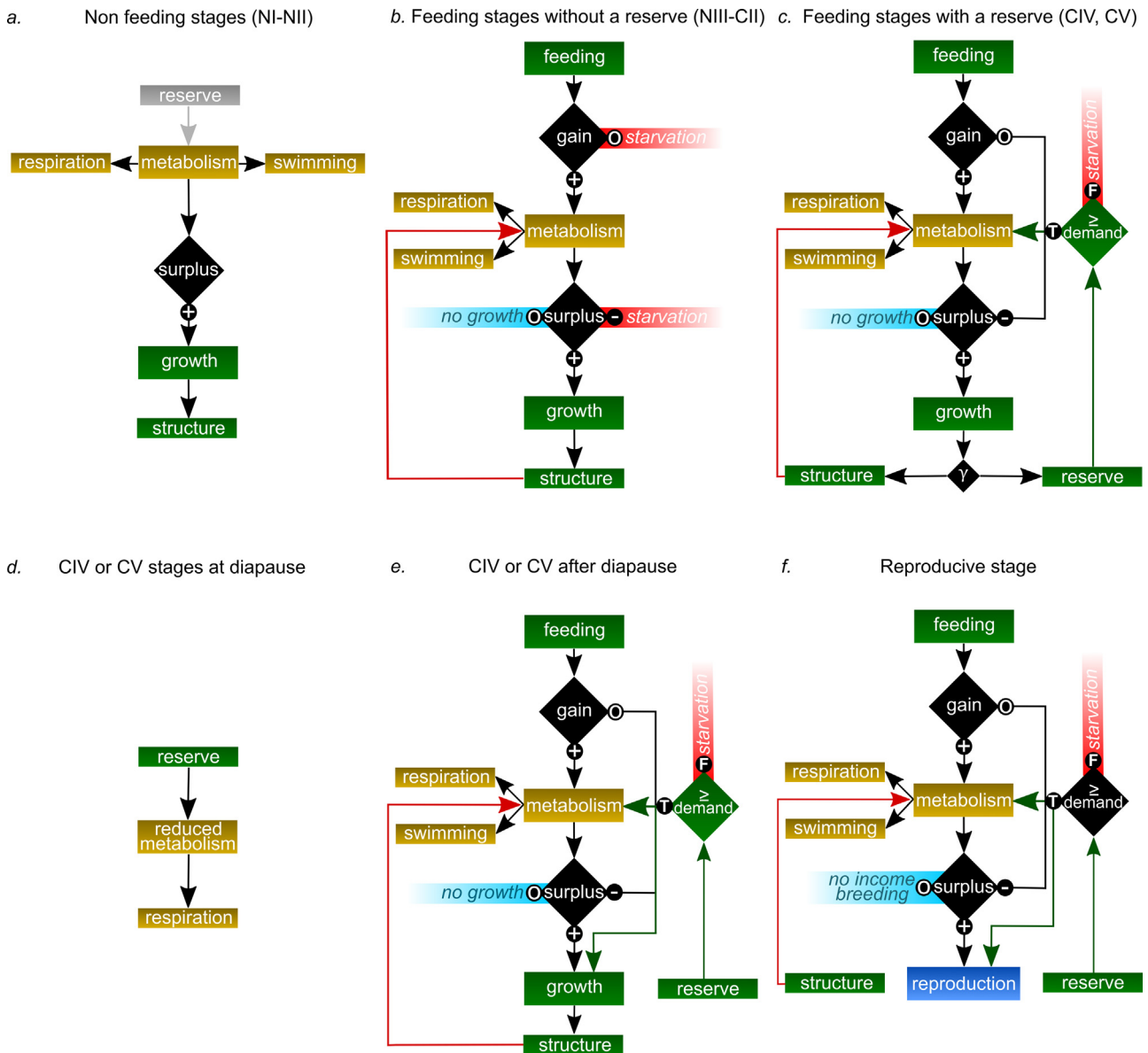


Fig. 3. Simplified physiological pathways modelled in this study. Some life stages are grouped together due to their similarities in energy allocation patterns (a–f). Starvation (highlighted in red) triggers catabolic pathways marked in red. T and F are Boolean values true and false. γ is the growth allocation parameter (Table 2). A comparative summary is given in Table 5.

The detection efficiency of visually orientating planktivores increases with the size of their prey (Brooks and Dodson, 1965; Batty et al., 1990). For simplicity, we modelled the size-dependent visual predation risk using a linear model, assuming that the largest developmental stage is ca. 10 times more vulnerable to visual predators compared to the smallest developmental stage (Fig. 4a, Table 5, De Robertis, 2002). This was implemented using the scalar K ($1 > K > 0$) as;

$$(M_v)_{i,t,z} = I'_{t,z} \cdot K_{i,t} \tag{10}$$

The initial value of K (i.e. K value at the embryonic stage, range = 1×10^{-4} – 1.5×10^{-2}) was decided so that it produces hourly estimates of visual predator-induced mortality.

We assumed the mortality risk caused by non-visual predators (non-visual predation risk, M_n) to be 1% of the maximum visual predation risk and constant over time and depth (Eiane and Parisi, 2001).

2.6.2.2. Diel vertical migration. The copepod may perform DVM to trade off growth potential to minimize the visual predation risk. We used the photoreactive behavior as a proxy to estimate the timing and amplitude of DVM (e.g. Kerfoot, 1970; Carlotti and Wolf, 1998). Here, α , an evolvable light sensitivity parameter (Table 2) was used to define an irradiance threshold above which induces a negative phototactic response in the vertical swimming behavior (Bätnes et al., 2015; Cohen et al., 2015). At any given time, the copepod occupies a depth with an irradiance level ($I_{t,z}$) below α . From a series of possible depth bins that satisfy the $I_{t,z} < \alpha$ condition, we assumed that the copepod searches and occupies the depth that maximizes its growth potential. For simplicity, we further assumed that internal state-dependent factors, such as hunger and satiation have a negligible influence on the modelled DVM. The vertical search pattern was predicted using a biased random walk algorithm (Codling, 2003, Appendix A2 in Supplementary material), assuming that the copepod is neutrally buoyant and vertically moves in the water column at a maximum velocity (hereafter cruising veloc-

Table 5
Developmental stages, their maximum structural body masses (W_x) and stage-specific variability in several biological processes modelled in this study (cf. Fig. 3). Dashes indicate non-applicability.

Stage	W_x (μgC)	Feeding	Structural growth	Energetic Reserve	Respiration	Swimming	Egg production
Egg	0.55	–	–	–	–	–	–
NI	0.55	–	x	–	x	x	–
NII	0.68	–	x	–	x	x	–
NIII	0.91	x	x	–	x	x	–
NIV	1.84	x	x	–	x	x	–
NV	2.72	x	x	–	x	x	–
NVI	3.92	x	x	–	x	x	–
CI	6.01	x	x	–	x	x	–
CII	9.84	x	x	–	x	x	–
CIII	17.58	x	x	–	x	x	–
CIV	36.42	x ^a	x ^d	x	x ^c	x ^d	–
CV	110.03	x ^a	x ^d	x	x ^c	x ^d	–
Adult	332.27	x	–	x ^b	x	x	x

W_x values resemble those published for *C. finmarchicus* and *C. glacialis* by Båmstedt et al. (1991); and Campbell et al. (2001).

^a Feeding ceases during diapause.

^b Does not allocate surplus growth to develop the energetic reserve, but inherit the reserves from its developmental progression.

^c Reduces during diapause.

^d Not relevant during diapause.

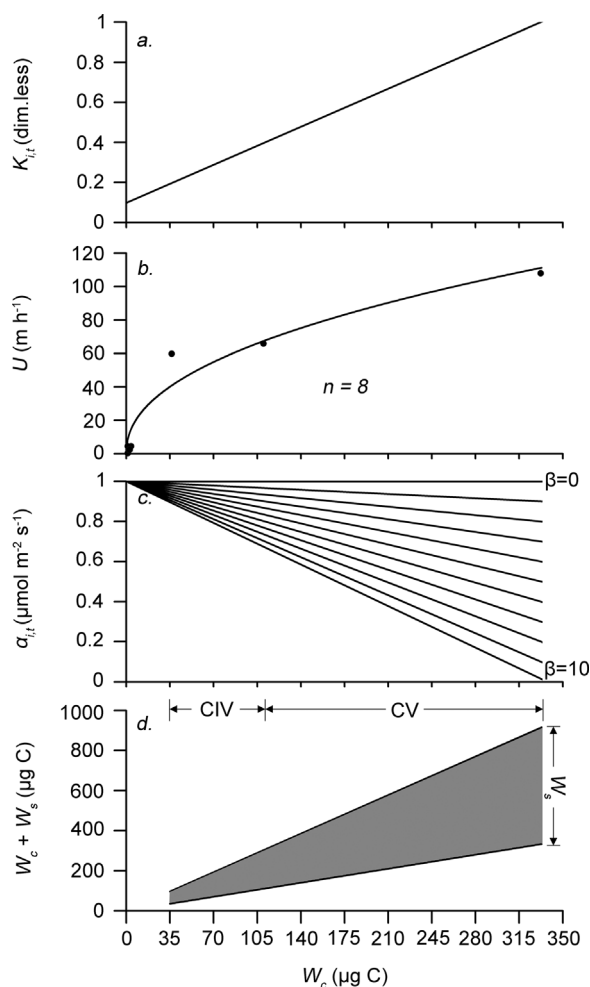


Fig. 4. Relationships of (a) visual predation risk scalar, (b) cruising velocity, (c) light sensitivity parameter and (d) the total body mass of the copepod with its structural mass (W_c). The cruising velocity (U) model was fitted using laboratory and field estimates of *Calanus* spp. from Hardy and Bainbridge (1954), Greene and Landry (1985) and Heywood (1996) (points in panel b). The different linear models for β , that scale the light sensitivity parameter (α) are optimized in the model (Table 2). The lower and upper border of the shaded polygon (panel d) represent the total body mass for growth allocation parameter (γ) = 0 and 1 respectively.

ity, U). We used several stage-specific cruising velocity estimates of *Calanus* spp. available in the literature (Fig. 4b), and related those to body mass as;

$$U_{i,t} = 8.0116 \cdot (W_c)_{i,t}^{0.4531} \quad (11)$$

We considered the size- or stage-specific variability of DVM as a response to size-dependent visual predation risk (Zaret and Kerfoot, 1975; Uye et al., 1990; Hays et al., 1994; Eiane and Ohman, 2004) and modelled it by scaling the light sensitivity parameter (α) with the body mass (W_c). As data on the light sensitivity of younger developmental stages (NI–CIII) of *Calanus* spp. is rare, we could not derive a general relationship between W_c and α . To address this, we defined an evolvable parameter β that describes the size specificity of α , which, at its maximum ($\beta = 10$) downscales α of the adult female to 10% of that of the egg/NI (Fig. 4c). Higher trajectories than $\beta = 10$ were not used, as it was shown in the trial runs that the model always converges on $\beta < 10$ even at highest levels of visual predation risk.

2.6.2.3. Energy storage. CIV and CV stages can allocate a specific fraction from surplus growth (evolvable growth allocation parameter: γ , Table 2) to build up an energy reserve (Fig. 3c) that possesses a maximum size of 70% of the total body mass (Fig. 4d, Fiksen and Carlotti, 1998).

2.6.2.4. Seasonal vertical migration. Similar to most high-latitude marine zooplankton, which descend to depths during the unproductive part of the year (reviewed in Conover, 1988; Hagen and Auel, 2001; Falk-Petersen et al., 2009), the copepod may perform SVM. We used the state of the energetic reserve as a proxy of timing of the SVM (cf. Visser and Jónasdóttir, 1999). Here, the copepod descends to a specific depth (evolvable overwintering depth ζ , Table 2) when the stores account for a specific fraction of the total body mass (evolvable seasonal descent parameter: δ , Table 2). Upon reaching the overwintering depth, the copepod remains stagnant at a diapause state (Hirche, 1996) with its metabolic rate reduced by 90% from that under normal conditions (Fig. 3d, Table 5, Pasternak et al., 1994; Varpe et al., 2007). The overwintering period terminates when a specific fraction (evolvable seasonal ascent parameter: ε , Table 2) of the energetic reserve is exhausted. After the overwintering period, surplus gains are not allocated to develop further energetic reserves, but may be used for structural growth and reproduction (Fig. 3e and f, Table 5).

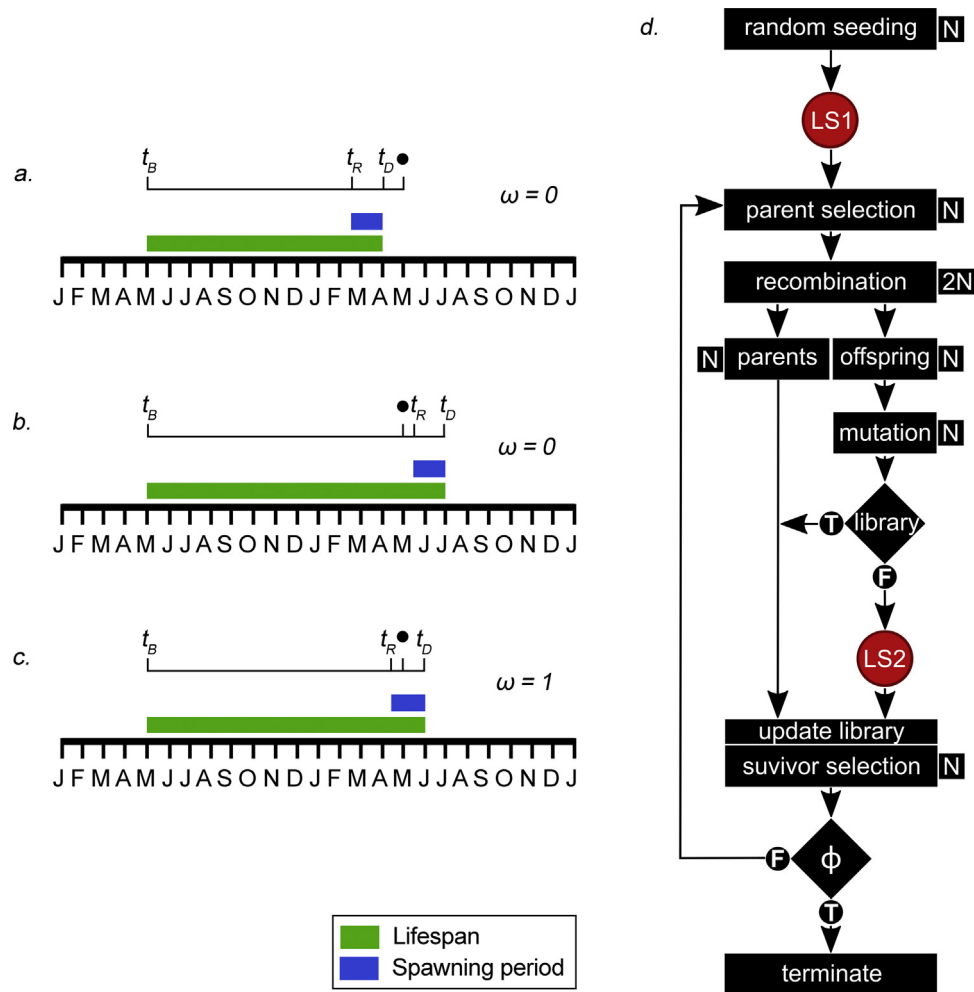


Fig. 5. (a–c) Mechanism of weighing fitness. Fitness of a copepod is multiplied by a binary weight $\omega = 0$ if its egg production season ($t_D - t_R$) does not overlap the time of birth (t_B , May 1 in this example, denoted by a black dot) and vice versa. (d) Simplified workflow of initialization and optimization steps of the model. The initial set of strategies enter the optimization loop after going through the first life cycle simulation (LS1). The GA optimizes seven evolvable (soft) parameters (Table 2) by repeatedly applying selection, recombination and mutation operators until a termination condition (ϕ) is satisfied. T and F are Boolean true and false conditions. No. of strategies (i.e. size of the GA-population, N or 2N) at each operation is indicated to the right.

2.6.2.5. Metabolism. The basal metabolic cost relates with the body mass and ambient temperature, expressed as $k \cdot W^m$ in Eq. (3) (terms as defined above and in Table 4). The metabolic cost of zooplankton vertical movements can account for 0–300% of the basal metabolic demand (Vlymen, 1970; Foulds and Roff, 1976; Morris et al., 1985; Dawidowicz and Loose, 1992). For simplicity, we assumed the cost of vertical movement to be 150% of the basal metabolic cost (mid-point of the above range). This additional cost is subtracted from the growth Eqs. (1) or (3). The energy reserve is used to balance the metabolic demands that cannot be sustained under low ambient food concentrations (Fig. 3c–f).

2.6.2.6. Starvation risk. When energy reserves are depleted, the metabolic demands that cannot be balanced by food intake are met by catabolizing structural body mass (Fig. 3b–f). This elevates the mortality risk due to starvation (starvation risk, M_s), which is defined as a probability that increases as a linear function of catabolized structural mass as;

$$(M_s)_{i,t} = 2 \cdot (W_q)_{i,t} \tag{12}$$

Here, W_q is the catabolized structural mass expressed as a proportion of the maximum structural mass prior to structural catabolization. W_q can reach a maximum of 0.5, during which M_s peaks following Eq. (12), and the copepod dies according the Chos-

sat’s rule (Chossat, 1843), which posits that starving animals may catabolize about half of their body weight before death. Irrespective of the age of this generalized rule, it has been used as a constraint in starvation studies of many vertebrate and invertebrate taxa (e.g. Threlkeld, 1976; Spencer, 1997; Costello, 1998; Loos et al., 2010).

2.6.3. Reproduction

We assumed that somatic growth ceases after the final molt, and all adults become sexually mature at a constant structural body mass (Fig. 3f, Table 5). Energetic input to egg production may be sourced from food intake (income breeding) or allocating a specific amount of matter (C) equivalent to the maximum growth rate (G_T : Eqs. (1) and (2)) from the remaining energetic reserve (capital breeding, cf. Varpe et al., 2009). The fecundity (R) from the time of sexual maturity ($t_R =$ time of final molt) to a given time horizon (t_X) is estimated using the matter allocated to egg production (W_R) and the unit egg mass ($W_E = 0.55 \mu\text{g C}$) as;

$$R_i = \sum_{t_R}^{t_X} \frac{(W_R)_{i,t,z}}{W_E} \tag{13}$$

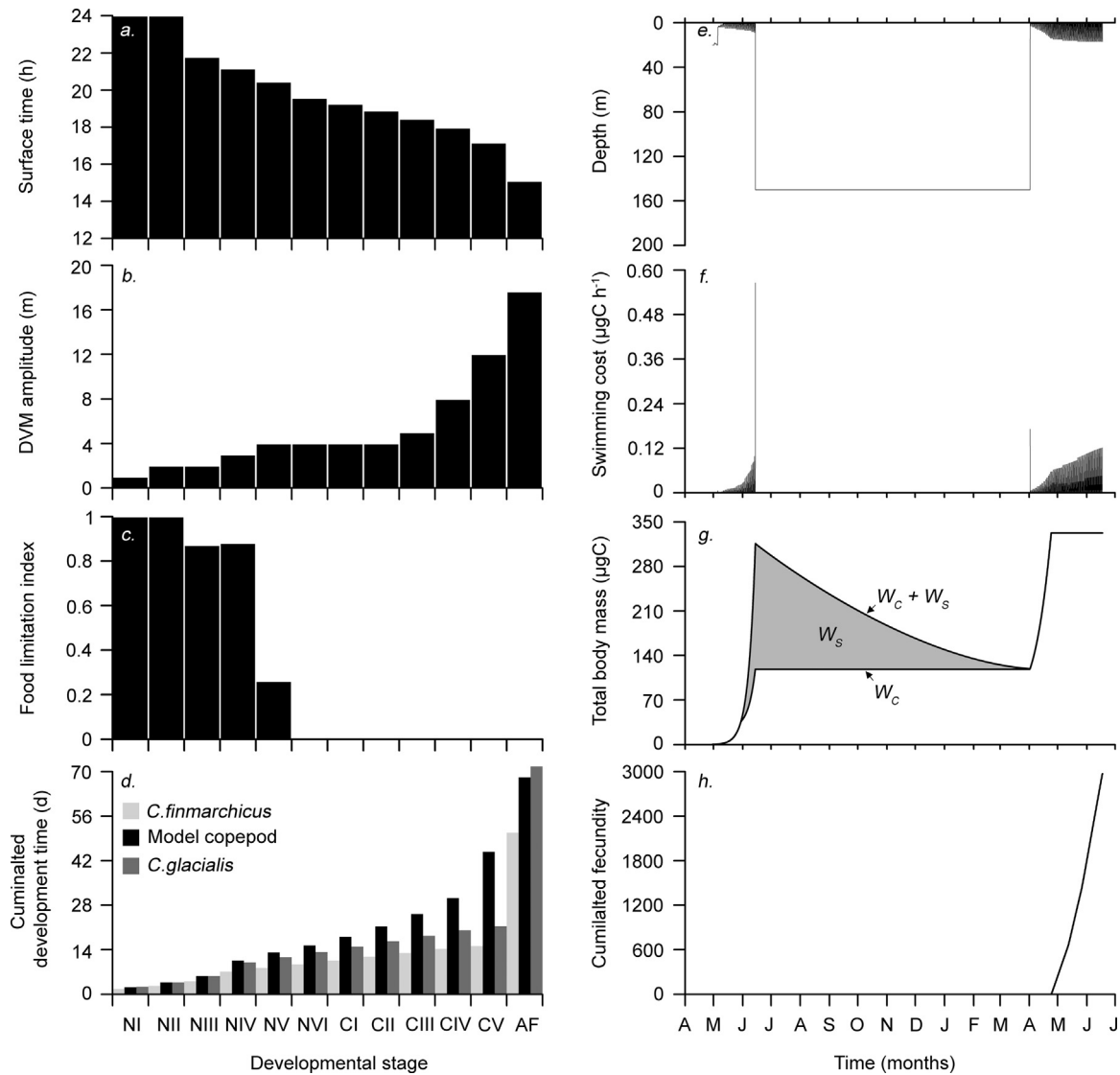


Fig. 6. Predicted optimal vertical strategy and associated growth and reproductive performances of the copepod in the basic run at Environment-L (cf. Fig. 1a–c). The surface time (a) is the stage-specific mean no. of hours per day that the copepod occupies food-rich surface waters, and amplitude (b) is its vertical range. Panel d compares predicted development times (excluding overwintering duration) to those estimated for *C. finmarchicus* and *C. glacialis* following Bělehrádek functions parameterized by Campbell et al. (2001) and Ji et al. (2012). W_c and W_s refer to structural body mass and size of the energetic reserve respectively.

2.6.4. Fitness function and optimization

To evaluate the performance of a vertical strategy, we derived a fitness estimate (Ω) as a function of survivorship and fecundity as;

$$\Omega_i = \left(\sum_{t_B}^{t_X} H_{i,t,z} \cdot R_{i,t,z} \right) \cdot \omega \quad (14)$$

Here, ω is a weight that adjusts fitness (see below) and H is the survivorship, i.e. the probability of survival from birth (t_B) to a given time horizon (t_X) estimated as a function of visual, non-visual and starvation risks (M_v , M_n and M_s) as;

$$H_{i,t,z} = \prod_{t_B}^{t_X} 1 - [(M_v)_{i,t,z} + (M_n)_{i,t,z} + (M_s)_{i,t}] \quad (15)$$

The term Ω technically resembles the net reproductive rate (e.g. Stearns, 1992), and is used in some optimization models (e.g. Kjørboe and Hirst, 2008) but may not bare the same interpretation given the strategy-oriented construct of this model. When the model predicts an optimal vertical strategy and time of birth

for a particular environment, we can assume that those predicted optima should persist from one generation to the next if the environment remains constant. If a copepod's spawning period lasts from time t_R to t_D (time of death) we assumed that it produces a series of offspring with the same vertical strategy, but born at different times of the year (ranging from t_R to t_D). However, only the offspring with a time of birth matching that of the mother can represent the entire evolvable (soft) parameter set of the mother, and guarantee its persistence from one generation to another (Fig. 5a–c). Therefore, we adjusted the fitness using a binary weight (ω) by setting $\omega = 0$ if the copepod's spawning season does not overlap its time of birth (Fig. 5a and b) and vice versa (Fig. 5c).

We used a Real-Coded Genetic Algorithm (RCGA) to derive heuristic estimates of optimal vertical strategy and time of birth that maximizes fitness in a given model environment (Fig. 5d). In the RCGA, six proxies of vertical strategies and the time of birth of the copepod that those are hardwired to (Table 2) are considered as genes on a single chromosome. The RCGA begins by selecting a mating pool of N chromosomes (=parents, i.e. N vertical strategies seeded in different times of the year) from the initial

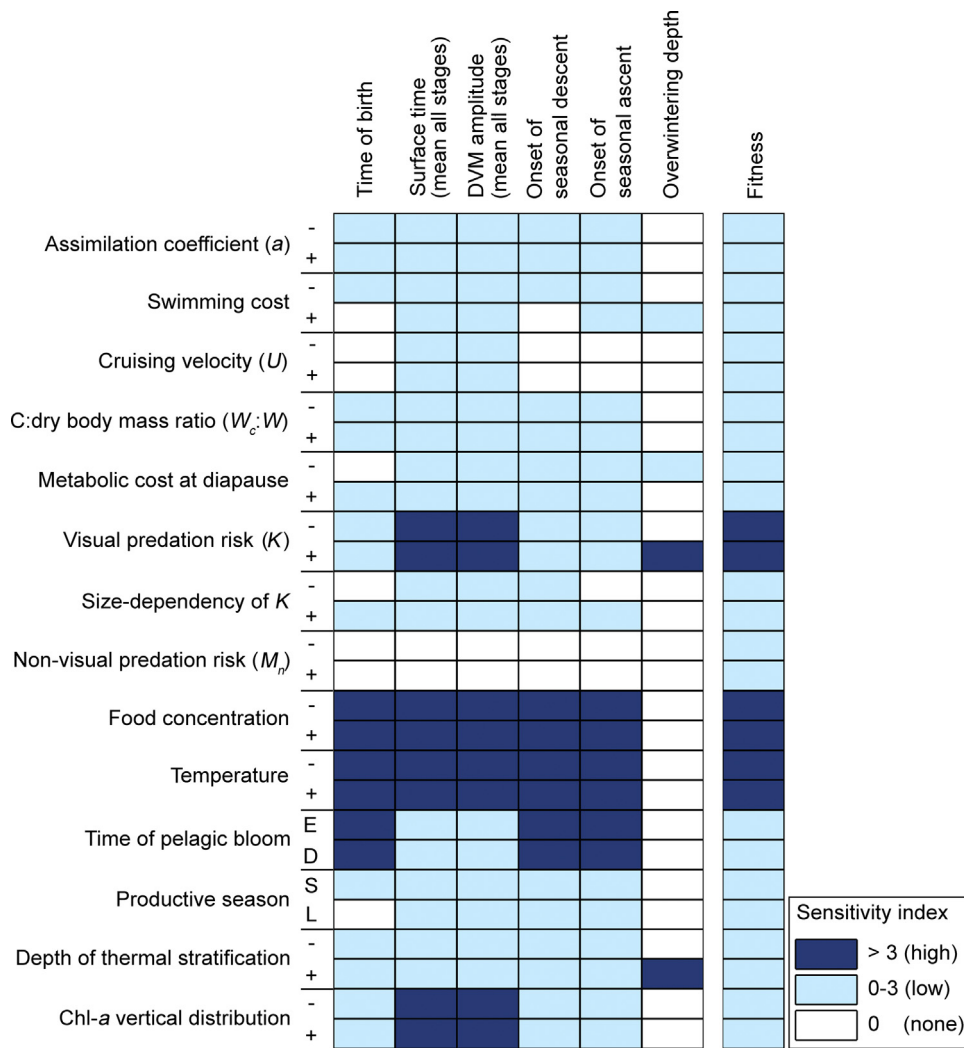


Fig. 7. Graphical summary of the sensitivity analysis. Model parameters and environmental variables tested for sensitivity are presented on the vertical dimension, and the model-predicted optima of time of birth and vertical strategy, and the associated fitness on the horizontal dimension. +/-: 25% increase/decrease in the parameter value, E/D: 15-d earlier/delayed and S/L: 15-d shorter/longer scenarios regarding timing and duration of the productive season (see Appendix A3 in Supplementary material).

seeds using a binary (two-way) deterministic tournament selection (Goldberg and Deb, 1991; Miller and Goldberg, 1995). Genes of two randomly selected parents from the mating pool are recombined through blend crossover following the BLX- α method (Eshelman and Schaffer, 1993), which produces two offspring (recombinants). Genes of the recombinants are mutated at a probability of 0.02 by random replacement (uniform mutation: Eiben and Smith, 2003; Haupt and Haupt, 2004). The population of strategies resulting from these operations comprises of N parents, whose fitness is known and N offspring, whose fitness is not yet known. Parents with unique gene combinations are selected to construct a library (hereafter, the reference library), which is updated at each iteration. Each offspring is compared with those in the reference library to assess their fitness. Fitness of the offspring with similar gene combination to those in the library are assigned *in-situ*, while the rest goes through the life cycle simulation to determine fitness (LS-2 in Fig. 5d). Once the fitness of all $2N$ individuals are known, N survivors are selected following a round-robin (all-play-all) tournament of size 10 (Harik et al., 1997; Eiben and Smith, 2003). This process is repeated for a minimum of 100 iterations, and terminated when the mean fitness of the population shows no improvement for 25 consecutive iterations (ϕ in Fig. 5d, Eiben and Smith, 2003).

2.7. Programming, execution and analysis of the model

We used R version 3.3.1 (R Core Team, 2016) and R Studio integrated development environment (IDE) version 1.0.136 (RStudio Team, 2016) along with the high-performance computing packages Rcpp (Eddelbuettel et al., 2011) and bigmemory (Kane et al., 2013) to construct, simulate and analyze the model.

A basic run was performed in the Environment-L using default values for model parameters (Table 4). In order to test the influence of model parameters and environmental variables on model-predicted vertical strategies and fitness, we performed a sensitivity analysis following (Jørgensen and Bendoricchio, 2001). Here, we calculated a sensitivity score (S_x) as;

$$S_x = \frac{(X_{BR} - X_M) / X_{BR}}{(P_{BR} - P_M) / P_{BR}} \quad (16)$$

where X is the predicted model output of the basic run (X_{BR}) and the modified run (X_M) for a given change ($\pm 25\%$) of input parameter value between the basic run (P_{BR}) and the modified run (P_M). We tested the sensitivity of vertical strategies and fitness for 13 different input parameters (Appendix A3 in Supplementary material). For the convenience of interpretation of these results, we presented the sensitivity scores under three categories: no-sensitivity ($S_x = 0$), low sensitivity ($0 < S_x \leq 3$) and high sensitivity ($S_x > 3$). Finally, we

Table 6
 Predicted variability of emergent behavioral strategies and life history traits (Fig. 2, Table 3) induced by model parameters and environmental variables analyzed for sensitivity (Fig. 7, Appendix A3 in Supplementary material). Only the input parameters with highest influence on fitness (sensitivity index >3) are tabulated.

Parameter	Variation	Time of birth (day)	Surface time ^a (h)	DVM amplitude ^a (m)	Onset of seasonal descent (day)	Onset of seasonal ascent (day)	Overwintering depth (m)	Development time ^b (d)	Food limitation index ^a	Body mass at seasonal descent (μgC)	Onset of egg production (day)	Fecundity	Breeding mode index	Longevity (d)	Absolute fitness
Basic run	0%	120	20.5	5	166	92	150	68.1	0.33	120.4	114	2972	0.000	414	4.10
Visual predation risk (K)	-25%	118	22.7	3	163	91	150	65.2	0.38	127.0	111	3834	0.000	430	21.2
	+25%	123	19.8	7	168	90	170	69.4	0.29	116.0	114	1446	0.000	389	0.90
Food concentration (F)	-25%	105	22.9	3	183	85	150	98.4	0.25	221.1	105	415	0.018	384	1.00
	+25%	122	19.8	6	157	99	150	59.5	0.63	120.8	123	3957	0.000	424	8.70
Temperature	-25%	103	23.0	2	181	82	150	99.1	0.21	229.7	103	654	0.022	387	1.20
	+25%	127	19.7	7	160	107	150	47.0	0.68	124.4	121	5285	0.000	412	16.1

W_c = structural mass, W_s = mass of energetic reserve.

^a Mean of all stages.

^b From egg to adult excluding the overwintering period.

studied the changes of heuristically optimized vertical strategies by performing model simulations in the three model environments at different levels of visual predation risk, while maintaining the rest of the parameter values at its default levels.

As GAs produce heuristic estimates of the maximum fitness, there is no guarantee that it would converge on the global maximum given a potentially diverse fitness landscape (Record et al., 2010). Therefore, we replicated each model run 10 times with different starting values assigned to the soft parameters (Table 2) to check if the algorithm converges on the same set of solutions. As the optimized parameter values showed little variability between replicate runs (<5%), we used the mean of the replicates for each parameter for analyses.

3. Results

3.1. The basic run

In the basic run, the life cycle emerging from the model began as an egg spawned at 20 m depth in late-April. The first two nauplii stages did not perform DVM, but DVM and the associated metabolic cost (swimming cost) increased ontogenetically from NIII onwards (Fig. 6a, b, e and f). The somatic growth of all developmental stages beyond NV occurred under food limitation (Fig. 6c), and because of reduced growth rates, their predicted development times were higher than those estimated from Bělehrádek temperature functions (Fig. 6d). As the energetic reserve reached 65% of the total body mass (W_s ≈ 196 μgC, Fig. 6g), the developmental stage CV descended to an overwintering depth of 150 m in mid-June, ca. 2 months before the pelagic primary production had terminated (Fig. 6e). It remained there for ca. 290 days, and ascended into near-surface waters again in early-April of the following year, ca. 10 days after the pelagic primary production had commenced (Fig. 1c), with fully depleted energetic reserves. Although the energetic cost of seasonal migration was quite high, the cost of ascent was ca. ¼ of that of the descent due to the loss of body mass during overwintering (Fig. 6f and g). The copepod developed to an adult female in mid-April, and thenceforth produced eggs (ca. 3000) via income breeding until mid-June and then died (Fig. 6h).

3.2. Sensitivity analysis

The model-predicted fitness was highly sensitive to visual predation risk, food concentration and temperature (Fig. 7). A 25% change in the visual predation risk ($K = 7.5 \times 10^{-3}$ and 1.25×10^{-2}) influenced the DVM, which intensified at the higher-end of K and vice versa (Table 6). Although the overwintering depth deepened by ca. 13% under higher visual predation risk, it did not change under lower visual predation risk. Furthermore, higher visual predation risk lowered the fecundity and longevity, and vice versa (Table 6).

A 25% change in food concentration ($F = 10$ and $6 \text{ mg m}^{-3} \text{ Chl.-a}$) notably influenced the DVM, timing of SVM and time of birth, but not the overwintering depth (Fig. 7). Under low food concentration, the DVM was less pronounced and the seasonal descent was delayed ca. 15 d compared to the basic run, possibly because of foraging later into the feeding season due to lower growth potential sustained under increased food limitation (Table 6). However, the copepod overwintered as a significantly large CV with elevated energetic reserves, and made ca. 7-d earlier spring ascent (late March), followed by spawning that preceded the pelagic bloom by ca. 2 days (cf. Fig. 1c). Here, ca. 2% of the total egg production was sourced from capital breeding (Table 6). Early seasonal ascent, capital breeding and early spawning thus appear as strategies employed to avoid seasonal peak in visual predation risk (cf. Fig. 1a, Eqs. (9) and (10)) when foraging efforts are elevated to cope with lower

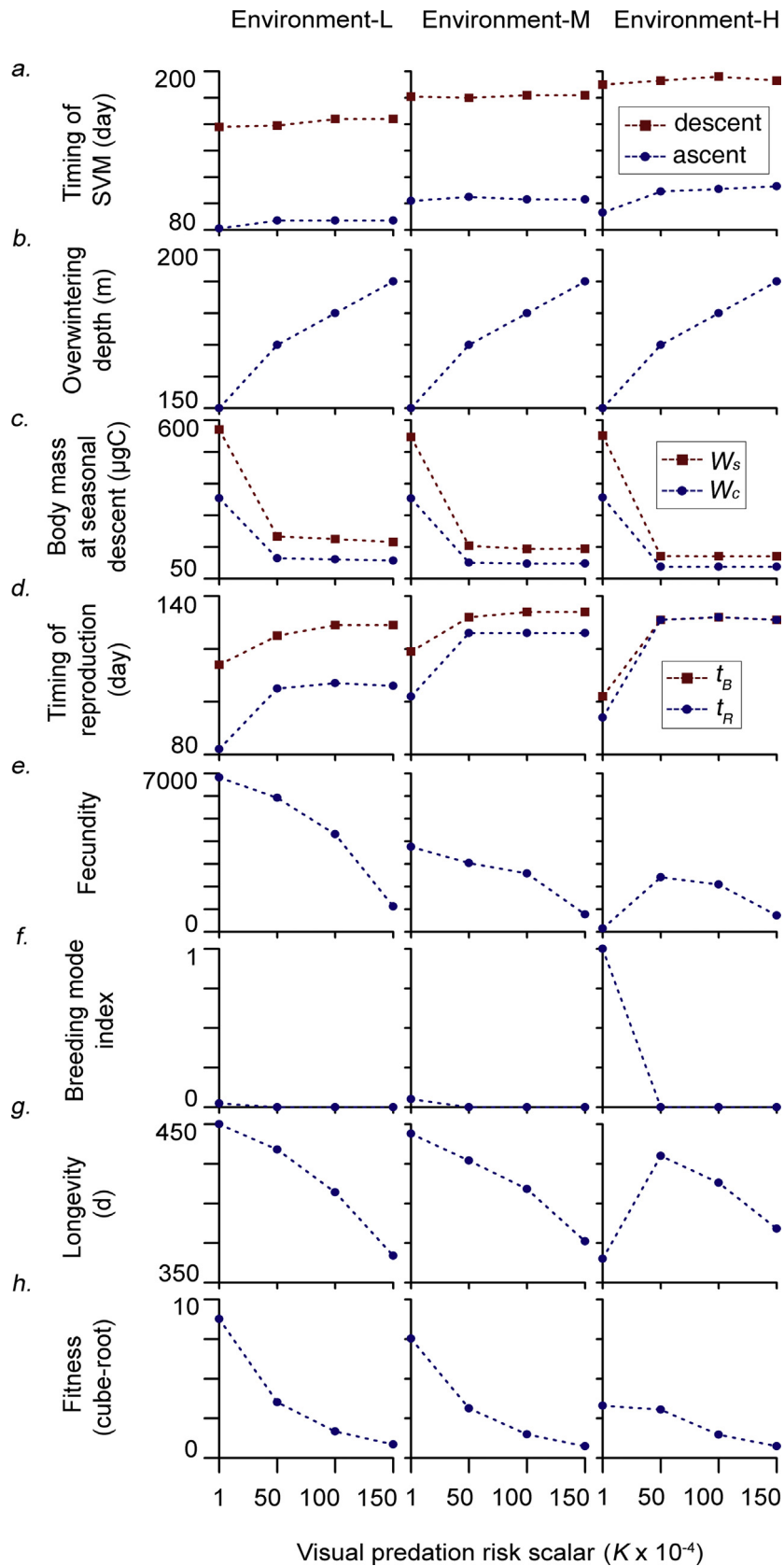


Fig. 8. Predicted optima of time of birth (t_B), vertical strategy and associated life history traits of the copepod in the three model environments under variable levels of visual predation risk. Visual predation risk is scaled by varying the parameter K in a range of $1-150 \times 10^{-4}$. The fitness is cube-root transformed for the convenience of visualization. Time is presented as day of the year, where day 1 = 1 January. W_c and W_s are structural and energy reserve masses, and t_R is time of first reproduction.

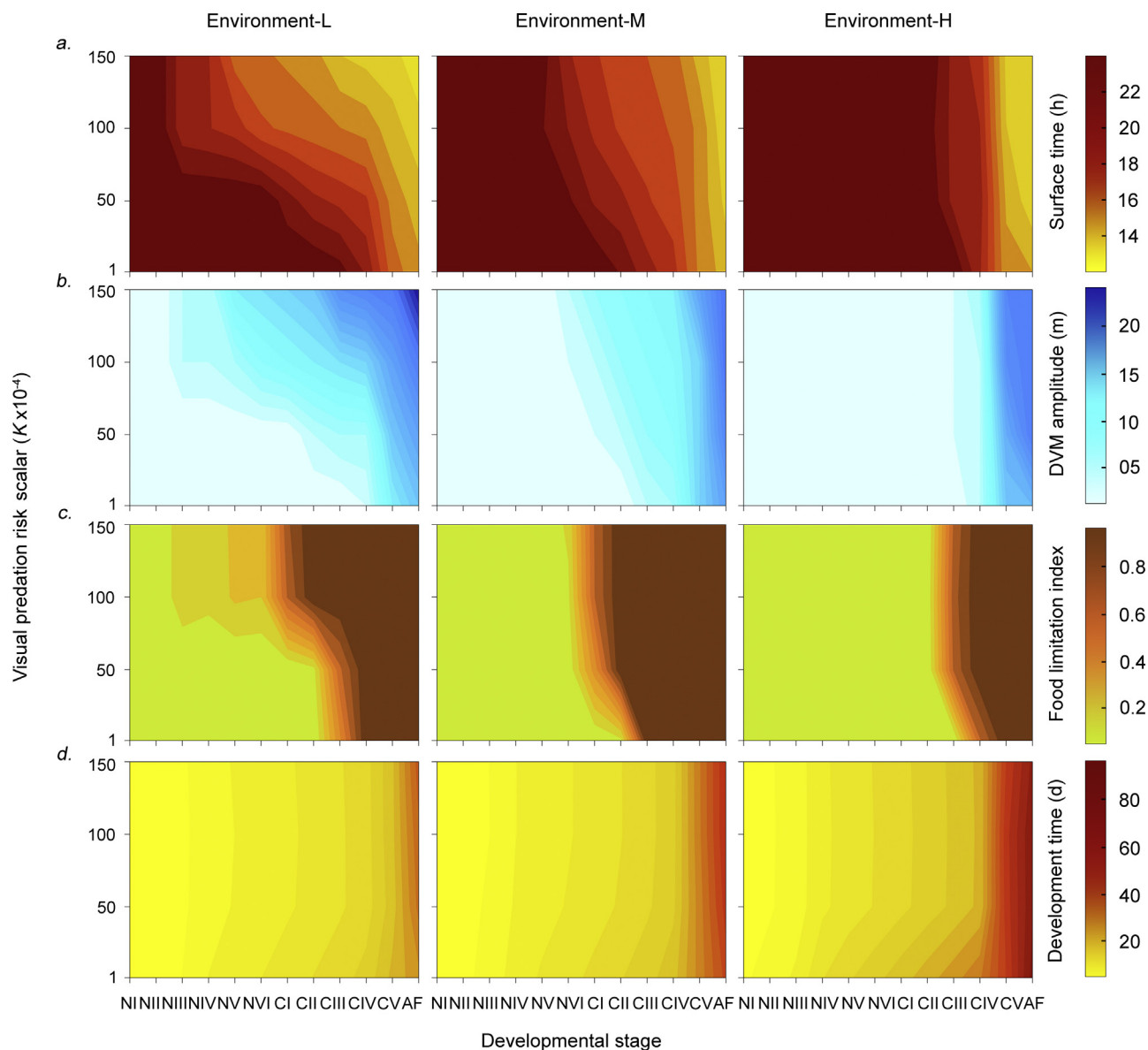


Fig. 9. Predicted stage-specific surface time (a), DVM amplitude (b), food limitation index (c) and the development times (egg to a given stage excluding the overwintering period, d) of the copepod in the three model environments under variable levels of visual predation risk.

growth potential. However, this came with a cost of decreased fecundity (80%) and longevity (7%) (Fig. 7, Table 6). Increased food concentration had the opposing effects on the predicted behavioral strategies and life history traits described above (Fig. 7, Table 6).

The influence of 4.5 °C change in temperature ($T=22.5$ and 13.5 °C) on copepod's behavior and life history followed the same general trends described for food concentration (Fig. 7, Table 6), and highlight the equally important roles played by food availability and temperature in growth and development.

3.3. Latitudinal environmental variability and visual predation risk

Model-predicted optimal time of birth, body mass (W_c and W_s) at seasonal descent, overwintering depth and longevity changed with visual predation risk (K), but showed less variability along the modelled latitudinal environmental gradient (Fig. 8b–d and g). The predicted optimal time of birth changed from late-April to mid-May with increasing visual predation risk, and was con-

stant across the three model environments, with the exception that it occurred the earliest in mid-April at Environment-H under the lowest visual predation risk ($K=10^{-4}$, Fig. 8d). At lower levels of visual predation risk, the copepod overwintered as relatively large CVs with elevated energetic reserves at relatively shallow depths (Fig. 8b and c). However, as the visual predation risk increased, the copepod overwintered as smaller CVs or CIVs with relatively less energetic reserves at greater depths. Although the predicted longevity decreased by ca. 80 d along the modelled visual predation risk gradient (Fig. 8g), longevity at Environment-H under $K=10^{-4}$ was lower (365 d) compared to those predicted under higher visual predation risk levels (384–430 d).

The predicted timing of SVM showed significant variability across the three model environments, but was less affected by visual predation risk (Fig. 8a). Both the descent and ascent occurred earliest at the lowest latitude environment, but happened later in the season at higher-latitude environments, with a shift of about a month. Although this reflects the delayed occurrence of the pelagic bloom along the modelled latitudinal gradient (Fig. 1c, f

and i, Appendix A1 in Supplementary material), the seasonal ascent at Environment-H under lowest visual predation risk ($K=10^{-4}$) occurred ca. 25 days before the pelagic primary production had commenced (Figs. 1i and 8a).

The predicted onset of spawning, fecundity and breeding mode index (Fig. 8d–f), along with the predicted timing and amplitude of DVM, food limitation index and development time (Fig. 9) varied with both the visual predation risk and latitudinal environmental gradient. In all model environments, the spawning started earliest under the lowest visual predation risk ($K=10^{-4}$, Fig. 8d). Here, spawning commenced ca. 5–7 days earlier than the onset of spring primary production in lower latitude environments (Fig. 1c and f), and ca. 2.5–5% of the total egg production were sourced from capital breeding (Fig. 8f). Spawning at Environment-H commenced ca. 25 d prior to the pelagic bloom, but lasted only for ca. 10 days (Figs. 1i, 8d and g). Consequently, the expected fecundity was the lowest (ca. 145 eggs, Fig. 8e) and all eggs were sourced from capital breeding (Fig. 8f). The onset of spawning shifted later into the season at higher latitude model environments at higher levels of visual predation risk, and occurred after the commencement of the pelagic primary production (Figs. 1c, f, i and 8d). Here, all eggs were produced via income breeding (Fig. 8f). At lower levels of visual predation risk ($10^{-4} \leq K \leq 5 \times 10^{-3}$), the predicted DVM pattern was similar across the three model environments, where developmental stages until early copepodites did not perform DVM (Fig. 9a and b). Although the model predicted the younger developmental stages (NIII onwards) to perform DVM under elevated visual predation risk at Environment-L, this effect gradually waned in higher-latitude model environments. The food limitation index strongly followed the DVM pattern, where developmental stages that performed DVM suffered from increased food limitation (Fig. 9c). Food limitation significantly reduced the growth rates (cf. Eqs. (1)–(3), see also Appendix A4 in Supplementary material), and consequently, the development times increased along the modelled environmental gradient (Fig. 9d). Further, in each model environment, lowest development times were predicted under the lowest level of visual predation risk.

4. Discussion

4.1. Influence of environmental variables on vertical strategies

4.1.1. Diel vertical migration

In this model, visual predation risk had the highest influence on the DVM, which diminished under low visual predation risk (Figs. 7 and 9, Table 6) and completely ceased when visual predation was removed from the model ($K=0$, data not presented). Conversely, under high visual predation risk, also younger developmental stages reduced the time spent in food-rich surface waters by performing low-amplitude (shallow) DVM (Fig. 9a and b). Food concentration and temperature also influenced the DVM (Fig. 7, Table 6). Lower food concentrations or temperatures produced low-amplitude DVM, possibly due the low growth potential attained in cold, food-limited conditions (Fig. 9, Eqs. (1)–(8), Appendix A4 in Supplementary material). Under these conditions, it appears that modelled copepods do not possess sufficient growth potential to trade off for survival and perform high-amplitude (deep) DVM, a conclusion also drawn in empirical work (e.g. Huntley and Brooks, 1982; Loose and Dawidowicz, 1994). Reduced or absence of DVM under low food concentrations and temperatures are reported from several other modeling studies on copepods and euphausiids (e.g. Andersen and Nival, 1991; Fiksen and Giske, 1995; Fiksen and Carlotti, 1998; Tarling et al., 2000) and from empirical work on marine copepods and freshwater cladocerans (e.g. Hardy and Gunther, 1935; Huntley and Brooks, 1982; Johnsen and Jakobsen,

1987). However, the largely exogenous-regulated DVM behavior emerging from this model does not render a complete view on the subject matter, as DVM can also be sensitive to internal (physiological) states of zooplankton (e.g. hunger and satiation, Hays et al., 2001; Pearre, 2003).

The effect of temperature on growth potential further explains the diminished influence of visual predation risk on the timing and amplitude of DVM predicted at higher latitude model environments (Fig. 9). Albeit similar food concentrations, the modelled temperatures decreased from lower- to higher-latitude model environments, reflecting a decreasing gradient of growth potential (Fig. 1, Appendix A4 in Supplementary material). Consequently, the model-predicted optimal DVM strategy for higher-latitude environments was to elevate the growth potential by spending more time foraging in near-surface waters (Fig. 9a and b). This effect was most pronounced among younger developmental stages (NIII–CI), whose DVM reduced from environment-L to -M, and completely ceased at environment-H. DVM of younger developmental stages (NIII onwards) are most commonly reported from lower latitudes for *Calanus* spp. (e.g. Huntley and Brooks, 1982; Uye et al., 1990; Huang et al., 1993; Osgood and Frost, 1994; Zakardjian et al., 1999) and *Metridia* spp. (e.g. Hays, 1995). While some field studies failed to detect notable DVM in high-latitudes (e.g. Blachowiak-Samolyk et al., 2006; Basedow et al., 2010), others reported ontogenetic increase of DVM (CI onwards, e.g. Dale and Kaartvedt, 2000; Daase et al., 2008) in *Calanus* spp. However, empirical data with high spatial, temporal and biological (i.e. developmental stage) resolution is needed to test the ontogenetic and latitudinal patterns of DVM predicted by our model.

The predicted DVM amplitudes spanned across the productive part of the water column (i.e. upper 30 m, cf. Figs. 1, 6 and 9), and showed a positive relationship with the vertical extent of food availability (Fig. 7). Moreover, strategies that involve higher-amplitude DVM lead to increased food-limitation, where younger developmental stages with no energetic reserves tend to suffer from starvation risk due to low temperatures and food concentrations that prevail in deeper parts of the model environments (Fig. 1). Therefore, low-amplitude DVM appears to be a strategy that efficiently trades off growth potential for survival, by balancing both the visual predation and starvation risks (Kerfoot, 1970; Fiksen and Giske, 1995; De Robertis, 2002). Although higher-amplitude DVM can be predicted either by not modeling starvation risk or imposing starvation tolerance (e.g. Andersen and Nival, 1991; Carlotti and Wolf, 1998; Zakardjian et al., 1999; Tarling et al., 2000), we did not follow these approaches because our model sufficiently represents the relative importance of DVM across the modelled environmental gradients.

4.1.2. Seasonal vertical migration

Food availability and temperature had the most notable influence on the model-predicted timing of SVM (Fig. 7, Table 6). The predicted shift in timing of seasonal descent and ascent coincided with those of the pelagic algal bloom and thermal stratification along the modelled latitudinal gradient (Figs. 1 and 8a). This agrees with the argument that food availability is the ultimate factor influencing the timing of seasonal vertical migration of *Calanus* spp. (Herman, 1983; Head and Harris, 1985; Hind et al., 2000). However, field estimates of timing of SVM of *Calanus* spp. from low to high latitudes do not point to a simple south–north gradient as predicted in our model (Table 7, see also Melle et al., 2014). This discrepancy of model predictions and field estimates underlies the differences between location-specific variability in hydrography, algal bloom dynamics and species composition (e.g. Hirche, 1991; Daase et al., 2013), diversity of generation lengths and breeding strategies (e.g. Conover, 1988; Falk-Petersen et al., 2009) and climate-driven and other stochastic oscillations of environmental

Table 7
Timing of seasonal ascent and descent of *Calanus* spp. estimated by several high-latitude field investigations. These estimates are based on observation of zooplankton populations oftentimes containing various combinations of *C. helgolandicus*, *C. finmarchicus*, *C. glacialis* and *C. hyperboreus*. Geographical location is approximate. Data for North-Atlantic are available in Melle et al. (2014).

Study	Lat.	Lon.	Onset of descent	Onset of ascent
Hirche (1984)	58° N	11° E	mid-October	mid-May
Bagøien et al. (2000)	59° N	10° E	July–August	March
Heath (1999)	61° N	4° W	–	May
Gislason and Aasthorsson (2000)	62° N	20° W	–	April
Astthorsson (2000)	64° N	28° W	–	April
Østvedt (1955)	66° N	2° E	July	April
Kosobokova (1999)	66° N	35° E	mid-July	mid-May
Astthorsson and Gislason (2003)	68° N	13° E	–	May
Madsen et al. (2001)	69° N	54° W	–	mid-April
Madsen et al. (2008)	69° N	54° W	late-September	early-April
Hirche (1997)	71° N	4° E	July–August	April
	74° N	1° E	July–August	April
Unstad and Tande (1991)	75° N	30° E	–	May
Arashkevich et al. (2002)	76° N	33° E	July	May
Hirche and Kosobokova (2011)	77° N	25° E	–	March
Hirche and Kosobokova (2003)	78° N	82.5° E	–	mid-May
Bandara et al. (2016)	78° N	16° E	July–August	mid-February
Dawson (1978)	84° N	112° W	August	June

conditions (e.g. Reid et al., 1998; Eiane and Parisi, 2001; Ji, 2011). However, direct comparisons between vertical migratory patterns predicted by a strategy-oriented model and field estimates should be done with caution, as an environment-specific optimal vertical strategy predicted by the model contrasts the diversity of vertical behavior exhibited by individuals of a zooplankton population.

Visual predation risk and depth of thermal stratification (summer–autumn) were the only environmental variables that influenced the model-predicted overwintering depth (Figs. 7 and 8b, Table 6). The overwintering depth deepened at higher levels of visual predation risk and deeper thermal stratification depths, and agrees with Hirche (1991), Kaartvedt (1996), Dale et al. (1999), Bagøien et al. (2000) and Astthorsson and Gislason (2003) that *Calanus* spp. prefer colder water masses with low predator abundance for overwintering. However, the overwintering depths predicted by or model underestimate those of field observations, which can extend well below 1000 m (e.g. Østvedt, 1955). Apart from the shallow bottom depths modelled, this discrepancy largely reflects how overwintering habitat selection of *Calanus* spp. is influenced by the buoyancy-effect of stored lipid reserves (Visser and Jónasdóttir, 1999), convective mixing of surface waters (Irigoin, 2004) and vertical distribution of water masses and predator populations, such as mesopelagic fish, predatory ctenophores and krill (Hirche, 1991; Kaartvedt 1996; Bagøien et al., 2000; Bandara et al., 2016).

4.2. Influence of vertical strategies on fitness and phenology

4.2.1. Diel vertical migration

In the model, high-amplitude DVM caused increased food limitation that led to slow growth and development and reduced fecundity (Figs. 7 and 9, Table 6). This ultimately resulted in lower fitness relative to that predicted for low-amplitude DVM. It is therefore apparent that decisions to fully or partly disregard DVM in models focusing SVM and other seasonal strategies should be made with caution, as our findings indicate that DVM can have a notable negative influence on growth and development of younger developmental stages, especially at lower latitudes.

At higher levels of visual predation risk, the model predicted up to a one month delay in the onset of spawning (Fig. 8d), and highlights the influence of predation risk on the reproductive phenology (Magnhagen, 1991; Stibor, 1992; Varpe et al., 2007). These late-spawned copepods appeared to possess higher fitness compared

to those spawned earlier (Fig. 8h). This seems counterintuitive as zooplankton are more vulnerable to visual predation risk later in the season due to the higher irradiance levels that persist in late-spring and summer. In this model, the early feeding season (i.e. until the time of peak pelagic bloom) is characterized by higher food concentrations and lower temperatures (Fig. 1). Although the food concentration decreases by ca. 20% by late spring or early-summer, the ambient temperature increases by ca. 2–4 times. Further, even at the onset of the productive season, the visual predation risk had reached ca. 70% of its maximum in all model environments (Fig. 1a, d and g). Therefore, it is likely that copepods born relatively later in the season use the higher temperatures to attain a higher growth (Eqs. (1)–(8), Appendix A4 in Supplementary material), which is then efficiently traded off for survival through DVM to counter the risk of increasing visual predation risk (Fig. 9). Conversely, due to lower temperatures, copepods born earlier in the season must elevate the time spent foraging in near-surface waters to attain higher growth rates, and become more vulnerable to visual predation risk. However, it should be noted that our model does not consider the ability of *Calanus* spp. to use the darker and seasonally ice-covered period of the year to attain growth with minimal influence from visual predators by feeding on alternative food sources, such as ice algae and microzooplankton (Conover and Siferd, 1993; Søreide et al., 2010).

Compared to the phenology of reproduction, the elevated visual predation risk had little influence on the timing of seasonal migration (Fig. 8a). The SVM strategy predicted by the model was to descend to overwintering depth approximately at the same period of the year (mid-late summer), but with ca. $\frac{1}{3}$ lesser the body mass (both structural and energetic reserve mass) compared to that under the lowest visual predation risk (Fig. 8a and c). As higher visual predation risk tends to intensify DVM in this model (Fig. 9a and b), it appears that trading off growth potential for survival makes an earlier seasonal descent unfavorable, as food-limitation and slower growth rates (Fig. 9c and d) demands more time to acquire sufficient energy reserves to overwinter, despite the smaller body mass of the overwintering stage. Further, as the model-predicted body mass of the overwintering stage reaches a lower threshold at $K > 10^{-4}$ (Fig. 8c), it is likely that modelled copepods overwintered with minimum reserves to last the overwintering duration, and therefore makes an earlier descent (with lesser energy reserves) nearly impossible. In the contrary, occupying near-surface waters later into the season and descend to

overwintering depths with elevated energetic reserves is also unfavorable as the visual predation risk in this model is not nullified even at the deepest parts of the water column (Eq. (10)). The lesser influence of visual predation risk on the timing of SVM does not align with Kaartvedt (2000) and Varpe and Fiksen (2010) who view predation by planktivorous fish as a key driver of generation lengths and timing of seasonal descent in *C. finmarchicus* in the Norwegian Sea. However, the consequences of DVM on the timing of seasonal descent presented here may diminish if there is an energetic benefit of DVM (e.g. McLaren, 1963; Enright, 1977), if copepods are capable of utilizing alternative food sources (e.g. Runge and Ingram, 1991; Hirche and Kwasniewski, 1997) or if there is a strong size selection against larger developmental stages by visual predators than the linear relationship applied in our model (cf. Fig. 4a with Brooks and Dodson, 1965; Batty et al., 1990; Langbehn and Varpe, 2017).

4.2.2. Seasonal vertical migration

SVM was essential for the wintertime survival of the modelled copepods, given its food source is only available during the primary production season (spring–autumn, Fig. 1). Different combinations of proxies (Table 2) yielded non-seasonally migrating strategies, in which the copepods developed to adults and reproduced within the same productive season. Although this strategy had the potential to produce more than one generation per year (especially in the relatively lower-latitude environment-L, Fig. 1a–c), we did not pursue this further, as our focus was on an annual life cycle (see the fitness weighing process, Eq. (14), Fig. 5a–c).

The body mass and the size of energetic reserve at seasonal descent together with the timing of seasonal ascent had a profound influence on the predicted timing of reproduction, breeding strategy and fecundity (Fig. 8). Overwintering as large CVs with elevated energetic reserves at lower visual predation risk enabled the copepod to allocate the post-overwintering surplus energetic reserves to capital breeding in the following year (Fig. 8c and f, and see also Sainmont et al., 2014a; Ejsmond et al., 2015; Halvorsen, 2015). As capital breeding emerged in environments with lower temperatures and food concentrations (Fig. 7, Table 6), it appears as a strategy that allows the new generation to feed from the very start of the feeding season, while avoiding the seasonal peak in visual predation risk later in the year (Fig. 1, see also Varpe et al., 2009). The proportional increase of capital breeding eggs from relatively lower-latitude environment-L to higher-latitude environment-H reflects the decreasing temperature gradient that occur at overwintering depths of these environments (Figs. 1, 8b and f). Overwintering in colder water masses reduces the metabolic costs and conserves the energetic reserve, which ultimately boosts the fecundity through capital breeding (Hirche, 1991; Hirche, 1996; Astthorsson and Gislason, 2003). The pure capital breeding strategy predicted at environment-H under the lowest visual predation risk more resembles the spawning strategy of *C. hyperboreus* than *C. finmarchicus* and *C. glacialis* (Conover, 1988; Falk-Petersen et al., 2009). The sensitivity of the model-predicted breeding strategy to visual predation risk indicates an extensive pre-breeding cost of capital breeding imposed by the size-dependent visual predation risk and acquisition and carriage of energy reserves (Jönsson, 1997; Varpe et al., 2009). Moreover, excess energy storage (i.e. more than to overwinter) and capital breeding do not emerge as dominant strategies in this model as the environmental parameters are modelled in a perfectly predictable manner, without any year-to-year variability. However, capital breeding and energy storage may possess a much larger adaptive significance in nature, where spatio-temporal environmental heterogeneity and unpredictability are more pronounced compared to our model (e.g. Jönsson, 1997; Fischer et al., 2009).

4.3. Concluding remarks

Findings of this study highlight the influence of environmental variables on vertical strategies, and suggest that in seasonal environments, DVM and SVM should be studied in concert, as these behavioral strategies can have profound and largely different effects on fitness and phenology of herbivorous zooplankton. Therefore, given the significance of biological information ensued, sacrificing computer time to adopt higher spatio-temporal resolution in behavioral and life-history models seems to be an appealing practice. However, strong recommendations should only be made after testing our model predictions further, especially, through improvements to cope with environmental stochasticity (e.g. Eiane and Parisi, 2001; Ji, 2011), and to incorporate the plasticity of feeding strategies, generation times and body sizes of *Calanus* spp. (e.g. Broekhuizen et al., 1995; Fuchs and Franks, 2010; Ji et al., 2012; Banas et al., 2016).

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Appendix A. Supplementary data

Supplementary data associated with this article can be found, in the online version, at <https://doi.org/10.1016/j.ecolmodel.2017.12.010>.

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this appendix accompanies additional material for

A high-resolution modeling study on diel and seasonal vertical migrations of high-latitude copepods

APPENDIX

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Appendix A1: Additional data on environmental parameters

A1.1 The Irradiance sub-model

The solar irradiance incident on the sea surface (I_0 , Eq. 1 in main text) was determined using the global horizontal irradiance (GHI) model of Robledo and Soler (2000) assuming clear sky conditions and discounting for Rayleigh scattering and various other measurable atmospheric parameters such as air pressure, temperature, perceptible water, ozone and aerosol concentrations.

$$I_0 = 1159.24 \cdot (\cos\theta_z)^{1.179} \cdot \exp(-0.0019 \cdot (90^\circ - \theta_z)) \quad \text{A1.1}$$

Here, θ_z is the solar zenith angle, calculated using,

$$\cos\theta_z = (\cos\phi \cdot \cos\theta_d \cdot \cos\theta_h) + (\sin\phi \cdot \sin\theta_d) \quad \text{A1.2}$$

where, ϕ is the latitude, θ_d is the declination angle (Eq. 3) and θ_h is the hour angle, which is an angular representation of local solar time (T_s , Eq. A1.4, A1.5) in degrees.

$$\theta_d = 23.45 \cdot \sin\left(\frac{360}{365}[D - 81]\right) \quad \text{A1.3}$$

$$\theta_h = 15 \cdot (T_s - 12) \quad \text{A1.4}$$

Here, D is the day of the year (as January 1 = day 1). The local solar time T_s was calculated using the local time (T_L), the difference between the local meridian (= longitude: λ) and local standard time meridian (λ_0 : Eq. 6), and the equation of time (E: Eq. A1.7), which is an empirical equation that accounts for the eccentricity of earth's orbit and the tilt.

$$T_s = T_L + \frac{4 \cdot (\lambda - \lambda_0) + E}{60} \quad \text{A1.5}$$

$$\lambda_0 = 15(T_L - T_{GM}) \quad \text{A1.6}$$

$$E = 9.87 \cdot \sin\left(\frac{360}{365}[D - 81]\right) - 7.53 \cdot \cos\left(\frac{360}{365}[D - 81]\right) - 1.5$$

$$\cdot \sin\left(\frac{360}{365}[D - 81]\right)$$
A1.7

Modelled irradiance roughly agrees with the field estimates (<https://data.met.no/>) and presented in Fig. 1 in the main text.

A1.2 Details of the three model environments

Table A1.1 Detailed description of the three model environments (Fig. 1 in main text)

Parameter	Attribute	Environment-L	Environment-M	Environment-H
Irradiance ($\mu\text{mol m}^{-2} \text{s}^{-1}$)	Min.	0	0	0
	Max.	1500	1300	1100
	Time of Max.	day 172 (June 21)	day 172 (June 21)	day 172 (June 21)
Temperature ($^{\circ}\text{C}$)	Min.	4	3	2
	Max. ($^{\circ}\text{C}$)	18	15	12
	Time of Max.	day 181 (July 1)	day 188 (July 7)	day 195 (July 14)
Food availability (mg m^{-3} Chl.a)	Min.	0	0	0
	Max.	8	8	8
	Time of Max.	day 105 (April 15)	day 121 (May 1)	day 135 (May 15)
	Productive season (duration)	167 d	153 d	137 d

The modeled irradiance is based on the GHI model of Robledo and Soler (2000) and verified by the field estimates of Norwegian Metrological Institute (<https://data.met.no/>). Modelled temperature ranges were adopted from Unstad and Tande (1991); Bagøien et al. (2000); Ingvaldsen and Loeng (2009); Basedow et al. (2010); and Melle et al. (2014), and field estimates made during the LoVE MarinEco (<http://love.arctosresearch.net>) and DWARF (<http://www.iopan.gda.pl/projects/Dwarf/>) projects, obtained via personal communication through Boris Espinasse and Slawek Kwasniwski respectively. Timing and duration of pelagic

bloom were approximations which were adopted from Falk-Petersen et al. (2009); Daase et al. (2013). All modelled environmental variables represent coastal locations along the southern and southeastern Norwegian shoreline (Northeast Atlantic).

Appendix A2: Predicting diel vertical trajectories using a biased random walk

A2.1 Background

Classic random walk models are simple in the sense that the direction of the movement is completely random, and the movement at current time is not dependent on that of the previous time(s). A change in probability of moving in a certain direction can make a global directional bias, and paths that contain a consistent bias towards a preferred direction is called a biased random walk (BRW, Codling 2003). Here, we developed a BRW algorithm to predict copepod's movement in the vertical dimension (1D), in which the bias was introduced by the gradients of environmental variables (e.g. Alt 1980).

A2.2 The model

The copepod's photoreactive behavior is defined by an individual-specific lower irradiance threshold (α , see Table 2 in the main text), beyond which induces a negative phototactic response (Båtnes et al. 2015; Cohen et al. 2015). At any given time, the copepod searches and occupies a position in the water column with an irradiance level below α that maximizes its growth potential. The objective of the BRW algorithm is to predict the copepod's search pattern (direction and moving distance) in each time-step.

The distance that the copepod can cover between each time-step of the model (1 h) is limited by its cruising velocity, which was modelled as a power function of copepod's structural body mass (see Fig. 4a in main text). If the copepod cruises at a vertical velocity of $U_{i,t}$ m h⁻¹, its movement at each 1 h time-step can be predicted using $U_{i,t}$ (≥ 1) no. of vertical moves.

Given $V_{i,t}$ is the no. of moves that copepod has at its disposal between time-step t and $t + 1$ (for example after the first move $V_{i,t} = U_{i,t} - 1$), z the depth, G the growth potential (see Eq.

1–8 in the main text), P_0 the probability of initial movement at time-step t , P_A the probability of advancing along the direction of the previous move, and $P_R (= 1 - P_A)$ the probability of retreating along the opposite direction to the previous movement, the probabilities of vertical movement are given as in Fig. A2.1.

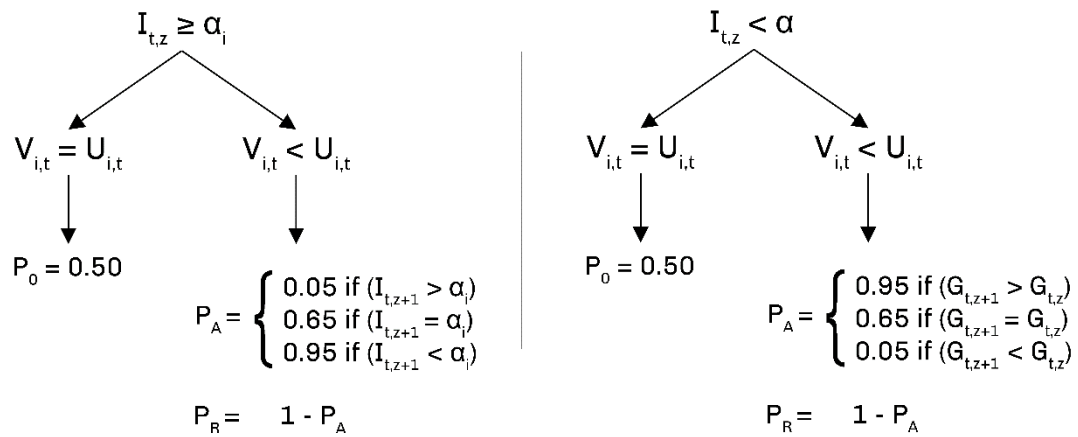


Fig. A2.1 Probabilities of the initial and subsequent vertical movements of the copepod, determined by the Irradiance and the growth potential (which is dependent on food availability and temperature). Decisions are made by drawing uniform random numbers (0–1, with the intervals of 0.01) and checking those against P_0 , P_A and/or P_R .

Tuning the decision-making probability cutoffs (0.05, 0.65 and 0.95) changes the influence of stochasticity in the copepod’s vertical movement. Using a directionally biased cutoff probability of 0.65 when environmental gradients do not exist is used to avoid the copepod getting stuck in unproductive waters or local food optima.

A2.3 Model predictions and drawbacks

The model predicts the diel vertical trajectories fairly well (Fig. A2.2). Given the growth sub-model (eq. 1–8 in the main text) and lack of starvation tolerance in our model, the predicted optimal diel migrations are restricted to the productive part of the water column (see section 4 in the main text). Constraints of daily feeding or some measure of starvation tolerance measure is needed to accurately predict the DVM amplitude and surface times, especially on older developmental stages.

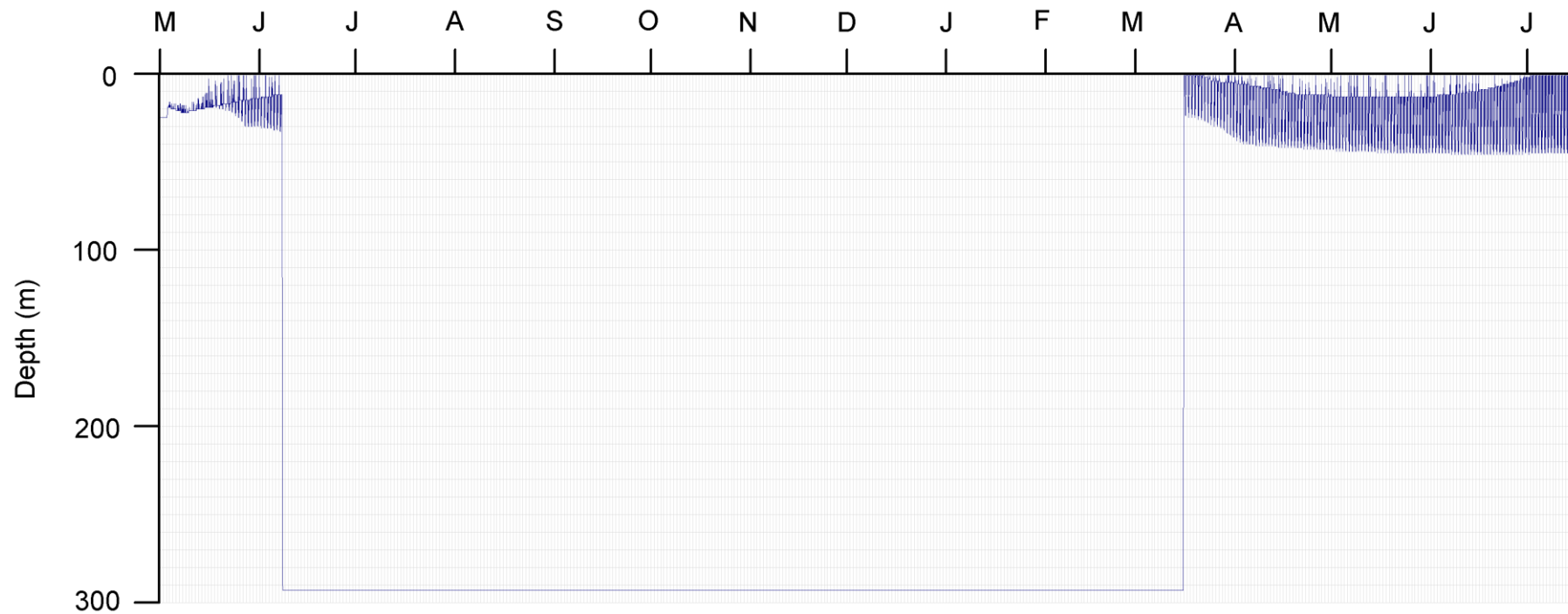


Fig. A2.2 Simulated diel and seasonal vertical trajectory of a copepod with an irradiance sensitivity parameter (α) of 1000 and size-selectivity of α (β) of 2 (see Table 2 and Fig. 4c in main text). Simulation was performed in the Environment-L (Fig. 2 in main text, and Appendix A1).

Appendix A3: Details of the sensitivity analysis

Table A3.1 List of input parameters and environmental variables tested for sensitivity, their relative changes and associated values. Basic run values of these parameters are available in the main text

Input parameter	%Change	Value	Notes
Assimilation coefficient (a)	-25%	0.575	
	+25%	0.875	
Swimming cost	-25%	1.125	Multiplier of basal metabolic cost
	+25%	1.875	
Cruising velocity	-25%	Eq.11	Increase or decrease in the mass-specific cruising velocity (U, Eq. 11)
	+25%	Eq.11	
Body Carbon composition (%)	-25%	30	
	+25%	50	
Visual predation risk (K)	-25%	0.0075	
	+25%	0.0125	
Size dependency of K	-25%	7.5	Changes the slope of the linear model in Fig. 4a in the main text so that the adult female is ca. 7.5 or 12.5 times more vulnerable to visual predators
	+25%	12.5	
Non-visual predation risk (M_n , %)	-25%	0.75	As a percentage of the max. visual predation risk
	+25%	1.25	
Food concentration (F, mg Chl.-a m^{-3})	-25%	10	
	+25%	6	
Temperature (T, °C)	-25%	22.5	Only the maximum temperature. Minimum, i.e. near-bottom temperatures not affected
	+25%	13.5	
Timing of pelagic bloom (d)	–	+15d	In order to maintain the relationship meaningful, the seasonal temperature distribution (Fig. 1 in the main text) was shifted with the algal bloom peak
	–	+15d	
Length of productive season (d)	–	+15d	In order to maintain the relationship meaningful, the seasonal temperature distribution (Fig. 1 in the main text) was expanded/contracted with the productive season expansion/contraction
	–	+15d	
Depth of thermal stratification (m)	–25%	75 m	
	+25%	125 m	
Chl-a vertical distribution (m)	-25%	-0.075	The exponent that scales the vertical scaling of Chl-a distribution (exponential decay with depth) was tuned (default value = -0.06)
	+25%	-0.045	

Appendix A4: Growth rate of the copepod; temperature and food relations

The food-limited growth rate (Eq. 3 in main text) is always lesser than the maximum food-independent growth rate (Eq. 1, 2 in main text). In the model, DVM caused copepods to move away from the regions with maximum food abundance on a daily basis, and resulted in food-limitation, which ultimately caused retardation of growth rates.

The food-limited growth rate increases with temperature and food availability. At any given food concentration, the food-limited growth rate (Eq. 3) decreases and the saturation food concentration (Eq. 8) increased with decreasing temperature (Fig. A4.1–3).

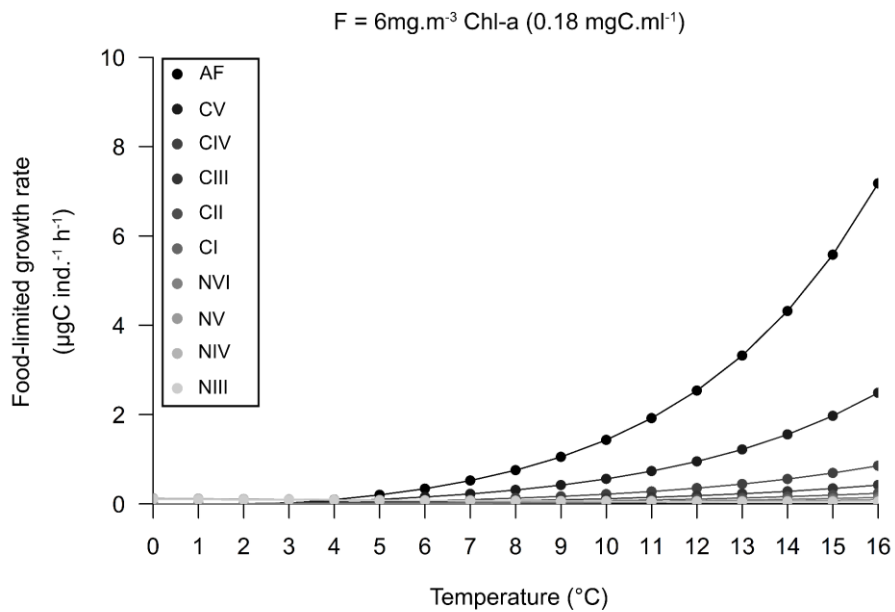


Fig. A4.1 Relationship of the food-limited growth rate and environmental temperature at lower food concentration (6 mg Chl.a m^{-3})

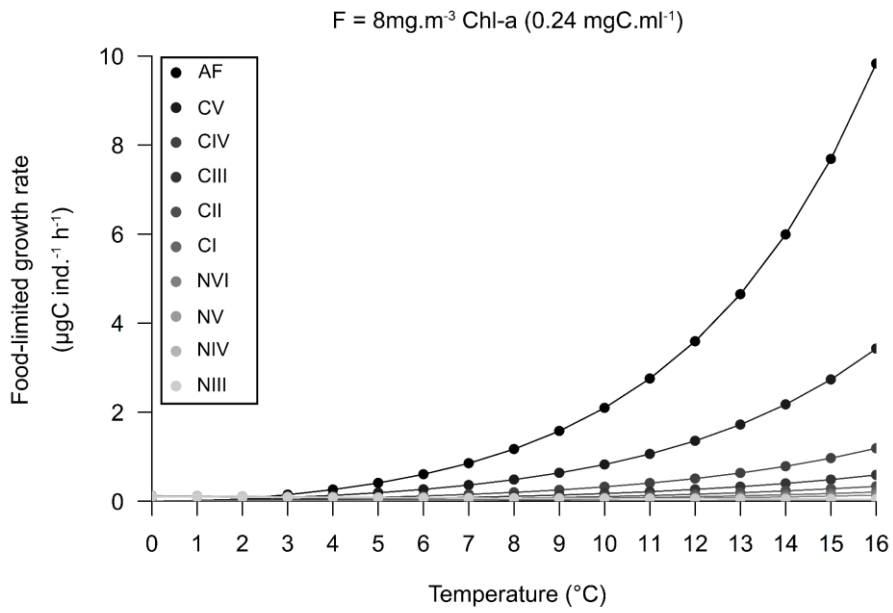


Fig. A4.2 Relationship of the food-limited growth rate and environmental temperature at a higher level of food concentration ($8\text{ mg Chl.-a m}^{-3}$)

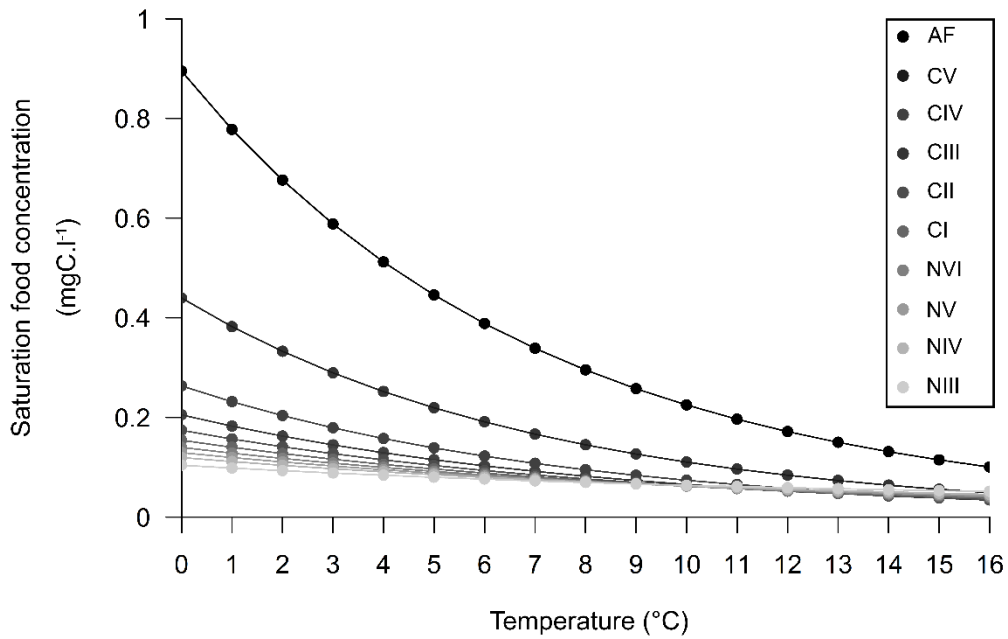


Fig. A4.3 Relationship of the saturation food concentration and environmental temperature at a fixed level of food concentration ($8\text{ mg Chl.-a m}^{-3}$)

The Fig. A4.3 above represents the Fig. 5 of Huntley and Boyd (1984). Here, the increase of the saturation food concentration (f , see Table 4 in main text) increases due to the exponential relationship between the assimilation coefficient (b , see Eq. 4, 8 and Table 4 in main text) and environmental temperature.

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Paper-III

1 **Artificial evolution of behavioral and life history strategies of Arctic *Calanus* spp. in**
2 **response to bottom-up and top-down selection pressures**

3

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12 **Abstract**

13 Strong seasonality of resources and risks act as bottom-up and top-down selection
14 pressures in high-latitudes, under which numerous behavioral and life history strategies
15 evolve. Such seasonal strategies are well-documented for high-latitude marine
16 zooplankton. However, little is known about the separate effects of bottom-up and top-
17 down selection pressures in the shaping up of their seasonal behavioral and life history
18 strategies. Here, we present a model that allows partitioning of bottom-up (i.e. food
19 availability and temperature) and top-down (i.e. visual predation risk) selection
20 pressures to study how behavioral and life history strategies evolve. In the model,
21 differential timing, amplitude and ontogenetic trajectories of diel and seasonal vertical
22 migrations were defined as behavioral strategies. Body size, generation time and birth
23 time comprised the life history strategy. Numerous combinations of behavioral and life
24 history strategies were hardwired to copepods belonging to three model-species
25 representing the Arctic *Calanus* species. In a given model environment, strategies were
26 evaluated for growth, survival and reproductive performances using a fitness estimate,
27 which was heuristically maximized using an evolutionary algorithm. Model simulations
28 were performed in three low- to high-Arctic deterministic seasonal environments at
29 various levels of visual predation risk. At lower visual predation risk, species-specific
30 behavioral and life history strategies were largely influenced by food availability and
31 temperature. However, as visual predation risk increased, the influence of bottom-up
32 selection pressures diminished, and irrespective of the modelled latitude, all model-
33 species employed largely similar strategies under top-down selection pressure. Modest
34 increase of visual predation risk increased the diel vertical migration behavior. Further
35 increase of visual predation risk was associated with the decrease of body size, which
36 created a significant impact on the observed behavioral and life history strategies
37 through allometric processes. We conclude that top-down selection pressures serve a
38 significant role in the evolution of behavioral and life history strategies of high-latitude
39 zooplankton.

40

41 Keywords: seasonal strategies, life history evolution, vertical migration, pelagic
42 environments, genetic algorithm, optimization model

43 1. Introduction

44 High-latitude pelagic environments are characterized by strong seasonal oscillations
45 of irradiance, which drives seasonal patterns of temperature, primary production and
46 predation risk. These impose strong bottom-up and top-down selection pressures on
47 pelagic inhabitants ([Hunter & Price 1992](#), [Power 1992](#), [Varpe 2017](#)) and lead to a wide
48 range of behavioral and life history adaptations ([Ohman 1988](#), [Conover 1992](#), [Szulkin et
49 al. 2006](#), [Williams et al. 2017](#)). Seasonal adaptations are usually linked with trade-offs,
50 as all adaptive traits cannot be simultaneously improved without compromising each
51 other, especially in seasonally resource-limited environments with elevated predation
52 risk ([Stearns 1989](#), [Fabian & Flatt 2012](#), [Varpe 2017](#)).

53 Seasonal behavioral and life history adaptations and associated trade-offs are well-
54 documented among marine zooplankton (reviewed in, [Conover & Siferd 1993](#), [Hagen &
55 Auel 2001](#), [Varpe 2012](#)). These involve adaptations to cope with both the productive and
56 unproductive parts of the year. During the productive season (spring–summer),
57 zooplankton tend to feed in the warmer, food-rich, near-surface waters to grow and
58 develop rapidly toward attaining sexual maturity ([Hopkins et al. 1984](#), [Huntley & Lopez
59 1992](#), [Lee et al. 2003](#), [Escribano et al. 2014](#)). However, occupation of surface waters
60 elevates the mortality risk through visual predation. This is usually countered by trading
61 off growth potential for survival through diel vertical migration (DVM) ([Lampert 1989](#),
62 [Loose & Dawidowicz 1994](#), [Hays 2003](#)). Further, structural growth of late developmental
63 stages (e.g. copepodite III onwards) of many high-latitude zooplankton is traded off to
64 build up energy reserves ([Lee et al. 2006](#)). Such trade-offs together with shorter
65 productive season and seasonal peaks of visual predation risk make it usually impossible
66 for zooplankton to attain sexual maturity and reproduce within the same calendar year
67 ([Hirche 1996b](#), [Hagen 1999](#), [Kaartvedt 2000](#), [Varpe & Fiksen 2010](#)). Instead, as the
68 unproductive season (autumn–winter) approaches, zooplankton perform seasonal
69 vertical migrations (SVM) to deeper waters and overwinter with minimal biological
70 activity (i.e. diapause, [Carlisle & Pitman 1961](#), [Hirche 1996a](#)).

71 The inability to maintain high biological activity during the unproductive part of the
72 year tends to increase generation times ([Conover 1988](#), [Falk-Petersen et al. 2009](#)) and
73 consequently elevates body sizes of most high-latitude zooplankton ([Hall et al. 1976](#),
74 [Gillooly et al. 2002](#)). Based on their inverse relationships with temperature, it can also
75 be predicted that generation time and body size of high-latitude zooplankton increase
76 from lower- to higher-Arctic locations (see, [Rohde 1992](#), [Blackburn et al. 1999](#)).
77 However, since these two life history traits show plasticity to top-down selection
78 pressures ([Brooks & Dodson 1965](#), [Gillooly 2000](#), [Jeppesen et al. 2004](#)), it is interesting
79 to investigate how growth and reproductive advantages of a longer lifespan and larger
80 body size ([McLaren 1966](#)) are traded off for survival at elevated levels of size-dependent

81 predation risk. In addition, as body size changes driven by bottom-up or top-down
82 selection pressures can directly influence physiological and behavioral activity through
83 allometric relationships ([Brown et al. 2004](#)), it is crucial to study how these processes
84 influence seasonal behavioral and life history adaptations of high-latitude zooplankton.

85 Predominantly herbivorous zooplankton occupy a crucial trophic position between
86 primary producers and higher-order consumers, and are well-suited for studying the
87 influences of bottom-up and top-down selection pressures ([Hays et al. 2005](#)). In the
88 Arctic, three congeners of *Calanus*, i.e. *C. finmarchicus*, *C. glacialis* and *C. hyperboreus*
89 usually dominate the herbivore biomass ([Eiane & Tande 2009](#)). Despite the largely
90 similar morphologies, these three species exhibit diverse behavioral and life history
91 strategies that are plastic to environmental variability (Table 1). These, coupled with
92 latitudinal patterns of species composition and abundance from lower- to higher-Arctic
93 locations ([Conover 1988](#), [Daase & Eiane 2007](#)) make *Calanus* an ideal model-species for
94 studying ecological implications of spatio-temporal dynamics of the abiotic and biotic
95 environments ([Beaugrand et al. 2002](#), [Hays et al. 2005](#), [Ji et al. 2012](#), [Espinasse et al.
96 2017](#)). In this study, we present a model of behavioral and life history strategies of Arctic
97 *Calanus* species. In the model, species-specific optimal behavioral and life history
98 strategies are artificially evolved in a deterministic model environment. By performing
99 model simulations along a latitudinal gradient at variable levels of predation risk, we
100 investigate how species-specific behavioral and life history responses of Arctic *Calanus*
101 spp. emerge in response to bottom-up and top-down selection pressures of the
102 environment.

103

104 2. Materials and Methods

105 Although the model is not strictly individual-based, it is described using the Overview,
106 Design concepts and Details (ODD) protocol ([Grimm et al. 2006](#), [Grimm et al. 2010](#)) to
107 improve reproducibility.

108

109 2.1 Purpose

110 The purpose of the model is to investigate species-specific behavioral and life history
111 responses of Arctic *Calanus* spp. against bottom-up and top-down selection pressures
112 mediated by the environment.

113

114 2.2. Entities, State variable and scales

115 The model possesses three entities: strategies, model organism and the model
116 environment.

117 Strategies are of two types: behavioral strategy and life history strategy. The
118 behavioral strategy (or vertical strategy) defines the timing, amplitude and the
119 ontogenetic trajectories of DVM and SVM. These are represented by three evolvable
120 (soft) parameters (Table 2). The life history strategy represents a collection of life history
121 traits (i.e. birth time, body mass, generation length, size at diapause onset, age and size
122 of sexual maturity, onset of spawning, breeding mode and fecundity) and their size- or
123 stage-specific variability. From these, the birth time, body size and generation length are
124 represented by three evolvable parameters (Table 2). Other life history traits emerge as
125 the evolvable parameters are optimized in the model. Strategies are hardwired to the
126 model organism, i.e. copepods belonging to three model-species.

127 The model organism characterizes hypothetical, semelparous female copepods of
128 species-CF, species-CG and species-CH, that represent the Arctic *Calanus* complex (*C.*
129 *finmarchicus*, *C. glacialis* and *C. hyperboreus*). Although these species are distributed
130 throughout the Arctic, only *C. glacialis* and *C. hyperboreus* are considered as species with
131 a true Arctic origin, where *C. finmarchicus* is a boreal species that primarily inhabit the
132 North Atlantic ([Fleminger & Hulsemann 1977](#), [Conover 1988](#)). *Calanus* spp. possess a 13-
133 developmental stage ontogeny, which includes an embryonic stage, six naupliar stages
134 (NI–NVI), five copepodite (CI–CV) stages and an adult. Older developmental stages can
135 store lipids, which act as energy reserves that are used to meet the metabolic demands
136 during the diapause ([Hirche 1996a](#), [Hagen & Auel 2001](#)). However, overwintering stage
137 composition, size of energy reserves and potential diapause duration varies between
138 species (Table 1 and see also, [Falk-Petersen et al. 2009](#), [Maps et al. 2013](#)). Reproduction
139 of *Calanus* spp. usually occur in the spring, but the timing and the sources of energy

140 allocation vary between the three species (Table 1). *C. hyperboreus* has the longest life
141 cycle duration (usually 3 years) and largest body size, while the relatively small *C.*
142 *finmarchicus* and *C. glacialis* usually complete their life cycles within 1 or 2 years (Table
143 1).

144 The model runs in three 1000-m deep artificial seasonal environments that roughly
145 represent the expected environmental variability along a latitudinal gradient extending
146 from the north Atlantic to the Arctic (ca. 60–80 °N). These model environments do not
147 refer to specific geographic locations, but the modeled environmental parameters were
148 adopted from field measurements taken in this region (Appendix A1). Since water mass
149 characteristics of this region (e.g. [Swift 1986](#)) were not modelled for simplicity, the
150 model environments represent typical annual oceanographic characteristics of deep
151 Arctic fjords ([reviewed in Cottier et al. 2010](#)). The Environment-L characterizes the lower
152 end of the modelled latitudinal gradient (ca. 60°N). Here, the modelled irradiance,
153 temperature and primary production show pronounced seasonal and vertical variability
154 (Fig. 1A–C), but are assumed constant between years. The modelled sea-surface
155 irradiance follows the global clear-sky horizontal irradiance formulations of [Robledo and](#)
156 [Soler \(2000\)](#), and peaks at ca. 1500 $\mu\text{mol m}^{-2} \text{s}^{-1}$ (Fig. 1A). The modelled temperature
157 peaks at 15°C in the summer and distributes evenly across the surface mixed layer (Fig.
158 1B). The depth of surface mixed layer follows the seasonal pattern described by [Mann](#)
159 [and Lazier \(2006\)](#), and reaches a maximum of 500 m during the winter. Temperature
160 below the mixed layer decreases with depth and converges to a minimum of 2°C. The
161 pelagic primary production extends from mid-February to late-September, with a
162 chlorophyll-*a* peak (6 mg m⁻³) in mid-April (Fig. 1C). We manipulated the environmental
163 parameters of Environment-L to formulate two additional seasonal environments, i.e.
164 Environment-M (ca. 70°N, Fig. 1D–F) and Environment-H (ca. 80°N, Fig. 1G–I),
165 representing the mid-point and the higher end of the modelled latitudinal gradient
166 (Appendix A1). For simplicity, we did not model the sea ice dynamics at any of these
167 higher-latitude environments.

168 Copepods are characterized by four state variables: vertical location (depth),
169 structural body mass, energetic reserve and developmental stage. The model has a
170 temporal coverage of an annual cycle and a unidimensional (vertical) spatial coverage
171 of 1000 m. The time and space consist of 1 h and 1 m discrete intervals.

172

173 2.3 Process overview and scheduling

174 In each timestep, the model follows the evolvable proxies of behavioral and life
175 history strategies hardwired to copepods spawned in different times of the year and
176 simulates their growth, survival and reproduction separately for each model-species.

177 State variables are updated simultaneously. Strategies are evaluated using a fitness
178 function based on the expected survival and reproductive performances. The fitness is
179 heuristically maximized using a Genetic Algorithm ([Holland 1975](#)) to derive species-
180 specific optimal behavioral and life history strategies for a given set of environmental
181 conditions (Fig. 2).

182

183 2.4 Design concepts

184 2.4.1 Basic principles

185 The modeling framework follows a previous approach from [Bandara et al. \(2018\)](#) to
186 model DVM and SVM over the entire life cycle of a high-latitude copepod in higher
187 spatial and temporal resolution. In the above model, copepod diel and seasonal vertical
188 behavior were modelled as strategies that maximize fitness (hence, the term vertical
189 strategies). Vertical strategies were represented using multiple proxies that define the
190 timing and amplitude (vertical extent) of DVM and SVM. From a large number of vertical
191 strategies seeded at different times of the year, an optimal vertical strategy that
192 maximizes fitness was heuristically estimated for a given set of environmental
193 conditions. The present model was implemented to address a key limitation of the
194 above, which is the lack of species-specific patterns of behavioral and life history
195 strategies (Table 1). We built this to the present model via two design changes: first, the
196 species-specific parameterization of growth survival and reproductive processes
197 (section 2.6), and second, the inclusion of two evolvable parameters to allow the body
198 size and generation time to emerge (Table 2, Fig. 2). The complex multi-dimensional
199 optimization problem ensued was heuristically solved using a Real-Coded Genetic
200 Algorithm ([Davis 1989](#), [Lucasius & Kateman 1989](#), [Herrera et al. 1998](#)).

201 The strategy-oriented construct of the present model shares the same limitations of
202 [Bandara et al. \(2018\)](#). First, sacrificing biological (= individual) resolution to
203 accommodate higher spatial and temporal resolution, and second, the lack of
204 population level responses such as density dependence. Consequently, the modelled
205 behavioral and life history strategies do not show quantitative feedbacks with the model
206 environment (e.g. impact of grazing on food concentration and duration of the
207 productive season).

208

209 2.4.2 Emergence

210 The behavioral and life history strategies emerging from the model are presented in
211 Fig. 2 and described in Table 3.

212

213 2.4.3 Adaptation and sensing

214 Copepods are sensitive to their internal states (i.e. structural body mass, mass of the
215 energetic reserve developmental stage) and external stimuli (i.e. irradiance,
216 temperature, food concentration and depth). Altogether, these determine the size- or
217 stage-specific patterns of growth, survival and reproduction (section 2.6).

218

219 2.4.4 Objectives

220 The model uses a fitness estimate that evaluates the expected reproduction and
221 survival performances rendered by different behavioral and life history strategies
222 (section 2.6.4).

223

224 2.4.5 Prediction and stochasticity

225 The vertical position of eggs within the convective mixed layer and the overwintering
226 depth selection are modelled as stochastic processes (section 2.6.2.4). Further,
227 stochasticity plays a central role in the model initialization (section 2.5) and selection,
228 recombination and mutation operators of the GA (section 2.6.4).

229

230 2.4.6 Observations

231 For a given model environment, the model produces heuristic estimates of the
232 optimal behavioral and life history strategies (Fig. 2, Table 3).

233

234 2.5 Initialization

235 The model initializes with seeding of $N (= 2.5 \times 10^6)$ eggs at random times of the year
236 to random depths (< 50 m) of the water column. Each seed represents an embryonic
237 stage of a copepod following a specific behavioral (vertical) strategy, ontogenetic body
238 mass trajectory and a generation time. These are determined by randomly assigning
239 values to the evolvable parameters listed in Table 2.

240

241 2.6 Submodels

242 2.6.1 Growth and development

243 We modelled species-specific somatic growth in Carbon units following a simple
244 growth formulation, which defines the growth rate (G , $\mu\text{g C ind}^{-1} \text{h}^{-1}$) as the balance
245 between assimilation and metabolic rates ([Pütter 1920](#), [Von Bertalanffy 1938](#)) as,

$$G_{i,s,t,z} = a \cdot A_{i,s,t,z} - B_{i,s,t,z} \quad \text{Eq. 1}$$

246 Here, the assimilation rate is a product of the ingestion rate (A , $\mu\text{g C ind}^{-1} \text{h}^{-1}$) and the
247 assimilation coefficient (a) [Huntley and Boyd \(1984\)](#), where B ($\mu\text{g C ind}^{-1} \text{h}^{-1}$) is the
248 metabolic rate (section 2.6.2.6). Further, i is the individual, s is the species, t is time and
249 z is depth (definitions, units and references of all the terms are listed in Table 4). At a
250 hypothetical reference temperature -2°C , the ingestion rate relates with the structural
251 mass (W_c , $\mu\text{g C}$) as,

$$A_{i,s,t} = b_s \cdot (W_c)_{i,t}^{m_s} \quad \text{Eq. 2}$$

252 where b and m are species-specific mass coefficient and exponent of ingestion (Table 5).
253 The ambient temperature (T , $^\circ\text{C}$) elevates the ingestion rate following the exponential
254 function,

$$A_{i,s,t,z} = A_{i,s,t} \cdot c_s \cdot \exp(n_s \cdot T_{t,z}) \quad \text{Eq. 3}$$

255 where c and n are species-specific temperature coefficient and exponent of ingestion
256 (Table 5). Parameter values for coefficients and exponents of body mass and
257 temperature were estimated following the growth model of [Maps et al. \(2011\)](#) (Table 5,
258 Appendix A2). The temperature-dependent ingestion rate is scaled by the ambient food
259 concentration (F , $\mu\text{g C l}^{-1}$) into a range of 0–1 following a Holling's type-II (disk) function
260 ([Holling 1959](#)) as,

$$A_{i,s,t,z} = A_{i,s,t,z} \cdot \frac{d_{i,t} \cdot F_{t,z}}{1 + d_{i,t} \cdot F_{t,z}} \quad \text{Eq. 4}$$

261 Here, the parameter d (selected range = 0.1–0.3) defines the food concentration at
262 which the asymptotic value of the above relationship is reached (Fig. 3A, B), and relates
263 with the structural mass (Fig. 3C) as,

$$d_{i,t} = 0.3 \cdot (W_c)_{i,t}^{-0.138} \quad \text{Eq. 5}$$

264 This produces size-specific satiation food concentrations in the range of 75–125 $\mu\text{g C l}^{-1}$,
265 (Fig. 3A, B) which are comparable to those estimated by [Huntley and Boyd \(1984\)](#),

266 [Campbell et al. \(2001\)](#) and [Maps et al. \(2011\)](#). In the above calculations, we used Chl.-
 267 α :C ratio of 30 ([Båmstedt et al. 1991](#), [Sakshaug et al. 2009](#)).

268 This growth submodel cannot be applied to first and second nauplii stages (NI and NII)
 269 which do not feed, but utilize the reserves from the embryo to meet energetic demands
 270 ([Marshall & Orr 1972](#), [Mauchline 1998](#)). Although catabolization of reserves may lead to
 271 loss of body mass of non-feeding stages ([Maps et al. 2011](#)), for simplicity, we assumed
 272 that the structural masses of NI and NII remain constant during development.

273 The temperature-dependent development times (h) of eggs and non-feeding NI and
 274 NII stages are estimated following a Bělehrádek function ([Corkett et al. 1986](#)) as,

$$D_{s,t,z} = 24 \cdot [q_s \cdot (T_{t,z} + r_s)^{-2.05}] \quad \text{Eq. 6}$$

275 Here, species-specific values for the parameters q and r were adopted from [Campbell et](#)
 276 [al. \(2001\)](#) and [Ji et al. \(2012\)](#) (Table 5). The development of feeding stages (NIII–Adult,
 277 i.e. from stage j to $j + 1$, where $4 \leq j \leq 12$) occurs only if the structural mass (W_c) exceeds
 278 a stage-specific critical molting mass (W_j , $\mu\text{g C}$) as,

$$j_{i,s} = \begin{cases} j_{i,s} + 1 & \text{if } (W_c)_{i,s,t} > (W_j)_{i,s} \\ j_{i,s} & \text{if } (W_c)_{i,s,t} \leq (W_j)_{i,s} \end{cases} \quad \text{Eq. 7}$$

279 For each model environment, species-specific maximum and minimum estimates of W_j
 280 (W_j^{min} and W_j^{max}) were estimated following the growth model of [Maps et al. \(2011\)](#) (Fig.
 281 4, Appendix A2). To maintain the intra-specific plasticity of body sizes in the model, we
 282 introduced an evolvable parameter α (body size parameter, range = 0–1, Table 2), which
 283 defines the stage-specific critical molting masses of any given copepod as,

$$(W_j)_{i,s} = (W_j^{min})_{i,s} + [(W_j^{max})_{i,s} - (W_j^{min})_{i,s}] \cdot \alpha_i \quad \text{Eq. 8}$$

284 Therefore, based on the parameter value of α , the ontogenetic body mass trajectories
 285 of copepods tend to occupy a fixed fraction of the environment- and species-specific
 286 ranges (Fig. 4).

287

288 2.6.2 Survival

289 2.6.2.1 Predation risk

290 We modelled the predation risk (M_v) as a probability function. Here, the depth-
291 specific visual predation risk scales with the downwelling irradiance (I) following [Eiane](#)
292 [and Parisi \(2001\)](#) as,

$$I_{t,z} = I_{t,0} \cdot \exp(-\psi \cdot z) \quad \text{Eq. 9}$$

293 where I_z and I_0 are irradiance at depth z and surface at a given time, and ψ ($= 0.06 \text{ m}^{-1}$)
294 is the attenuation coefficient for downward directed irradiance in the water column. To
295 express this as a probability, we remapped the downwelling irradiance (I) in a range
296 between 0.1–0.9 (I') so that visual predation risk offers non-zero probability of survival
297 at the highest possible irradiance level, and non-zero probability of death at the lowest
298 level, expressed as,

$$(M_v)_{i,t,z} = I'_{t,z} \cdot K_{i,t} \quad \text{Eq. 10}$$

299 Here, K is a parameter that scales the visual predation risk to produce hourly estimates
300 of mortality.

301 The detection efficiency of visually orientating planktivores increases with the
302 size of their prey ([Brooks & Dodson 1965](#), [De Robertis 2002](#), [Aljetlawi et al. 2004](#)). Most
303 metrics of detection efficiency, such as predator visual range, reaction distance and
304 electivity index are modelled in a way that it increases rapidly with the initial increase of
305 prey size, while reaching a summit or a plateau as prey size increases further (e.g. [Zaret](#)
306 [& Kerfoot 1975](#), [Confer et al. 1978](#), [Pastorok 1981](#), [Aksnes & Giske 1993](#)). This is likely
307 due to elevated handling time, prey escape responses and gape-limitations driven by
308 larger prey sizes ([Werner 1974](#), [Fields & Yen 1997](#), [Devries et al. 1998](#), [Kiørboe 2011](#)).
309 We followed this logic and modeled the size-dependent visual predation risk as an
310 asymptotic exponential relationship between the body mass (W_c) and K (Fig. 5A),
311 assuming that the largest developmental stage (female of Species-CH) is ca. 25 times
312 more vulnerable to visual predation risk compared to the smallest developmental stage
313 (eggs of species-CF) (Fig. 4, Fig. 5D–E). This scaling accounts for the inclusion of *C.*
314 *hyperboreus* in this model (represented by model-species-CH), compared to a previous
315 model of smaller *C. finmarchicus* and *C. glacialis* ([Bandara et al. 2018](#)), which used a
316 maximum 10-fold size-dependent visual predation risk scaling.

317 We modeled the mortality risk imposed by the non-visual predators (M_n) constant
318 over time and depth ([Eiane & Parisi 2001](#)), and assumed to account for 10% of the K .

319

320 2.6.2.2 Diel vertical migration

321 We used the photoreactive behavior as a proxy to estimate the timing and amplitude
322 of DVM (e.g. [Kerfoot 1970](#), [Carlotti & Wolf 1998](#)). Here, an evolvable parameter β
323 (irradiance threshold parameter, Table 2) defines an irradiance threshold, above which
324 a negative phototactic response on the vertical swimming behavior is induced ([Båtnes et](#)
325 [al. 2015](#), [Cohen et al. 2015](#)). Consequently, at any given time, copepods occupy a depth
326 with an irradiance level ($I_{t,z}$) below β . From all possible depth bins that satisfy the $I_{t,z} < \beta$
327 condition, we assumed that copepods occupy the depth that maximizes the growth
328 potential (Eq. 1). We predicted this depth deterministically, assuming that copepods are
329 neutrally buoyant and swim vertically in the water column at a constant cruising velocity
330 (U , mh^{-1}) obtained from [Bandara et al. \(2018\)](#) as,

$$U_{i,t} = 8.0116 \cdot (W_c)_{i,t}^{0.4531} \quad \text{Eq. 11}$$

331 For simplicity, we further assumed that internal state-dependent factors, such as hunger
332 and satiation have a negligible influence on the modelled DVM.

333 To represent the size- or stage-specific variability of DVM (e.g. [Zaret & Kerfoot 1975](#),
334 [Huntley & Brooks 1982](#), [Hays 1995](#), [Eiane & Ohman 2004](#)), we defined an irradiance
335 sensitivity parameter (L , selected range = 1–2.5) that relates positively with structural
336 mass (W_c) following an asymptotic exponential relationship (Fig. 5B). The size-
337 dependent increase of irradiance sensitivity causes the irradiance threshold parameter
338 (β) to decrease as,

$$\beta_{i,t} = \beta_i \cdot \frac{1}{L_{i,t}} \quad \text{Eq. 12}$$

339 The minimum irradiance sensitivity thresholds produced by this model (i.e. $1.4 \times 10^{-7} \mu$
340 $\text{mol m}^{-2} \text{s}^{-1}$ for species-CF, $5.92 \times 10^{-8} \mu \text{mol m}^{-2} \text{s}^{-1}$ for species-CG and $3.2 \times 10^{-8} \mu \text{mol m}^{-2}$
341 s^{-1} for species-CH) agree with those published for *Calanus* spp. by [Båtnes et al. \(2015\)](#).

342

343 2.6.2.3 Energy storage

344 The developmental stage CIV and CV of species-CF and -CG, and stages CIII, CIV and
345 CV of species-CH can allocate an evolvable fraction γ (energy allocation parameter,
346 Table 2) from surplus growth to build up an energy reserve. The reserve can occupy up
347 to 70% of the structural mass ([Fiksen & Carlotti 1998](#), [Jónasdóttir 1999](#)). As a
348 consequence of the body size plasticity allowed in this model, it was observed in the trial
349 runs that copepods always followed the lowest body mass trajectories and overwintered
350 (see below) at a significantly smaller size (W_c ca. $10 \mu\text{g C}$ for species-CF). This not only

351 disagrees with the body mass estimates of overwintering *Calanus* spp. ([Båmstedt et al.](#)
 352 [1991](#), [Pepin & Head 2009](#)), but undermines the concept that a reasonable structural
 353 mass should be attained to allow space for lipid sac to be harbored ([Miller et al. 2000](#),
 354 [Lee et al. 2006](#)). Therefore, we defined a minimal structural mass (W_c , $\mu\text{g C}$) below which
 355 no stores can be maintained as,

$$(W_s)_{i,s,t,z} = \begin{cases} 0 & \text{if } (W_c)_{i,t} < 38 \\ G_{i,s,t,z} \cdot \gamma_i \cdot \left[\frac{1}{1 + \exp((600 - (W_c)_{i,t})/400)} \right] & \text{if } 38 < (W_c)_{i,t} < 159 \\ G_{i,s,t,z} \cdot \gamma_i & \text{if } (W_c)_{i,t} \geq 159 \end{cases} \quad \text{Eq. 13}$$

356 where W_s ($\mu\text{g C}$) is the mass of the energy reserve and G is the surplus growth (Eq. 1).
 357 Here, energy storage capacity exponentially increases from 38 $\mu\text{g C}$ and reaches a
 358 horizontal asymptote at 159 $\mu\text{g C}$ (Fig. 5C). These lower and upper W_c values were
 359 estimated from lipid sac volume to body size relationships published by [Miller et al.](#)
 360 [\(2000\)](#) and [Vogedes et al. \(2010\)](#).

361

362 2.6.2.4 Seasonal vertical migration

363 We used the state of the energetic reserve as a proxy of timing of the SVM ([Visser &](#)
 364 [Jónasdóttir 1999](#)). Here, copepods descend to an overwintering depth when the stores
 365 account for an evolvable fraction δ (seasonal descent parameter: Table 2) of the
 366 structural mass. For simplicity, we made three general assumptions for selecting
 367 overwintering depths. First, copepods always overwinter below the maximum depth of
 368 the convective mixed layer (i.e. 500 m) to avoid being circulated back to the surface
 369 ([Visser & Jónasdóttir 1999](#), [Irigoién 2004](#)). Second, the specific overwintering depth
 370 below the mixed layer is selected by a gaussian distribution (mean = 750, SD = 50). Third,
 371 internal and external environmental variability has no influence on the overwintering
 372 depth selection. Although the third assumption does not hold true in nature (e.g. [Hirche](#)
 373 [1991](#), [Kaartvedt 1996](#), [Astthorsson & Gislason 2003](#)), we used it here for simplicity,
 374 because it was shown by [Bandara et al. \(2018\)](#) that the use of an evolvable overwintering
 375 depth parameter had a little influence on the fitness and phenology of the modeled
 376 copepod. After descending to overwintering depths, copepods switch to a diapause
 377 state ([Carlisle & Pitman 1961](#), [Hirche 1996b](#)), where growth, development, vertical
 378 movements and reproduction cease (see also section 2.6.2.6). The diapause terminates
 379 and copepods ascend from overwintering depths upon exhausting an evolvable fraction
 380 ε from the energy reserve (seasonal ascent parameter, Table 2).

381

382 2.6.2.5 Generation time

383 We introduced an evolvable parameter η (generation time parameter, Table 2) to
384 represent the variability of generation times commonly reported for *Calanus* spp. (Table
385 1). Here, η ranges between 1–3, which indicates the generation time in number of years.
386 This shows species-specific patterns, where $\eta = 1$ for species-CF, η ranges between 1–2
387 for species-CG and 1–3 for species-CH (cf. Table 1). Generation times > 1 year are
388 characterized by several subsequent seasonal migrations, which follows the same
389 patterns of energy allocation, and same proxies of ascent and descent described above.
390 After the final diapause, copepods do not allocate surplus growth to maintain energy
391 reserves.

392

393 2.6.2.6 Metabolism

394 The metabolic rate (B , $\mu\text{g C}$) is the sum of the basal metabolic rate (B_b) and the active
395 metabolic rate (B_a). At the hypothetical reference temperature of -2°C , B_b relates with
396 the total body mass ($W = W_c + W_s$) as,

$$(B_b)_{i,s,t} = f_s \cdot (W)_{i,t}^{o_s} \quad \text{Eq. 14}$$

397 where f and o are mass coefficient and exponent of respiration (Table 5). Ambient
398 temperature elevates B_o following the exponential function,

$$(B_b)_{i,s,t,z} = (B_b)_{i,s,t} \cdot g_s \cdot \exp(p_s \cdot T_{t,z}) \quad \text{Eq. 15}$$

399 where g and p are temperature coefficient and exponent of metabolism (Table 5).
400 Parameter values for above respiration coefficients and exponents were estimated from
401 [Maps et al. \(2011\)](#) (Appendix A2). In the model, B_a consists of swimming costs (i.e.
402 metabolic costs of vertical migrations) and assumed to be 150% of the B_o ([Bandara et al.](#)
403 [2018](#)). During diapause, B_a is nullified (since copepods are assumed stagnant) and the B_b
404 is assumed to reduce by 75% in all species ([Maps et al. 2010](#)).

405

406 2.6.2.7 Starvation risk

407 Metabolic demands that cannot be sustained by food intake are balanced by energy
408 reserves. In case of an absent or depleted energy reserves, structural mass is
409 catabolized. This induces mortality risk through starvation (starvation risk, M_s).
410 However, we assumed that copepods are tolerant to modest ($< 10\%$) loss of structural
411 mass ([Threlkeld 1976](#)). Structural catabolization beyond the above threshold causes the

412 starvation risk to increase linearly and peaks as 50% of structural mass is lost and causes
 413 death ([Bandara et al. 2018](#)) as,

$$(M_s)_{i,t} = \begin{cases} 0 & \text{for } (W_x)_{i,t} \leq 0.1 \\ 2 \cdot (W_x)_{i,t} & \text{for } 0.1 < (W_x)_{i,t} \leq 0.5 \\ 1 & \text{for } (W_x)_{i,t} \geq 0.5 \end{cases} \quad \text{Eq. 16}$$

414 Here, W_x ($\mu\text{g C}$) is the catabolized structural mass expressed as a proportion of the
 415 maximum structural mass prior to structural catabolization.

416

417 2.6.3 Reproduction

418 In this model, somatic growth of copepods ceases after the final molt (e.g. [Fiksen &](#)
 419 [Giske 1995](#), [Fiksen & Carlotti 1998](#), [Varpe et al. 2007](#)), and the matter gained through
 420 feeding and catabolizing energy reserves is only allocated for meeting metabolic
 421 demands and reproduction. We modeled the energy input to reproduction as a species-
 422 specific process. Here, the reproduction of species-CF represents the pure income
 423 breeding strategy of *C. finmarchicus* (Table 1), where the energy input is sourced solely
 424 from food intake. Reproduction of species-CH represents the pure capital breeding
 425 strategy of *C. hyperboreus* (Table 1), where the energy input is sourced entirely from
 426 remaining reserves, by allocating a specific amount of matter (C) equivalent to the
 427 temperature-dependent growth rate (Eq. 1–3) from the remaining energetic reserve.
 428 Species-CG represents an intermediate reproductive strategy alike *C. glacialis* (Table 1),
 429 where energy inputs for reproduction may be sourced from both food-intake and energy
 430 reserves. The fecundity (R) from is estimated using the matter allocated to egg
 431 production (W_R , $\mu\text{g C}$) and the species-specific unit egg mass (W_E , $\mu\text{g C}$) as,

$$R_{i,s,t,z} = \frac{(W_R)_{i,s,t,z}}{(W_E)_s} \quad \text{Eq. 17}$$

432 Here, the W_E vary with the species (Table 5), but we assumed that it is not affected by
 433 environmental variability.

434

435 2.6.4 Fitness function and optimization

436 To evaluate the performance of behavioral and life history strategies, we used the
 437 fitness estimate (Ω) developed by [Bandara et al. \(2018\)](#) as,

$$\Omega_{i,s} = \left(\sum_{t_B}^{t_X} H_{i,t,z} \cdot R_{i,s,t,z} \right) \cdot \omega \quad \text{Eq. 18}$$

438 where H is the survivorship, i.e. the probability of survival from birth (t_B) to a given time
439 horizon (t_X), estimated as a function of visual, non-visual and starvation risks (M_v , M_n and
440 M_s) as,

$$H_{i,t,z} = \prod_{t_B}^{t_X} 1 - [(M_v)_{i,t,z} + (M_n)_{t,z} + (M_s)_{i,t}] \quad \text{Eq. 19}$$

441 The fitness is adjusted using a binary weight ω ([Bandara et al. 2018](#)), as the fitness
442 function is not robust to evaluate generation times less than one year.

443 We used a Real-Coded Genetic Algorithm to derive environment- and species-specific
444 heuristic estimates of the optimal behavioral and life history strategies that maximize
445 fitness. In the RCGA, the seven proxies of behavioral and life history strategy (Table 2)
446 are considered as genes on a single chromosome. The optimization process begins by
447 selecting a mating pool of N chromosomes (parents) from the initial seeds using a binary
448 (two-way) deterministic tournament ([Goldberg & Deb 1991](#), [Miller & Goldberg 1995](#)).
449 Genes of two randomly selected parents from the mating pool are recombined following
450 the Laplace crossover method ([LX, Deep & Thakur 2007](#)), which produces two offspring
451 (recombinants). Genes of the recombinants are mutated at a probability of 0.02 per-
452 gene following the Makinen, Periaux and Toivanen mutation ([MPTM, Toivanen et al.](#)
453 [1999](#)). The population of strategies resulting from these operations comprises of N
454 parents, whose fitness is known and N offspring, whose fitness is not yet known. Parents
455 with unique gene combinations are selected to construct a reference library, which is
456 updated at each iteration. Each offspring is compared with those in the reference library
457 to assess their fitness. Fitness of the offspring with similar gene combination to those in
458 the reference library are assigned *in-situ*, while the rest goes through the life cycle
459 simulation to determine fitness. Once the fitness of all $2N$ individuals are known, N
460 survivors are selected following a round-robin (all-play-all) tournament of size 10 ([Harik](#)
461 [et al. 1997](#), [Eiben & Smith 2003](#)). This process is repeated for a minimum of 400
462 iterations, and terminated when the mean fitness of the population of strategies shows
463 no improvement for a 100 consecutive iterations thenceforth ([Eiben & Smith 2003](#)).

464

465 2.7 Model development, execution and analysis

466 The model was developed, executed and analyzed using the R™ v.3.3.1 ([R Core Team](#)
467 [2016](#)) and RStudio™ integrated development environment (IDE) v.1.0.136 ([RStudio](#)
468 [Team 2016](#)), along with the high-performance computing (HPC) packages Rcpp
469 ([Eddelbuettel et al. 2011](#)) and bigmemory ([Kane et al. 2013](#)).

470 A basic run (BR) with default values for model parameters (Table 4) was performed
471 for each model-species in the Environment-L. To test the influence environmental

472 variables on the predicted behavioral and life history strategies, we performed a
473 sensitivity analysis following [Jørgensen and Bendoricchio \(2001\)](#), which produces a
474 sensitivity scores (S_x) as,

$$S_x = \frac{(X_{BR} - X_M)/X_{BR}}{(P_{BR} - P_M)/P_{BR}} \quad \text{Eq. 20}$$

475 where X is the predicted model output of the basic run (X_{BR}) and the modified run (X_M)
476 for a given change ($\pm 25\%$) of input parameter value between the basic run (P_{BR}) and the
477 modified run (P_M). Sensitivity analysis was performed separately for each model-species.

478 By performing model runs along the modelled latitudinal gradient at variable levels of
479 visual predation risks we investigated how species-specific behavioral and life history
480 strategies emerge under the influences of bottom-up (i.e. temperature and food
481 availability) and top-down (predation risk) selection pressures ([cf. Bandara et al. 2018](#)).
482 Although the modelled food concentration was constant across the model
483 environments, the decreasing duration of the modelled productive season and
484 temperatures ensued a decreasing gradient of growth potential from lower latitude
485 Environment-L to higher latitude Environment-H (Fig. 1, Appendix A1). A gradient of
486 visual predation risks was created by varying the scalar K in between 10^{-6} – 10^{-2} (i.e. 10^{-6} ,
487 10^{-5} , 10^{-4} , 5×10^{-4} , 10^{-3} , 2.5×10^{-3} , 5×10^{-3} , 7.5×10^{-3} and 10^{-2}). To enhance visualization,
488 we transformed the visual predation risk scalar (K) to its fourth root (K').

489 As Genetic Algorithms produce heuristic estimates of the maximum fitness, there is
490 no guarantee that it would converge on the global maximum given a potentially diverse
491 fitness landscape ([Zanakis & Evans 1981](#), [Rardin & Uzsoy 2001](#), [Strand et al. 2002](#), [Record
492 et al. 2010](#)). Therefore, we replicated each model run 10 times with different starting
493 values for the evolvable parameters (Table 2) to check if the algorithm converges on the
494 same set of solutions. As the optimized parameter values showed little variability
495 between replicate runs ($< 7\%$), we used the mean of the replicates for each parameter
496 for analyses.

497

498 3. Results and Discussion

499 3.1 Emergent strategies of the basic run

500 An annual life cycle was predicted for all modelled species in the basic run (*BR*).
501 However, the model-predicted optimal behavioral and life history strategy of species-CF
502 was distinctly different from the other model-species.

503 The predicted optimal birth times for species-CF occurred in mid-June, when the
504 irradiance and the temperature of the model environment (Environment-L) were at its
505 peak, and the food concentration had decreased by ca. 50% compared to its annual
506 maximum (Fig. 6A). This seems counterintuitive, as the optimal birth time coincides the
507 annual maximum of visual predation risk (Eqs. 9 and 10). However, in the model,
508 species-CF possesses the smallest body size (Fig. 4), and hence is the least vulnerable to
509 visual predation risk (Fig. 5A, F). In addition, due to the smaller size, species-CF become
510 satiated at lower food concentrations (Fig. 5C, F) and is less likely to suffer from food-
511 limitation (Fig. 6L). Therefore, it is likely that the smaller body size of species-CF allowed
512 it to utilize higher summertime temperatures to grow and develop faster (Fig. 6P). The
513 higher growth rates of late developmental stages (i.e. copepodite stage IV onwards)
514 likely dampened by two trade-off strategies. First, to minimize the visual predation risk,
515 relatively large developmental stages had to periodically abandon food-rich near-
516 surface waters to perform DVM (down to 50–60 m, Fig. 6D, H). DVM leads to reduced
517 growth and development rates ([Houston et al. 1993](#), [Bandara et al. 2018](#)), as growth
518 potential is traded off for survival ([Lampert 1989](#), [Hays 2003](#)). Second, in order to survive
519 the winter, CIV and CV stages had to allocate a fraction of surplus growth to build up
520 energy reserves. The optimal energy allocation parameter (γ , Table 2) predicted for
521 species-CF in the *BR* was 0.1, which translates to a 10% decrease in structural growth.
522 Consequently, species-CF grazed toward the end of the productive season and
523 descended to overwintering depths in late-July (Fig. 6A, E) with partly filled energy
524 reserves ($W_s/W_c = 0.54$, where maximum = 0.70, Fig. 6I). The ascent from overwintering
525 depths occurred in late-January of the following year, before the primary production
526 had commenced, and while the other model-species were still in diapause (Fig. 6A, E–
527 G). This doesn't agree with the well-established notion that the seasonal ascent of *C.*
528 *finmarchicus* occurs after those of *C. glacialis* and *C. hyperboreus* (e.g. [Madsen et al.](#)
529 [2001](#), [Astthorsson & Gislason 2003](#), [Søreide et al. 2008](#), [Bandara et al. 2016](#)). However,
530 the predicted early ascent of species-CF appears to possess an advantage, as it used the
531 post-overwintering surplus energy reserves for structural growth to reach sexual
532 maturity by the time when the pelagic algal bloom had commenced (W_c of overwintering
533 CV $\approx 258 \mu\text{g C}$, W_c at sexual maturity $\approx 271 \mu\text{g C}$, Fig. 6I). This somewhat resembles the
534 pre-bloom seasonal ascent and spawning patterns observed for *C. finmarchicus* ([Diel &](#)
535 [Tande 1992](#), [Melle & Skjoldal 1998](#), [Richardson et al. 1999](#)).

536 These predictions about species-CF in the *BR* points to a life strategy that attempts to
537 elevate structural growth at the expense of energy reserves. This strategy is expected
538 from a species which does not use energy reserves for egg production, such as *C.*
539 *finmarchicus* ([Tande et al. 1985](#), [Niehoff et al. 2002](#), [Madsen et al. 2008](#)). However, it
540 may also be that the optimized strategy for species-CF in the *BR* is an attempt to attain
541 sexual maturity and reproduce within the same productive season. This resembles the
542 life cycle of *C. finmarchicus* in lower latitudes, where it completes several generations
543 per year (e.g. [Fish 1936](#), [Lie 1965](#), [Matthews et al. 1978](#), [Gislason & Astthorsson 1996](#),
544 [McLaren et al. 2001](#), [Bagøien et al. 2012](#)). As our model does not allow generation times
545 < 1 year to be simulated, the ability to maintain multiple generations per year and its
546 adaptive significance in the modelled environments remain unclear.

547 The model-predicted birth times for species-CG and -CH occurred ca. 1 month
548 earlier than species-CF between late-April and May. Unlike species-CF, these two model-
549 species did not employ early birth as a strategy to utilize the seasonal temperature peak
550 to attain higher growth rates. This was likely caused by the increased vulnerability to
551 visual predation risk and the higher satiation food concentrations associated with their
552 relatively large body sizes (Figs. 3, 5A, E and F). Consequently, species-CG and -CH were
553 characterized by slower growth rates, pronounced food limitation and 1.5–2 times
554 longer development times compared to species-CF (Fig. 6L and P). Possibly due to
555 occupying a time of the year with lower growth potential, their DVM did not increase
556 much compared to species-CF (Fig. 6D and H), in line with observations made by
557 numerous other empirical and modeling work (e.g. [Hardy & Gunther 1935](#), [Huntley &](#)
558 [Brooks 1982](#), [Andersen & Nival 1991](#), [Fiksen & Giske 1995](#), [Tarling et al. 2000](#), [Bandara](#)
559 [et al. 2018](#)). However, irrespective of the lower growth rates, species-CG and -CH
560 descended to overwintering depths with maximum possible energy reserves ($W_s/W_c =$
561 0.70 , Fig. 6J, K). To attain such large reserve loads, older developmental stages (CIII, CIV
562 and CV) allocated up to 40% of the surplus growth to reserve build-up, while grazing
563 until the very end of the productive season (Fig. 6F, G).

564 It appears that species-CG and -CH are adopting a more conservative strategy that
565 prepares themselves for an upcoming overwintering period, than pushing themselves
566 toward attaining sexual maturity. We interpret this conservative life strategy as a classic
567 adaptation to seasonality in the Arctic pelagic environments ([Conover & Siferd 1993](#),
568 [Hagen & Auel 2001](#)). The decision to store more and overwinter possesses a significant
569 pay-off in the following year, where surplus reserves can be allocated to egg production.
570 Species-CG and -CH ascended in mid-February, with the commencement of pelagic
571 primary production (Fig. 6F, G). As the food-availability until mid-April (peak bloom) was
572 relatively low, species-CG used the post-overwintering surplus energy reserves as a
573 capital for egg production (Fig. 6B, C, J, K), which agrees well with the egg production
574 strategy described for *C. glacialis* ([Swift 1986](#), [Melle & Skjoldal 1998](#), [Niehoff et al. 2002](#),

575 [Søreide et al. 2010](#)). The profitability of the mixed income and capital breeding strategy
576 of species-CG was such that its total fecundity was higher compared to the other model-
577 species (Fig. 6M–O). The egg production of species-CH ceased ca. 10 d before the peak
578 pelagic bloom, as the reserves were spent on producing nearly 1700 eggs (equivalent to
579 capital input of ca. 950 μgC , Fig. 6K, O). This agrees with the capital breeding strategy
580 described for *C. hyperboreus* ([Dawson 1978](#), [Matthews et al. 1978](#), [Smith 1990](#), [Hirche](#)
581 [1997](#), [Scott et al. 2000](#), [Niehoff et al. 2002](#), [Hirche & Kosobokova 2003](#)). After spawning,
582 the species-CH females lived toward the end of the productive season, without serving
583 any adaptive benefit (Fig. 6G, K, O). This hints at a possibility that if not constrained by
584 our model, these females could have acquired energy reserves and probably spawned
585 again in the following year. Such iteroparous breeding has been commonly reported for
586 *C. hyperboreus* ([Hirche & Kwasniewski 1997](#), [Arnkvaern et al. 2005](#), [Hirche 2013](#)).

587

588 3.2 Sensitivity of emergent strategies to environmental variability

589 3.2.1 Food concentration

590 At a 25% higher food concentration ($F = 225 \mu\text{g C l}^{-1} \approx 7.5 \text{ mg Chl.-a m}^{-3}$), the predicted
591 optimal birth time of species-CF occurred ca. 2 days later compared to *BR* on June 18
592 (Fig. 7A, Table 6). It seems that species-CF used the higher food concentrations and
593 temperatures later in the year to speed-up structural growth, but allocated less for
594 energy reserves and entered diapause with 3% less energy reserves compared to *BR*
595 ($W_s/W_c = 0.52$). The timing of seasonal descent did not change much from the *BR*, mainly
596 because of late birth time, slightly elevated DVM (which tend to reduce growth
597 potential) and ca. 2% larger size of the overwintering CVs (Table 6). However, higher
598 food concentration influenced the income breeding egg production of species-CF, which
599 was elevated by ca. 15% compared to the *BR*.

600 The influence of elevated food concentration on the species-CG and -CH was notably
601 different from that of species-CF. The predicted birth times of species-CG and -CH
602 occurred ca. 2 days earlier (Fig. 7B, C, Table 6), and the growth allocation parameter was
603 ca. 11% higher at 25% higher food concentration compared to the *BR*. Thus, it seems
604 that these two model-species used the higher growth potential resulted from elevated
605 food concentration to develop energy reserves rather than structural growth. This was
606 achieved by two means: firstly, through rapid accumulation of reserves, shown by the
607 2–4 days earlier timing of seasonal descent, and secondly, by elevating the size of the
608 overwintering CVs by ca. 2% compared to the *BR* (Table 6), which allows a little extra
609 space for energy storage at the maximum W_s/W_c ratio of 0.7 (cf. Fig. 5C). Consequently,
610 the number of capital breeding eggs increased for both model-species, atop of which,

611 species-CG produced more income breeding eggs, using the higher food concentrations
612 (Fig. 7B, and Table 6).

613 These findings point to how a small improvement in growth potential was
614 differentially utilized by the three model-species. This was largely driven by their
615 breeding modes. Since there is no added advantage of carrying extra energy reserves
616 for the purely income breeding species-CF, it directed the higher growth potential
617 towards accelerating structural growth, which likely reflect a short-term motivation
618 toward attaining sexual maturity within the same calendar year ([Jönsson 1997](#),
619 [Sainmont et al. 2014](#), [Barta 2016](#)). However, since the duration of the productive season
620 remained the same as *BR*, SVM and diapause yet appeared to be inevitable for this
621 species (Fig. 1A–C). On the other hand, the capital breeding strategy of species-CG and
622 -CH made them use the elevated growth potential as a long-term investment ([Varpe et
623 al. 2009](#), [Eismond et al. 2015](#)) that increased fecundity in the following year (Table 6).

624 At a 25% lower food concentration ($F = 135 \mu\text{g C l}^{-1} = 4.5 \text{ mg Chl.-a m}^{-3}$), the differences
625 between the two emergent strategies of the *BR* appeared to diminish. Here, the
626 predicted optimal birth times of all species occurred ca. 12–15 d earlier in the year (Fig.
627 7A–C, Table 6). Due to lowered growth potential, they occupied more time in food-rich
628 near-surface waters with reduced DVM, and grazed later into the productive season
629 before descending to diapause in late-August. Further, ca. 40% of the surplus growth
630 was allocated to building up of energy reserves. Irrespective of the breeding mode,
631 fecundity of all model-species decreased by 5–11% (Table 6). The attempt to elevate
632 energy reserves by the purely income breeding species-CF indicates a shift towards a
633 more conservative strategy. Storing additional reserves compared to *BR* reflects the
634 difference in the structural masses of the overwintering CV and the adult, which was
635 spanned by allocating the reserves to structural growth in the following year. This
636 plasticity of life strategy of species-CF reflects that of *C. finmarchicus* in Arctic locations,
637 which can maintain viable populations in cold and food-limited environments (e.g.
638 [Hirche & Kwasniewski 1997](#), [Madsen et al. 2001](#), [Arnkværn et al. 2005](#), [Hirche &
639 Kosobokova 2007](#), [Bandara et al. 2016](#)).

640

641 3.2.2 Temperature

642 Temperature had the highest influence on the strategies emerging from the *BR* (Fig.
643 7D–F). A 25% temperature increase throughout the water column (i.e. $T_{max} = 18.75 \text{ }^\circ\text{C}$,
644 $T_{min} = 2.5 \text{ }^\circ\text{C}$) caused the birth times of all model-species to delay by 2–35 days (Table 6).
645 This delay became more pronounced in species-CG and -CH, which reflects their
646 motivation to utilize the higher growth potential by allowing their younger
647 developmental stages to thrive in warmer, food-rich waters of the late-spring and

648 summer. Here, the elevated visual predation risk was countered by performing
649 pronounced DVM (Table 6). In all model-species, the predicted development times and
650 sizes of the overwintering stages and females decreased (Table 6). Instead of prioritizing
651 structural growth, all model-species allocated more of the surplus growth to building up
652 of energy reserves and entered diapause with nearly full lipid reserves ($W_s/W_c > 0.69$).
653 This seems to be a driven by the elevated temperature at overwintering depths, which
654 tends to exhaust energy reserves faster than those used in the *BR* (2.5 °C vs 2.0 °C, cf.
655 Eq. 15). Consequently, the number of capital breeding eggs decreased in species-CG and
656 -CH. However, due to the elevated assimilation efficiency at higher temperatures (Eq.
657 3), income breeding appeared more profitable for species-CF and -CG, which is shown
658 by 15%–20% increase of fecundity (Table 6). This suggests that in a warmer ocean, *C.*
659 *finmarchicus* would use the elevated fecundity as an advantage to increase their
660 populations in the high-Arctic (see also, [Beaugrand et al. 2002](#), [Chust et al. 2014](#)). Under
661 similar circumstances, the purely income breeding strategy of *C. hyperboreus* could
662 become disadvantageous, as energy requirements at diapause elevates, and leaves less
663 reserves for egg production ([Hirche 1991, 1997](#), [Maps et al. 2013](#)). However, the flexible
664 breeding strategy of *C. glacialis* will enable it to cope with either warming or cooling
665 scenarios, as loss of fecundity through decreased capital breeding would be
666 compensated from income breeding and vice versa ([Falk-Petersen et al. 2007](#), [Daase et](#)
667 [al. 2013](#), [Grote et al. 2015](#)).

668 At lower temperatures (i.e. $T_{max} = 11.25$ °C, $T_{min} = 1.5$ °C), all species were born 2–27
669 days earlier in the year and occupied more time in near-surface waters with reduced
670 DVM (Fig. 7, Table 6). This reflects the longer time it takes to develop to diapause stage
671 due to lower growth potential attained in colder waters (Eqs. 1–3). Further, grazing
672 continued toward the very end of the productive season (late August in most cases), and
673 the overwintering CVs of all model-species were ca. 2–4% larger than predicted in the
674 *BR* (Table 6). At the time of seasonal descent, CVs of species-CF had partly filled energy
675 reserves ($W_s/W_c \approx 0.51$). This reflects the decreased diapause metabolic costs at lower
676 temperatures (1.5 °C vs 2.0 °C, cf. Eq. 15). In contrast, species-CG and -CH entered
677 diapause with full energy reserves ($W_s/W_c \approx 0.70$), and used the post-overwintering
678 surplus reserves to elevate capital breeding (Table 6). However, only the purely income
679 breeding species-CH had a ca. 10% fecundity gain, while species-CF and -CG suffered a
680 12%–15% loss of income breeding potential (and hence the total fecundity, Table 6), due
681 to decreased assimilation efficiency rendered at lower temperatures (Eq. 3).

682

683 3.2.3 Predation risk

684 Compared to food concentration and temperature, a 25% change in non-visual and
685 visual predation risks had a negligible influence on the emergent behavioral and life
686 history strategies of the *BR* (Fig. 7G–L). low sensitivity to non-visual predation risk has
687 also been observed by ([Bandara et al. 2018](#)). The low sensitivity to visual predation risk
688 may be due to that it operates on a larger range than manifested in the sensitivity
689 analysis (see below).

690

691 3.3 Emergent strategies under bottom-up and top-down selection pressures

692 3.3.1 Emergent strategies at low visual predation risk

693 At the lowest level of visual predation risk ($K' = 0.032$), the species-specific behavioral
694 and life history strategies emerging from the model appear to be heavily influenced by
695 the patterns of food availability and temperature. Here, in each model environment, the
696 predicted optimal birth times of all model-species occurred the earliest (Fig. 8C1–C9).
697 The DVM was less pronounced (Fig. 9A1–A9), and hence they suffered least from food-
698 limitation (Fig. 9B1–B9). All model-species developed relatively slowly due to lower
699 temperatures that occurred earlier in the season (Fig. 1B, E and H), and developed to CV
700 stage with the highest structural and energy reserve masses possible (Fig. 8B1–B9).
701 These large CVs descended to overwintering depths earlier in the year (Fig. 8A1–A8) for
702 diapause. This was followed by an earlier seasonal ascent and reproduction (Fig. 8C1–
703 C8). At the onset of reproduction, females were older compared to higher predation risk
704 levels (Fig. 8D1–D8) and their body masses were the highest (Fig. 8E1–E8). The larger
705 females of species-CF and -CG could assimilate more efficiently (Eqs. 1 and 2), and
706 produced the highest number of eggs (Fig. 8F1–F6). The extensive energy reserves of
707 larger overwintering CVs of species-CG and -CH were used for higher capital breeding
708 output (Fig. 8F4–F8).

709 At the lowest level of visual predation risk, environment and species-specific patterns
710 were also apparent. In species-CF, the birth times shifted ca. 45 days later into the year
711 from mid-April to early-June along the modelled latitudinal gradient (Fig. 8C1–C3). These
712 match the exact times at which the peak pelagic bloom occurred in these model
713 environments (Fig. 1C, F and I). Because of the decreasing gradient of growth potential
714 encountered along the modelled latitudinal gradient and the ca. 20% increase of the size
715 of the overwintering CVs (Fig. 8B1–B3), the timing of seasonal descent shifted by ca. 78
716 days from late-June to early-September (Fig. 8A1–A3). The timing of seasonal ascent and
717 onset of reproduction also followed an increasing trend along the modelled latitudinal
718 gradient, and aligned with the time of the onset of pelagic algal bloom (Figs. 1C, F and I,
719 and 8C1–C3). Because of delayed birth times and elevated development times

720 associated with lower temperatures, the age of sexual maturity increased along the
721 modelled latitudinal gradient (Fig. 8D1–D3). Further, the size at sexual maturity also
722 increased by ca. 20% (Fig. 8E1–E3). However, the increased size of female could not
723 compensate for the decreased assimilation rates induced by lower temperature at
724 higher latitude model environments, as the fecundity decreased by ca. 30% (Fig. 8F1–
725 F3). These findings suggest that species-CF timed its reproduction to match the timing
726 of the pelagic algal bloom along the modeled latitudes. This has been a common
727 observation for *C. finmarchicus*, and reflects the strong dependency of its reproduction
728 on the food availability ([Tande & Hopkins 1981](#), [Aksnes & Magnesen 1983](#), [Hirche &
729 Kosobokova 2003](#), [Madsen et al. 2008](#)).

730 Unlike species-CF, the birth times of species-CG and -CH did not change much along
731 the modelled latitudinal gradient (Fig. 8C4–C9). At the lower latitude Environment-L,
732 predicted birth times of species-CG and -CH roughly aligned with the timing of the peak
733 pelagic algal bloom (Figs. 1C, 8C4 and C7). At higher latitude environments, their birth
734 times were predicted ca. 7–25 days ahead of the peak algal bloom (Figs. 1F, I, 8C5, C6,
735 C8 and C9). These temporal offsets roughly align with the cumulative development times
736 estimated from eggs to first feeding NIII stage estimated from Bělehrádek temperature
737 functions (Eq. 6, Table 5), and thus reflect the classical descriptions of capital breeding
738 strategies of Arctic *Calanus* species (reviewed in, [Falk-Petersen et al. 2009](#), [Søreide et al.
739 2010](#), [Daase et al. 2013](#)), where egg production is timed so that first feeding stages can
740 feed under non-limiting food concentrations.

741 In species-CG, the size of the overwintering CVs increased by ca. 25% along the
742 modelled latitudinal gradient (Fig. 8B4–B6). Given the lower temperatures modelled at
743 higher latitude environments (Fig. 1B, E and H) these larger CVs had to graze toward the
744 end of the productive season to gain energy reserves required for overwintering (Fig.
745 8A4–A6). The timing of seasonal ascent and reproduction of species-CG showed a delay
746 of ca. 60 days along the modelled latitudinal gradient (Fig. 8A4–A6 and C4–C6), and
747 reflects the coupling of its reproduction with the timing of the pelagic bloom (Fig. 1C, F
748 and I) despite being a partly a capital breeder (Fig. 8G4–G6). Although capital breeding
749 of species-CG accounted for an increasing fraction of the total egg production along the
750 modelled latitudinal gradient (7%–11%), the total fecundity decreased by ca. 21% due
751 to the reduced income breeding potential ensued by lower growth potential sustained
752 in higher latitude model environments (Fig. 8F4–F6).

753 The variability in the timing of SVM and reproduction was least apparent for species-
754 CH. Its timing of seasonal descent showed the least variability along the modelled
755 latitudinal gradient (i.e. ca. 30 days from late-August to late-September, Fig. 8A7–A9),
756 while the timing of seasonal ascent and the onset of reproduction did not change (Fig.
757 8C7–C9). This reflects the decoupling of species-CH's reproduction from the timing of

758 pelagic bloom, which was driven by the purely capital breeding strategy. The early-
759 January seasonal ascent and the onset of reproduction predicted by our model for the
760 highest latitude model environment may be doubtful since it occurs in the midst of the
761 Arctic polar night with the pelagic algal bloom ca. 5–6 months away (Fig. 1G, I). However,
762 this aligns with recent field observations on the timing of seasonal ascent and
763 reproduction in several high-Arctic fjords in the Svalbard archipelago between 78°–80°
764 N ([Daase et al. 2014](#), [Błachowiak-Samołyk et al. 2015](#), [Bandara et al. 2016](#)). Given the
765 strict herbivory of the modelled copepods, the viability of these early eggs remains
766 questionable, as NIII emerges within ca. 15 days (at -1.5 °C) after being spawned, which
767 precedes the pelagic bloom by several months. However, in nature, these early born
768 copepods may survive, as they can feed on alternative food, such as microzooplankton
769 and ice algae ([Runge & Ingram 1991](#), [Søreide et al. 2008](#), [Campbell et al. 2016](#)).

770 Although we assumed $K' = 0.032$ as a reference value for extremely low visual
771 predation risk, it had a notable influence on species-CH at the Environment-H. Here,
772 unlike other model-species, the species-CH did not follow the maximum potential stage-
773 specific body mass trajectory (W_j^{max} , Fig. 4G–I). Consequently, the size of the
774 overwintering CV and the size at sexual maturity predicted for species-CH was ca. 11%
775 smaller compared to that predicted for Environment-M (Fig. 8B8, B9, E8 and E9). As the
776 size of the energy reserve was modelled as a fixed (yet, evolvable) fraction of the
777 structural mass, the decreased structural mass translated to a ca. 15% decrease of
778 fecundity (Fig. 8F8 and F9). This finding points to the fact that the adaptive advantage
779 of a larger body size at higher latitudes can be highly sensitive to top-down
780 environmental selection pressures (see below).

781

782 3.3.2 Emergent strategies at elevated visual predation risk

783 As the visual predation risk increased from its baseline level of $K' = 0.032$ toward 0.32,
784 the bottom-up influences described above diminished, and all model-species reacted to
785 visual predation risk in more or less the same manner. Altogether, two behavioral and
786 life history strategies were manifested to counter the elevated visual predation risk.

787

788 3.3.1.1 Plasticity of behavior: diel vertical migration

789 DVM was used to counter relatively modest levels of visual predation risk, i.e. $0.032 \leq$
790 $K' \leq 0.15$. Here, feeding stages (NII and onwards) of all model-species reduced the time
791 spent in warmer, food-rich near-surface waters by descending to depths typically
792 exceeding 100 m (Fig. 9A1–A9 and B1–B9). As reduced surface time decreases feeding
793 opportunities, diel migrants suffered from increased food limitation (Fig. 9C1–C9), which
794 led to reduced growth rates that ultimately elevated development times (Fig. 9D1–D9).

795 To compensate for the DVM-induced loss of growth potential, birth times of all model-
796 species were shifted later into the year (Fig. 8C1–C6), possibly to utilize the higher
797 temperatures that occur later in the season (Fig. 1B, E and H) and attain higher growth.
798 Consequently, they had to feed later into the productive season to fulfil the energy
799 requirements needed to survive the forthcoming unproductive winter, and descended
800 to overwintering depths in late-autumn (Fig. 8A1–A9). This late-birth strategy was such
801 successful that, despite the elevated DVM, fecundity of all model species remained
802 largely unchanged (Fig. 8F1–F9). An exception to the above-mentioned phenological
803 shifts was observed for species-CH at the Environment-H. Here, timing of birth, SVM,
804 reproduction did not change for the initial increase of visual predation risk despite its
805 larger body size (Fig. 8A9, B9 and C9). At Environment-H, younger developmental stages
806 (up to NVI) of species-CH did not perform notable DVM (Fig. 9A9, B). This was caused by
807 the early seasonal ascent and reproduction of this species, which allowed its younger
808 developmental stages to elevate foraging efforts in near-surface waters in a period with
809 minimal irradiance, that reduces the risks imposed by visually orientating planktivores
810 (Fig. 1G).

811 Although DVM is the most immediate response against elevated visual predation
812 (reviewed in, [Lampert 1989](#), [Hays 2003](#), [Brierley 2014](#)), the associated phenological
813 shifts were brought into prominence in the recent work of [Bandara et al. \(2018\)](#). Findings
814 of the above study and those of this investigation agrees well, but do not align with the
815 belief that increased visual predation risk drives earlier seasonal descents in *Calanus*
816 spp. ([Kaartvedt 2000](#), [Varpe & Fiksen 2010](#)).

817

818 3.3.1.2 Plasticity of body size

819 As the visual predation risk increased further ($0.15 \leq K' \leq 0.22$), the trading-off of
820 growth potential for survival became unprofitable. This was caused by the inability to
821 further delay the birth times (Fig. 8C1–C9) in response to elevated DVM, as the growth
822 and development of later developmental stages became constrained by the duration of
823 the productive season. At this point, model-species began to adjust the size (i.e.
824 structural and energy reserve masses, W_c and W_s) of overwintering stages. Here, instead
825 of overwintering as larger CV stages with full energy reserves, all species entered
826 diapause as CIV and CIII stages with 50%–90% lesser structural and energy reserve
827 masses. This strategy did not notably reduce the DVM nor the food limitation effects
828 ensued (Fig. 9), but allowed the model-species to reduce visual predator-induced
829 mortality risk during the ca. 200–350 days long diapause, as in this model, visual
830 predation risk was not nullified even at greater depths (Eqs. 9 and 10). Similar
831 predictions have been made previous models (e.g. [Fiksen & Carlotti 1998](#), [Varpe et al.](#)
832 [2007](#), [Bandara et al. 2018](#)). The timing of reproduction of all species were significantly

833 delayed by this strategy, as the smaller overwintering stages must use the post-
834 overwintering residual energy reserves or gains from food intake to elevate their
835 structural mass to attain sexual maturity. This reduced the capital breeding potential of
836 species-CG and -CH. Further, at $K' \geq 0.18$, species-CG could not produce any capital
837 breeding eggs, and switched its reproductive strategy to pure income breeding (Fig.
838 8G4–G6).

839 Further increase of visual predation risk, i.e. $0.22 < K' \leq 0.32$ was dealt with the
840 reduction of body masses at each developmental stage. In the model, this was achieved
841 by evolving smaller values for the body size parameter α (Table 2). Modelled copepods
842 with smaller body masses could reduce the vulnerability of younger developmental
843 stages to visual predation, and hence the DVM was restored back to the levels observed
844 at lowest level of visual predation risk (Fig. 9A1–A9, and B1–B9). As the copepods could
845 occupy more time on the food-rich near-surface waters, the food limitation of younger
846 developmental stages also decreased (Fig. 9C1–C9). However, birth times did not shift
847 back in time to occur earlier in the year, probably reflecting the need for the smaller size
848 copepods to occupy warmer waters to elevate their assimilation efficiency (Eqs. 1–3).
849 Consequently, adult females attained sexual maturity at smaller body sizes, and the
850 expected fecundity decreased dramatically by 20%–60% among income breeding
851 species-CF and -CG (Fig. 8F1–F6). As the capacity to carry energy reserves decreases with
852 body size (Fig. 5C), the capital breeding capacity of species-CH was severely hampered,
853 and its fecundity decreased by ca. 40% at the lower latitude Environment-L, and ca. 96%
854 at the higher latitude Environment-H (Fig. 8F7–F9).

855

856 3.4 Concluding remarks

857 The artificial evolution of body sizes observed in this study resembles the classic field
858 observations of rapid evolution of zooplankton body sizes in response to size-selective
859 predation by planktivorous fish in smaller freshwater lakes (e.g. [Brooks & Dodson 1965](#),
860 [Wells 1970](#), [Zaret & Kerfoot 1975](#)). Although such observations are rare in the Arctic
861 marine realm, recent hypotheses suggest that the evolution of differential body size
862 pattern among *Calanus* spp. is largely due to extensive size-selective predation induced
863 by mega planktivores ([Berge et al. 2012](#), [Falk-Petersen et al. 2015](#)). A key difference
864 between the above studies and our model is that body size plasticity evolves as the ‘last
865 resort’, when the increasing visual predation risk could not be countered with behavioral
866 strategies. However, if visual predation risk elevates beyond the limits of behavioral
867 toleration, it can have dramatic consequences on zooplankton life strategies which can
868 easily outweigh those induced in the bottom-up (Figs. 8 and 9).

869 Further, elevated visual predation risk obscured the apparent south to north trends
870 in ontogenetic body size patterns observed under the minimum visual predation risk
871 (Fig. 8E1–E9, cf. Fig. 4), leading to increased overlap of body size ranges (especially
872 between species-CF and -CG) irrespective of the modelled latitude. Therefore, top-down
873 selection pressures, such as the presence of resident or seasonally migrating
874 populations of planktivorous fish (e.g. [Varpe et al. 2005](#), [Renaud et al. 2012](#)) should be
875 considered as an important factor when assessing the potential for misidentifying
876 coexisting *C. finmarchicus* and *C. glacialis* populations using morphometric methods
877 (e.g. [Parent et al. 2011](#), [Gabrielsen et al. 2012](#)).

878 At all scenarios tested, the annual life cycle was the only generation time emerged in
879 this model. Upon further testing we found that > 1 year generation times do emerge
880 when the duration of pelagic productive season is cut down by ca. 40% under lower
881 levels of visual predation risk modelled here ($K < 5 \times 10^{-3}$). Therefore, it is likely that the
882 influences of bottom-up selection pressures become more apparent in higher-latitude
883 seasonal environments where resource limitation and year to year stochastic
884 environmental variability is more pronounced ([Roff 1980](#), [Fiksen 2000](#), [Ji 2011](#)).

885 Bottom-up and top-down environmental variability acts as selection pressures that
886 operates interactively to drive seasonal adaptations of high-latitude pelagic inhabitants
887 ([Varpe 2017](#)). There is often a discrepancy in views about which selective force holds
888 the primacy ([Hunter & Price 1992](#), [Power 1992](#), [Baum & Worm 2009](#)). Using a model
889 that allows partitioning the two selection pressures and artificial evolution of seasonal
890 strategies, we argue that top-down selective forces are more significant in shaping up
891 of behavioral and life history strategies of Arctic *Calanus* species. However, the
892 influences of alternative food sources, sea-ice, and spatio-temporal heterogeneity of the
893 environment should be considered toward drawing stronger conclusions.

894

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900 **Table 1** The inter-and intra-specific diversity of some life history traits/attributes of
 901 *Calanus* spp. Body mass estimates are from prosome length (PL) to dry mass (DM)
 902 relationships published by [Robertson \(1968\)](#). Cited literature only serve as examples.
 903 See [Falk-Petersen et al. \(2009\)](#), and [Bandara \(2014\)](#) for an extensive review on some of
 904 these life history traits.

Trait/attribute	<i>C. finmarchicus</i>	<i>C. glacialis</i>	<i>C. hyperboreus</i>
Center of distribution	North Atlantic ^[5,11]	Arctic (shelf) ^[5,11]	Arctic (oceanic) ^[11]
Body size			
Length (mm PL)	2.2 ^[4] –3.2 ^[14,30,17]	3.2 ^[14,17,25] –4.6 ^[17,28]	3.9 ^[17] –6.7 ^[28]
Mass (µg DM)	204–557 ^[3]	533–1742 ^[3]	1016–5947 ^[3]
Timing of reproduction	In synchrony with pelagic bloom ^[8,16,31]	Before or in synchrony with pelagic bloom ^[12,22,26,32]	Before the pelagic bloom ^[6,7,12,26]
Reproductive strategy	Income breeding ^[18]	Income or capital breeding ^[32]	Capital breeding ^[2,19]
Most common overwintering stages	CIV ^[8,31,33] CV ^[15,22,28]	CIV ^[23,17,29] CV ^[13,21,24]	CIII ^[13,27] CIV ^[6,7,29] CV ^[20,25,28] Females ^[13,21,29]
Most common generation times (years)	1 ^[9,21,29]	1–2 ^[10,23,24,29]	1–3 ^[1,13,20,23,25]

[\[1\] Conover \(1965\)](#), [\[2\] Conover \(1967\)](#), [\[3\] Robertson \(1968\)](#), [\[4\] Jaschnov \(1972\)](#), [\[5\] Fleminger and Hulsemann \(1977\)](#), [\[6\] Dawson \(1978\)](#), [\[7\] Matthews et al. \(1978\)](#), [\[8\] Tande and Hopkins \(1981\)](#), [\[9\] Aksnes and Magnesen \(1983\)](#), [\[10\] Tande et al. \(1985\)](#), [\[11\] Conover \(1988\)](#), [\[12\] Smith \(1990\)](#), [\[13\] Hirche \(1991\)](#), [\[14\] Unstad and Tande \(1991\)](#), [\[15\] Diel and Tande \(1992\)](#), [\[16\] Plourde and Runge \(1993\)](#), [\[17\] Hirche et al. \(1994\)](#), [\[18\] Hirche \(1996b\)](#), [\[19\] Hirche and Niehoff \(1996\)](#), [\[20\] Hirche \(1997\)](#), [\[21\] Hirche and Kwasniewski \(1997\)](#), [\[22\] Melle and Skjoldal \(1998\)](#), [\[23\] Falk-Petersen et al. \(1999\)](#), [\[24\] Scott et al. \(2000\)](#), [\[25\] Madsen et al. \(2001\)](#), [\[26\] Niehoff et al. \(2002\)](#), [\[27\] Astthorsson and Gislason \(2003\)](#), [\[28\] Hirche and Kosobokova \(2003\)](#), [\[29\] Arnkværn et al. \(2005\)](#), [\[30\] Daase and Eiane \(2007\)](#), [\[31\] Madsen et al. \(2008\)](#), [\[32\] Søreide et al. \(2010\)](#), [\[33\] Hirche and Kosobokova \(2011\)](#)

905

906 **Table 2** List of evolvable (soft) parameters optimized by the Genetic Algorithm. The
 907 parameters β , γ , δ and ε are proxies that define the behavioral strategy (vertical
 908 strategy), while the rest are proxies that define some key aspects of the life history
 909 strategy. These proxies (parameters 1–6) of behavioral and life history strategies are
 910 hardwired to copepods spawned in different times of the year (t_B).

Term	Definition	Range	Interval	Unit
α	Body size parameter	0–1	0.01	dim. less
β	Irradiance threshold parameter	10^{-7} –1	*	$\mu\text{mol m}^{-2} \text{s}^{-1}$
γ	Energy allocation parameter	0–1	0.01	dim. less
δ	Seasonal descent parameter	0–1	0.01	dim. less
ε	Seasonal ascent parameter	0–1	0.01	dim. less
η	Generation time parameter	1–3	1	years
t_B	Birth time	1–8760	1	h

*Intervals = 10^{-7} , 10^{-6} , 10^{-5} , 10^{-4} , 10^{-3} , 10^{-2} , 10^{-1} , 1

911

912 **Table 3** Description of the emergent behavioral and life history strategies of the model.
 913 Note that unlike [Bandara et al. \(2018\)](#), the amplitude of the SVM (overwintering depth)
 914 is not an emergent property of this model.

Trait/attribute	Units	Description
Time of birth	Day of the year	Time of being spawned
Surface time	h	Unified estimate representing the timing of DVM, i.e. the stage-specific mean no. of hours per day occupied in waters with highest growth potential (usually the surface waters)
Amplitude of diel vertical migration	m	The vertical range corresponding to the above
Time of seasonal descent and ascent	Day of the year	Separate estimates representing the timing of SVM (ascent and descent)
Size at seasonal descent	$\mu\text{g C}$	Structural and energetic reserve mass at the onset of diapause
Age of sexual maturity	d	Time since birth to the first egg production
Size at sexual maturity	μgC	Structural mass at first egg production
Onset of spawning	Day of the year	Time of first egg production
Fecundity	No. of eggs	No. of eggs produced during the lifetime
Breeding mode index	dim.less	Proportion of capital breeding eggs (0 = pure income breeding, 1 = pure capital breeding)
Food limitation index	dim.less	Stage-specific total no. hours with food-limited growth (Eq. 3) as a fraction of stage duration (0 = no food limitation, 1 = total food limitation)
Development time	d	From egg to a given stage
Longevity	d	Duration of the life cycle, from birth to death

915

916 **Table 4** Definitions, values and units of the non-evolvable (hard) parameters used in the
 917 model. See Table 2 for a description of evolvable parameters

Term	Definition	Value/Reference	Units
a	Assimilation coefficient	0.6 ^[3]	–
$A_{i,s,t,z}$	Ingestion rate	Eqs. 2–4	$\mu\text{g C ind.}^{-1} \text{h}^{-1}$
b_s	mass coefficient of ingestion	Eq. 2, Table 5	–
$B_{i,s,t,z}$	Metabolic rate	Eqs. 14–15	$\mu\text{g C ind.}^{-1} \text{h}^{-1}$
$(B_a)_{i,s,t,z}$	Active metabolic rate	$1.5 \cdot B_b$	$\mu\text{g C ind.}^{-1} \text{h}^{-1}$
$(B_b)_{i,s,t,z}$	Basal metabolic rate	Eqs. 6–7	$\mu\text{g C ind.}^{-1} \text{h}^{-1}$
c_s	Temperature coefficient of ingestion	Eq. 3, Table 5	–
$d_{i,t}$	Parameter for satiation food concentration	0.1–0.3	–
$D_{s,t,z}$	Development time	Eq. 6	h
$F_{t,z}$	Ambient food concentration	Fig. 1	$\mu\text{g C l}^{-1}$
f_s	mass coefficient of metabolism	Table 5	–
$G_{i,s,t,z}$	Growth rate	Eq. 1	$\mu\text{g C ind.}^{-1} \text{h}^{-1}$
g_s	temperature coefficient of metabolism	Table 5	–
$H_{i,t,z}$	Survivorship	Eq. 19	–
i	Individual	–	–
$I_{t,0}$	Irradiance incident on sea surface	Fig. 1	$\mu\text{mol m}^{-2} \text{s}^{-1}$
$I_{t,z}$	Irradiance at depth z	Eq. 9	$\mu\text{mol m}^{-2} \text{s}^{-1}$
$I'_{t,z}$	Remapped $I_{t,z}$	0.1–0.9	–
j	Developmental stage	0–12	Egg–Adult
$K_{i,t}$	Scalar for visual predation risk	10^{-6} – 10^{-2}	–
$L_{i,t}$	Irradiance sensitivity parameter	1–2.5	–
m_s	Mass exponent of ingestion	Eq. 2, Table 5	–
$(M_n)_{i,t,z}$	Non-visual predation risk	$0.1 \cdot K$	–
$(M_s)_{i,t}$	Starvation risk	Eq. 16	–
$(M_v)_{i,t,z}$	Visual predation risk	Eq. 10	–
n_s	Temperature exponent of ingestion	Eq. 3, Table 5	–
N	No. of initial seeds	2.5×10^6	strategies
o_s	mass exponent of metabolism	Table 5	–
p_s	temperature exponent of metabolism	Table 5	–
q_s	Development time parameter-1	Table 5 ^[1, 4]	–
r_s	Development time parameter-2	Table 5 ^[1, 4]	–
$R_{i,s,t,z}$	Fecundity	Eq. 17	No. of eggs
s	Species	CF, CG, CH	–
$T_{t,z}$	Ambient temperature	Fig. 1	$^{\circ}\text{C}$
t	Time	1–8760	h
t_R	Time of sexual maturity	1–8760	hour
t_X	Time horizon	1–8760	hour
$U_{i,t}$	Cruising velocity	Eq. 11	m h^{-1}
W_c	Structural body mass	–	$\mu\text{g C}$

continued...

continued...

$(W_E)_s$	Species-specific unit egg mass	Table 5	$\mu\text{g C}$
W_j	Stage-specific critical molting mass	Fig. 3	$\mu\text{g C}$
$(W_j^{max})_{i,s}$	Stage-specific minimum W_j	Fig. 3	$\mu\text{g C}$
$(W_j^{min})_{i,s}$	Stage-specific minimum W_j	Fig. 3	$\mu\text{g C}$
$(W_R)_{i,s,t,z}$	Matter allocated for egg production	Eq. 17	$\mu\text{g C}$
W_s	Mass of the energy reserve	–	$\mu\text{g C}$
	Catabolized structural mass		
$(W_x)_{i,s}$	(proportion of the maximum lifetime structural mass)	0.1–0.5	–
z	Depth	1–1000	m
φ	Termination condition of GA	Section 2.6.3	–
ψ	Water column light attenuation coefficient	$0.06^{[2]}$	m^{-1}
$\Omega_{i,s}$	Fitness	Eq. 18	–
ω	Weight for fitness	0 or 1 ^[5]	–

[\[1\] Campbell et al. \(2001\)](#), [\[2\] Eiane and Parisi \(2001\)](#), [\[3\] Maps et al. \(2011\)](#), [\[4\] Ji et al. \(2012\)](#), [\[5\] Bandara et al. \(2018\)](#)

919 **Table 5** Species-specific values parameter values for coefficients and exponents of
 920 ingestion and respiration ([estimated from Maps et al. 2011, Appendix A2](#)), along with
 921 species-specific egg masses and development time parameters. See Table 4 for term
 922 definitions.

Parameter	Species-CF	Species-CG	Species-CH
b_s	0.009283	0.01656	0.01319
m_s	0.7524	0.7518	0.7516
c_s	1.2382	1.1606	1.1833
n_s	0.0966	0.0673	0.0761
f_s	0.0008487	0.003292	0.001153
o_s	0.75	0.75	0.75
g_s	1.2956	1.1382	1.2065
p_s	0.1170	0.0585	0.0849
$(W_E)_s$	0.23 $\mu\text{g C}^{[2]}$	0.40 $\mu\text{g C}^{[3]}$	0.56 $\mu\text{g C}^{[4]}$
q_s (eggs)	595 ^[5]	839 ^[6]	1495 ^[6]
q_s (NI)	388 ^[5]	548 ^[6]	974 ^[6]
q_s (NIII)	581 ^[5]	819 ^[6]	1461 ^[6]
r_s (eggs, NI, and NII)	9.11 ^[1]	13.04 ^[1]	13.66 ^[1]

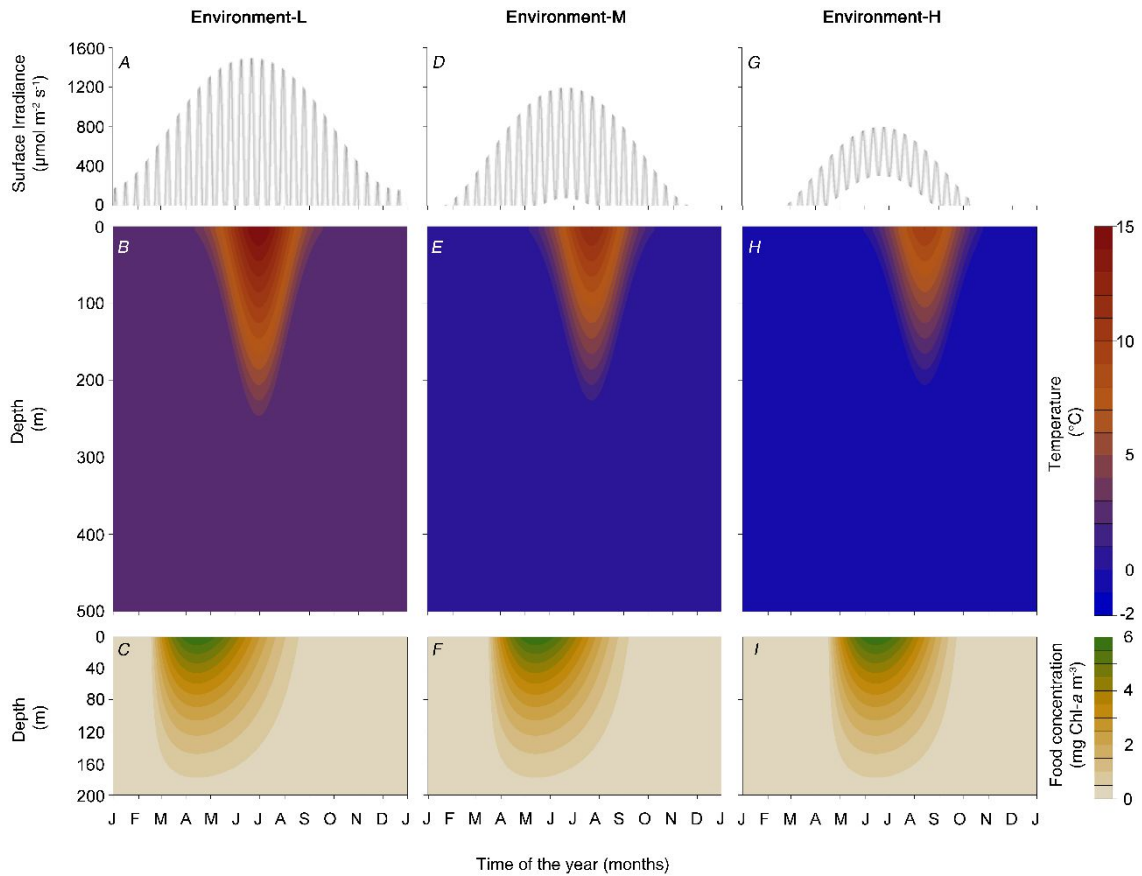
923 [\[1\] Corkett et al. \(1986\)](#), [\[2\] Hirche and Bohrer \(1987\)](#), [\[3\] Hirche \(1990\)](#), [\[4\] Smith \(1990\)](#),
 924 [\[5\] Campbell et al. \(2001\)](#), [\[6\] Ji et al. \(2012\)](#)

925

926 **Table 6** Influence of $\pm 25\%$ changes in food concentration (F) and temperature (T) changes on the behavioral and life history strategies
 927 emergent from the basic run (BR) for the three model-species. Variations of food concentration ranges between $F-$: $135 \mu\text{g C l}^{-1}$, $F+$: 225
 928 $\mu\text{g C l}^{-1}$, and temperature between $T-$: 11.25°C max and 1.5°C min, and $T+$: 18.75°C max and 2.5°C min. Influences of predation risk are
 929 not tabulated due to their lesser significance (cf. Fig. 7).

Species	Scenario tested	Absolute Fitness	Surface time (h) ¹	DVM amplitude (m) ¹	Food-limitation (dim. less) ¹	Development Time (d) ²	Time of seasonal descent	Time of seasonal ascent	Size at seasonal descent ($\mu\text{g C}$)	Size at sexual maturity ($\mu\text{g C}$)	Age at sexual maturity (d)	Onset of spawning (d)	Birth time	Fecundity	Breeding mode index (dim. less)
CF	<i>BR</i>	12227.31	19.63	16.12	0.96	44.13	Jul.30	Jan.24	258.31	271.65	244	Feb.16	Jun.16	12628	0.00
	<i>F+</i>	12870.88	19.40	16.36	0.97	40.19	Jul.28	Jan.29	274.12	276.94	242	Feb.15	Jun.18	13111	0.00
	<i>F-</i>	11253.92	19.77	16.02	0.94	63.16	Aug.03	Jan.14	249.21	264.43	261	Feb.15	Jun.01	11621	0.00
	<i>T+</i>	14739.06	19.58	16.09	0.97	32.60	Jul.20	Jan.31	260.55	262.21	242	Feb.15	Jun.18	15239	0.00
	<i>T-</i>	10337.61	19.56	16.22	0.95	69.75	Aug.13	Jan.14	272.41	283.22	254	Feb.14	Jun.04	10686	0.00
CG	<i>BR</i>	12100.56	18.53	21.87	0.95	76.21	Aug.10	Feb.16	554.38	554.38	266	Feb.16	May.26	13031	0.07
	<i>F+</i>	12941.16	18.46	21.89	0.97	80.17	Aug.12	Feb.15	563.85	563.85	267	Feb.15	May.24	13947	0.07
	<i>F-</i>	10838.82	18.82	21.23	0.94	85.08	Aug.05	Feb.16	537.29	537.29	278	Feb.16	May.12	11649	0.06
	<i>T+</i>	13732.62	18.01	22.47	0.96	59.83	Aug.09	Feb.14	530.66	530.66	248	Feb.14	Jun.11	14727	0.05
	<i>T-</i>	10797.90	18.67	21.64	0.95	100.91	Aug.07	Feb.14	579.93	579.93	291	Feb.14	Apr.29	11683	0.08
CH	<i>BR</i>	1450.29	18.05	29.51	0.93	135.46	Aug.28	Feb.16	1918.04	1918.04	308	Feb.16	Apr.15	1690	1.00
	<i>F+</i>	1479.60	17.80	30.13	0.95	133.01	Aug.24	Feb.16	1957.79	1957.79	309	Feb.16	Apr.13	1721	1.00
	<i>F-</i>	1376.78	18.19	28.69	0.93	141.62	Aug.21	Feb.16	1860.71	1860.71	319	Feb.16	Apr.02	1609	1.00
	<i>T+</i>	1366.20	17.77	30.12	0.94	96.25	Aug.25	Feb.17	1852.35	1852.35	272	Feb.17	May.21	1583	1.00
	<i>T-</i>	1516.58	18.21	28.92	0.93	152.38	Aug.23	Feb.14	1994.31	1994.31	328	Feb.14	Mar.24	1781	1.00

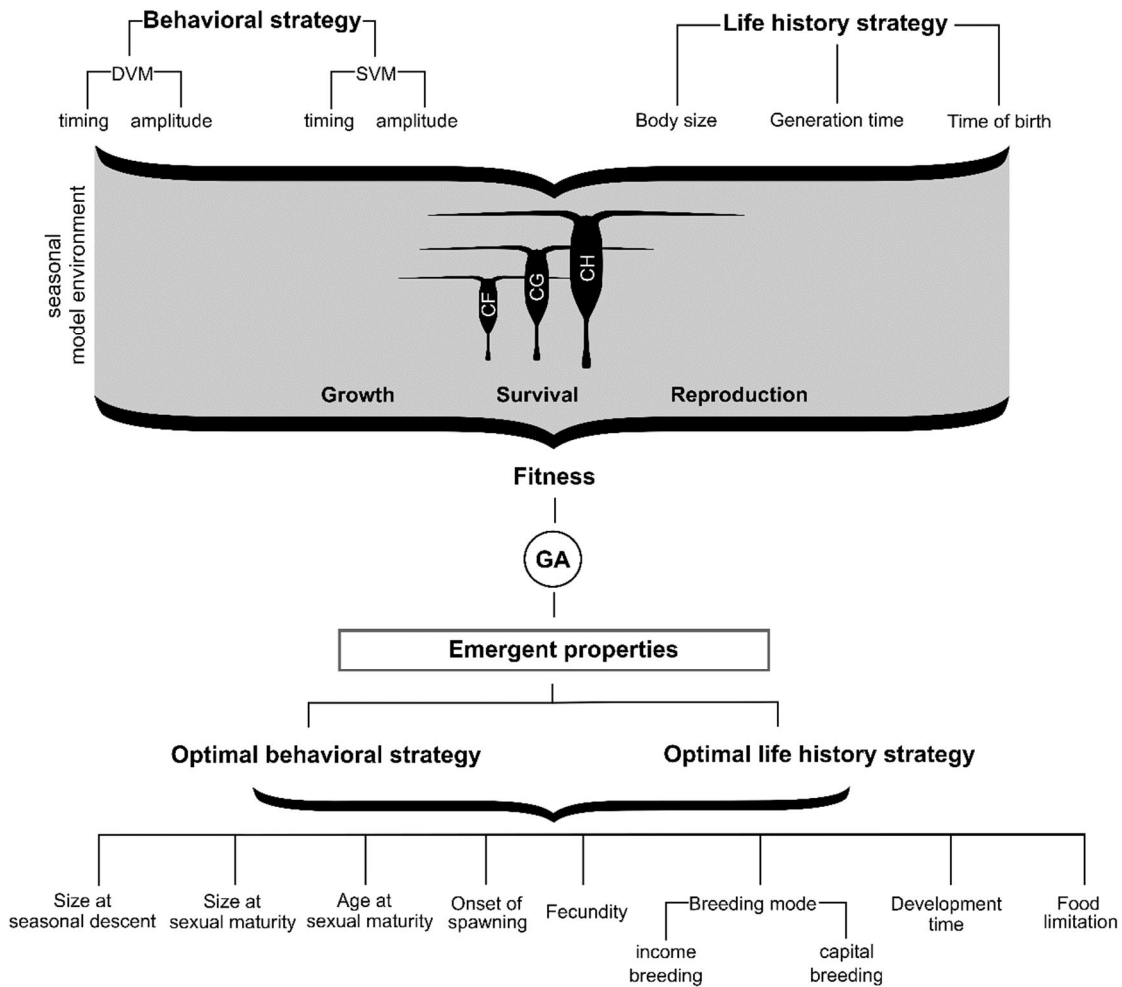
¹Mean for all developmental stages, ²From egg to adult excluding the overwintering duration



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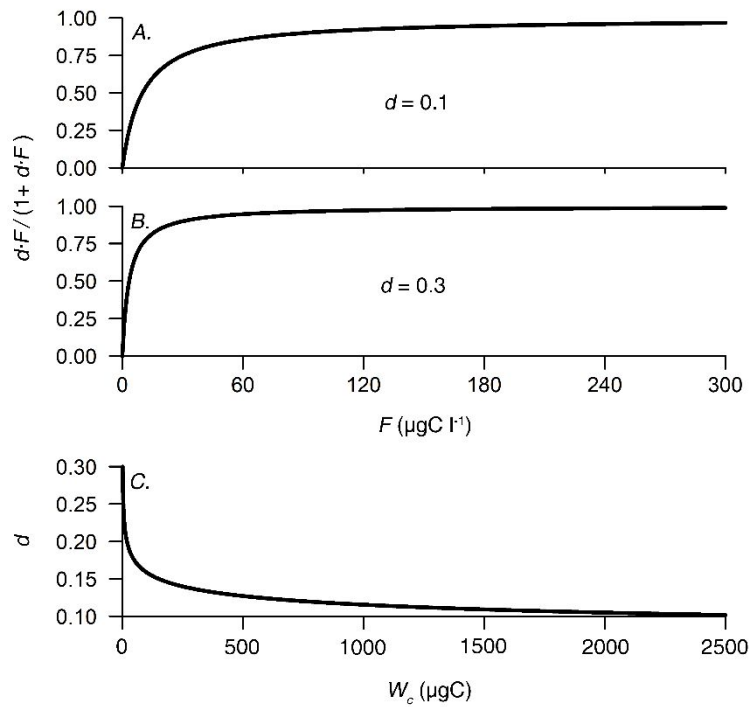
932 **Fig.1** Dynamics of modeled sea-surface irradiance (A, D, G), temperature (B, E, H) and
 933 food availability (C, F, I, as Chlorophyll-*a* biomass) in the three model environments. The
 934 bottom depth is 1000 m, but the ordinates of lower panels are cropped due to the
 935 vertical homogeneity of those parameters. See Appendix A1 for a detailed comparison.

936

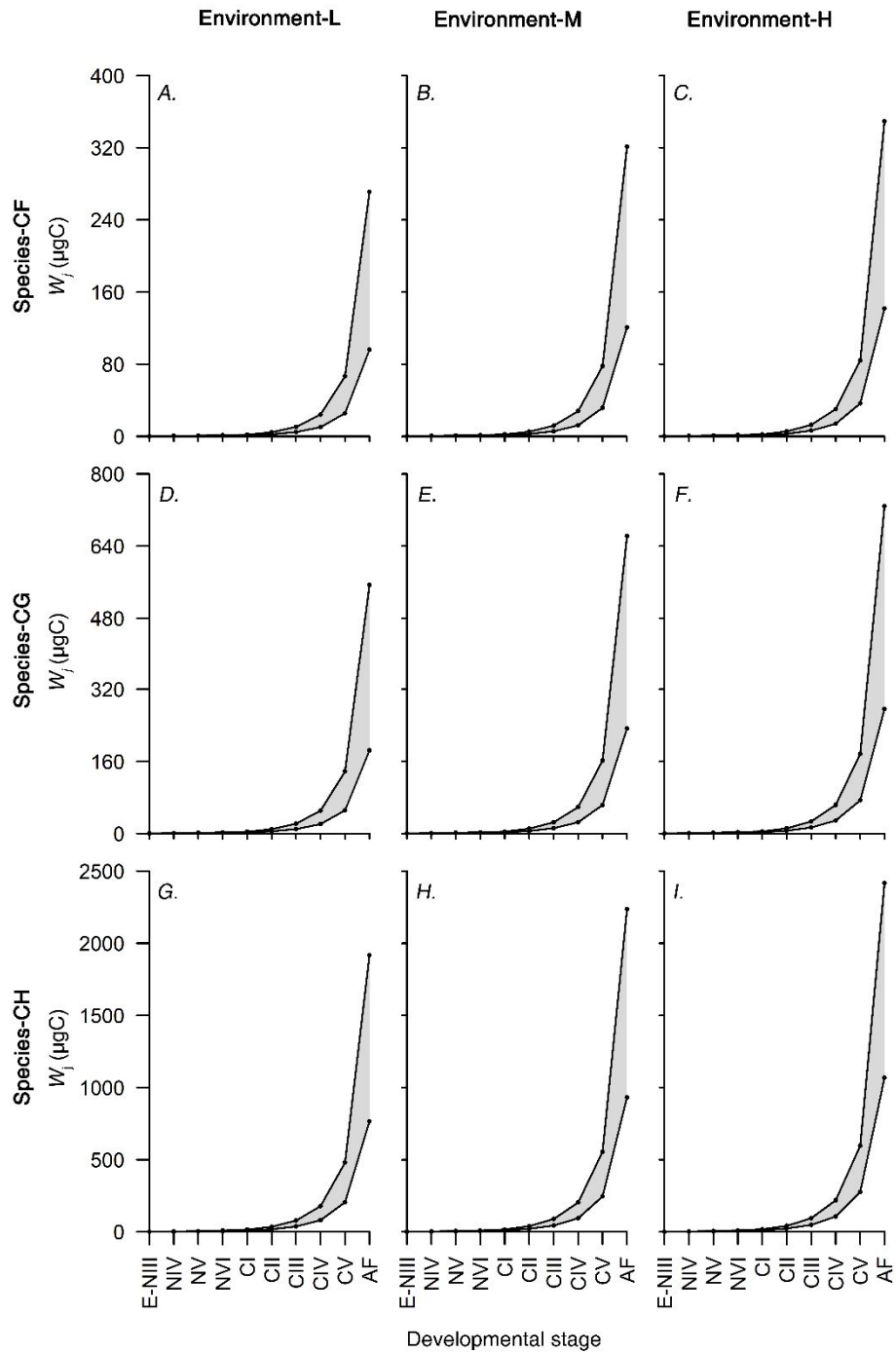


937 **Fig. 2** The model overview. The behavioral strategy (vertical strategy) and some key
 938 aspects of the life history strategy are defined by seven evolvable free parameters (cf.
 939 Table 2). These are hardwired to model copepods of three different species representing
 940 *C. finmarchicus*, *C. glacialis* and *C. hyperboreus*. For a given model environment, a
 941 Genetic Algorithm heuristically finds the optimal combination for these parameters
 942 through a fitness function of growth, survival and reproduction, and predicts the
 943 environment- and species-specific optimal behavioral and life history strategies.

944

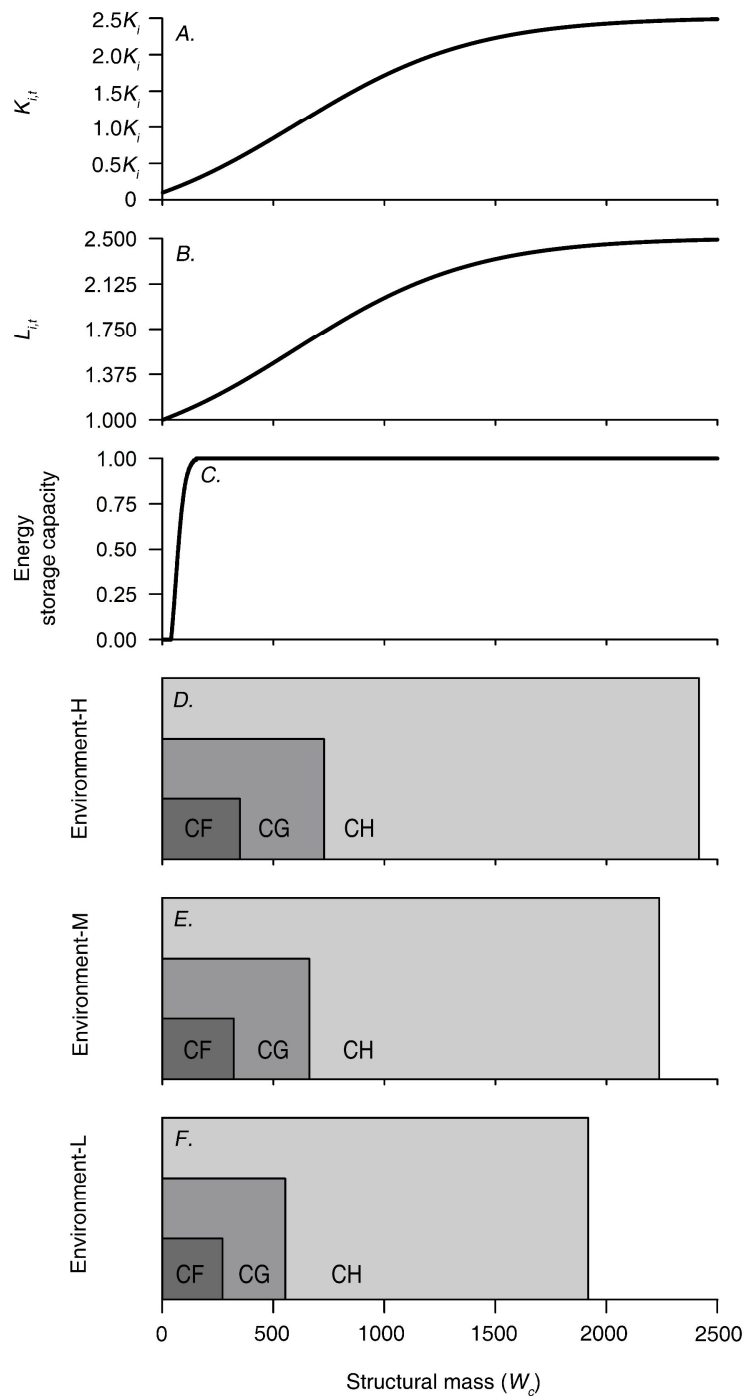


945 **Fig. 3** (A–B) The shape of the Holling’s type-II (disk) function (Eq. 4) at the higher (0.3)
 946 and lower (0.1) ends of parameter d , which describes the dependency of ingestion rate
 947 on the ambient food concentration. (C) The power function (Eq. 5) through which
 948 parameter d relates with the structural mass (W_c). In panels A and B, the y-intercept of
 949 the horizontal asymptote is ca. $75 \mu\text{gC l}^{-1}$ for $d = 0.3$ and $125 \mu\text{gC l}^{-1}$ for $d = 0.1$. This is
 950 the satiation food concentration, above which ingestion rate becomes solely
 951 temperature-dependent (Eq. 3).

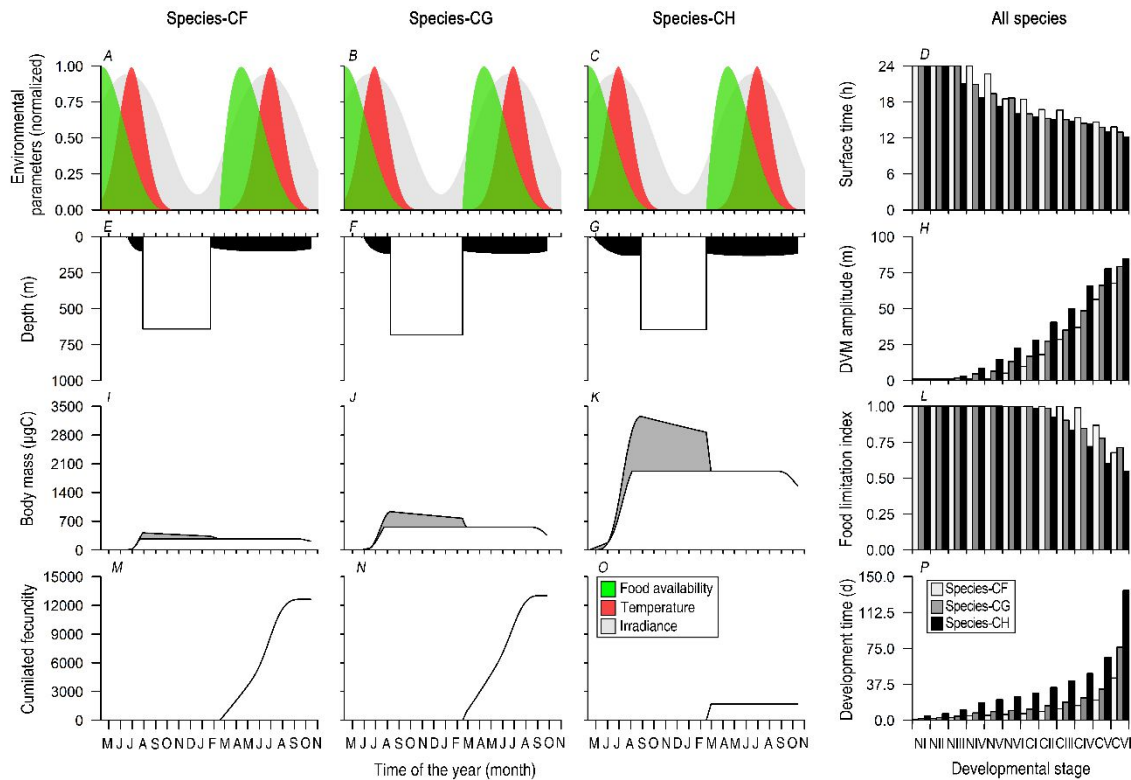


952 **Fig. 4** The species-specific maximum (W_j^{max} : upper line and points in each panel) and
 953 minimum (W_j^{min} : lower line and points in each panel) critical molting masses estimated
 954 for each model environment. Based on the value of the evolvable body size parameter
 955 (α), stage-specific critical molting masses for a given copepod occupies a fixed fraction
 956 between the minima and maxima, i.e. within the shaded area. See Appendix A2 for
 957 tabulated values.

958

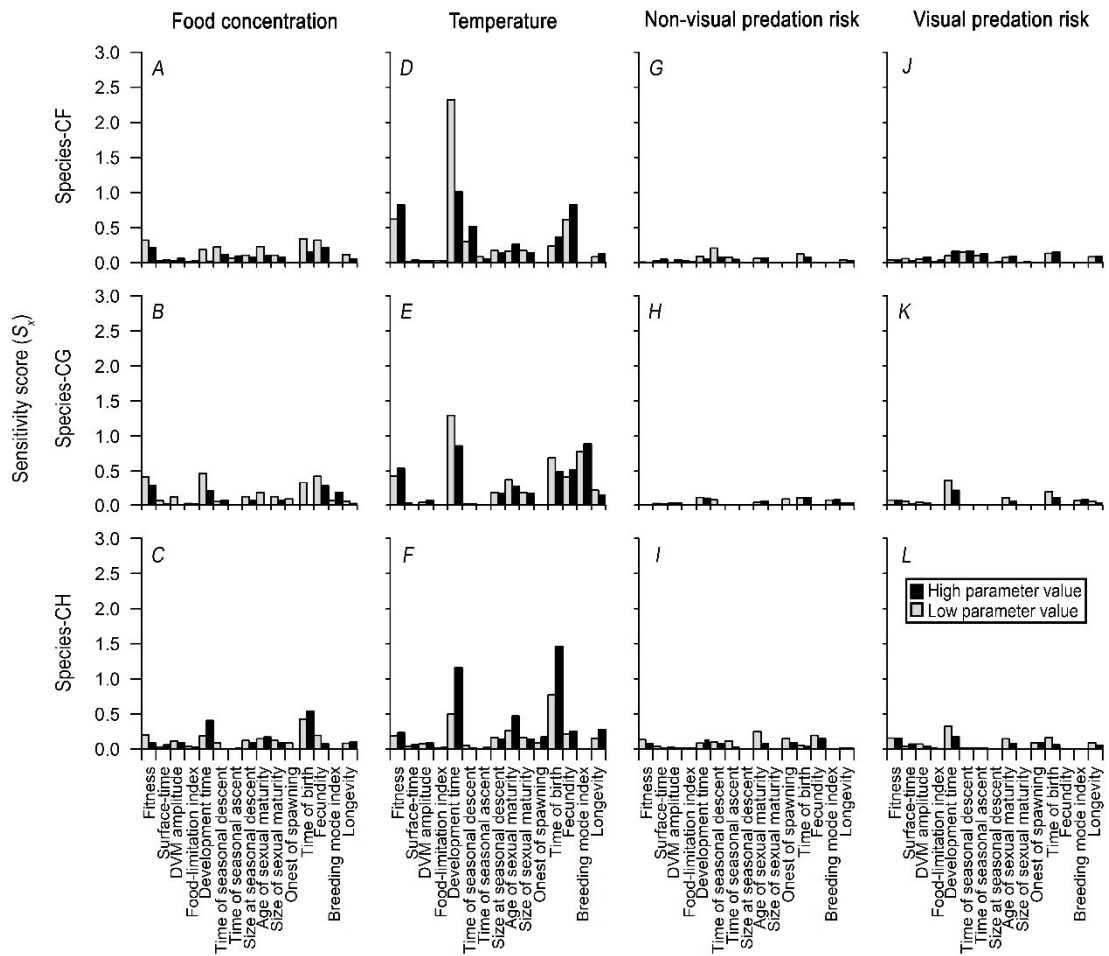


959 **Fig. 5** Size-dependent relationships of (A) visual predation risk scalar K , (B) irradiance
 960 sensitivity parameter L , and (C) the energy storage capacity (Eq. 13). Panels D–F provide
 961 a rough reference to how these size-specific patterns can influence species-specific
 962 processes in different model environments. Vertical axes of D–F have no dimensions,
 963 but the height of the boxes is to show the overlap of inter-specific size ranges.



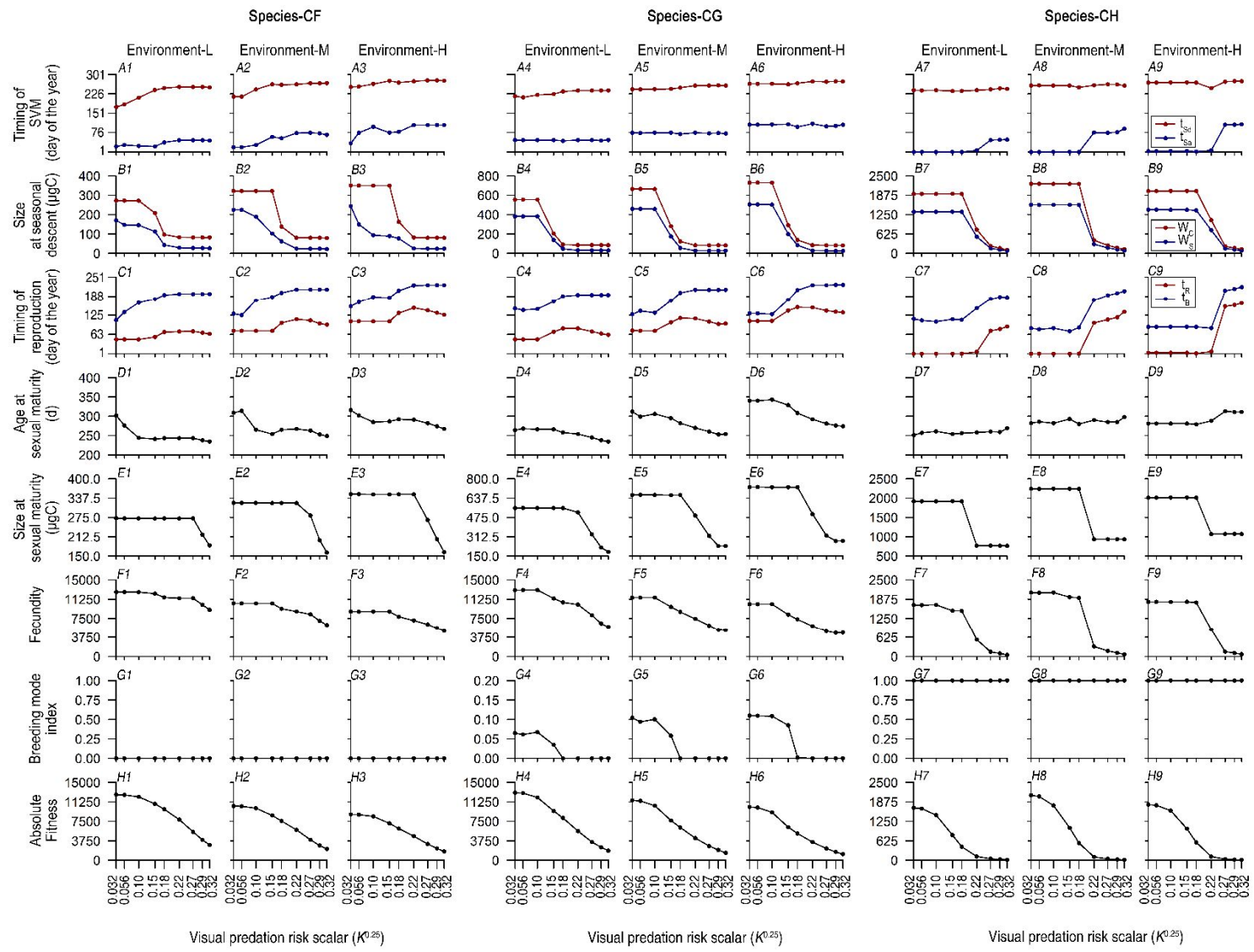
964 **Fig. 6** Some behavioral and life history traits/attributes of the three model-species
 965 traced in the basic run at Environment-L (A–C). To the left are predicted lifetime
 966 variability of the vertical trajectories (E–G), structural and energetic reserve masses (I–
 967 K) and fecundity (M–O). To the right are stage-specific attributes (D, H, L and P). Shaded
 968 regions of panels I–K represent the mass of energy reserve.

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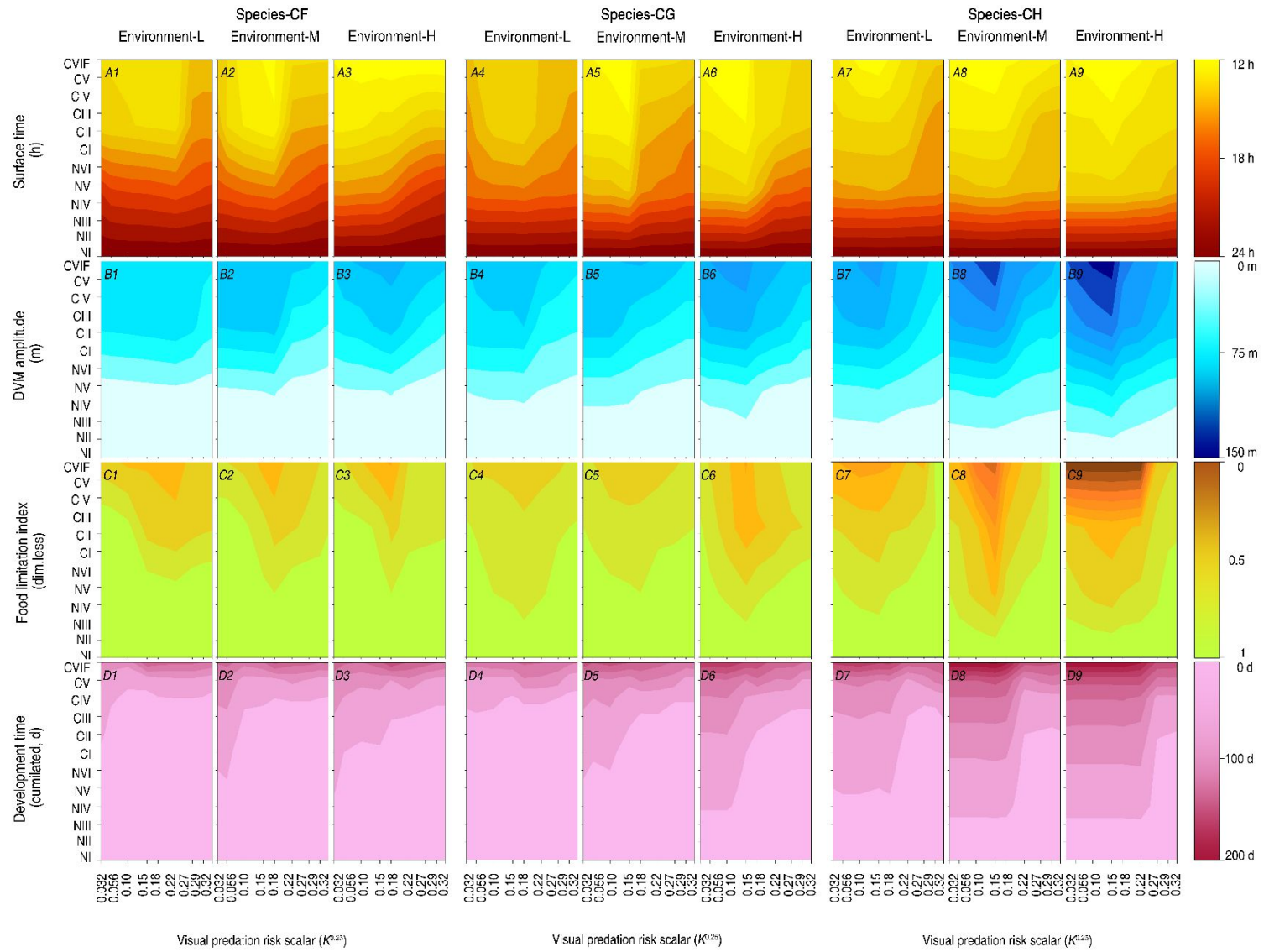
970 **Fig. 7** Graphical summary of the sensitivity analysis. Plotted as bars are absolute
 971 sensitivity scores, the height of which indicating the degree of sensitivity. Further details
 972 are provided in Table 6.

973



974 **Fig. 8** Predicted variability of species-specific life history traits along the modelled latitudinal gradient under variable levels of visual
975 predation risk. t_{sd} : time of seasonal descent, t_{sa} : time of seasonal ascent, W_c : structural mass, W_s : mass of energy reserve, t_B : optimal birth
976 time, t_R : time of first reproduction

977



978 **Fig. 9** Predicted stage-specific variability of (A1–A9) surface time, (B1–B9) DVM amplitude, (C1–C9) food limitation index and (D1–D9) the
979 development times (excluding overwintering duration) of each model-species along the modelled latitudinal gradient at variable levels of
980 visual predation risk. See table 3 for descriptions of the above variables.

981

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1 **APPENDIX**

2

3 **Artificial evolution of behavioral and life history strategies of Arctic *Calanus* spp. in**
4 **response to bottom-up and top-down selection pressures**

5

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14 **A1: Summary environmental parameters of the model environments**

15 The modeled irradiance was estimated following the global clear-sky horizontal
 16 irradiance model of [Robledo and Soler \(2000\)](#). A comprehensive account of the
 17 irradiance submodel is provided in [Bandara et al. \(2018\)](#). Estimated irradiance over the
 18 modeled environments roughly agree with the field estimates. Field estimates of
 19 temperature were adopted from [Swift \(1986\)](#), [Ingvaldsen and Loeng \(2009\)](#), [Daase and](#)
 20 [Eiane \(2007\)](#), [Daase et al. \(2013\)](#), [Bandara \(2014\)](#), [Bandara et al. \(2016\)](#). Further,
 21 temperature and Chlorophyll-a biomass data collected during the UNIS AB820 (2012-
 22 2016) cruise from Van Mijenfjorden, Isfjorden, Billefjorden, Kongsfjorden, and offshore
 23 stations around 78–81°N were used via Paul E. Renaud (course coordinator 2016). Year-
 24 round field data (temperature and Chlorophyll-a biomass observations) from Lofoten
 25 and Vesterålen regions were also obtained from mooring data via Boris Espinasse
 26 (<http://love.arctosresearch.net>). Finally, temperature data from southern and
 27 southeastern Norwegian fjords (60–70°N) were also obtained following
 28 communications with Slawek Kwasniwski. These data were considered when deciding
 29 the seasonal maxima, minima of temperature and maximum Chlorophyll-a biomass
 30 parameterizations.

31

32 **Table A1:** Comparison between model environments. Cf. Fig. 1 in main text

Parameter	Attribute	Env-L	Env-M	Env-H
Irradiance ($\mu\text{mol m}^{-2} \text{s}^{-1}$)	Min.	0	0	0
	Max.	1500	1200	800
	Time of Max.	day 172 (June 21)	day 172 (June 21)	day 172 (June 21)
Temperature (°C)	Min.	2	0	-1.5
	Max. (°C)	15	12	10
	Time of Max.	day 181 (July 1)	day 203 (July 21)	day 212 (Aug 1)
Food availability (mg m^{-3} Chl.a)	Min.	0	0	0
	Max.	6	6	6
	Time of Max.	day 105 (April 15)	day 135 (May 15)	day 165 (Jun 15)
	Productive season (duration)	229 d	208 d	180 d

33

34

35 **A2: Growth and development submodel**

36 [Maps et al. \(2011\)](#) have formulated a mechanistic model to describe growth and
37 development of several high-latitude calanoid copepod species. Their predictions
38 include *Calanus finmarchicus*, *C. glacialis* and *C. hyperboreus*. At constant food
39 concentration and constant temperatures, this growth model performs well. However,
40 in their model, growth ($\mu\text{g C}$) emerges as a function of development.

41 In this model, when temperature and food concentration varies over time, the
42 development times tend to shift. For example, a copepod performing DVM would
43 encounter variable temperature and food concentrations on daily if not hourly basis.
44 This variability of development times causes large amounts of Carbon to be assimilated
45 between the development stage j and $j + 1$. We observed that copepods performing
46 DVM attaining unrealistic structural masses as a result (e.g. females with structural
47 masses ca. $4 \times 10^4 \mu\text{g C}$ at -1 – 10°C and 0 – $180 \mu\text{g C l}^{-1}$). Due to this limitation, we could
48 not implement [Maps et al. \(2011\)](#) growth model as the growth submodel in our work.

49 However, given the usefulness of the above model at constant temperature and food
50 concentrations, we used it to parameterize a simple growth model that we formulated
51 (Eqs. 1–8 and 14–15). The temperature and mass coefficients and exponents of our
52 model were estimated from the temperature and mass specific growth predictions (at
53 satiation food concentrations) simulated by [Maps et al. \(2011\)](#)'s model. Predicted
54 development times of the above model and those predicted by ours at constant
55 temperatures and food concentrations are identical. However, at variable temperatures
56 and food concentrations (such experienced by diel migrating copepods) our model
57 produces more meaningful estimates, as development is a function of growth (the
58 concept of critical molting masses, e.g. [Fiksen & Giske 1995](#), [Fiksen & Carlotti 1998](#)). The
59 only down side to this is that we had to adopt a new evolvable parameter to describe
60 the body mass trajectory.

61 The critical molting masses (W_j) were calculated from running the [Maps et al. \(2011\)](#) at
62 minimum and maximum environmental specific temperatures (see Table A1 above)
63 under the minimum and maximum food concentration. Here, the W_j^{max} is given by
64 running the above model at minimum temperature at maximum food concentration.
65 W_j^{min} was extracted in vice versa scenario (Table A2).

66

67 **Table A2:** Minimum and maximum stage-specific critical molting masses ($\mu\text{g C}$) for each model-species for the three environments.
 68 Estimates are derived following the method described above. These are presented in Fig. 4 in the main text. Egg to NIII mass was assumed
 69 to remain constant during the development.

	Species-CF						Species-CG						Species-CH					
	Env.-L		Env.-M		Env.-H		Env.-L		Env.-M		Env.-H		Env.-L		Env.-M		Env.-H	
	W_j^{min}	W_j^{max}	W_j^{min}	W_j^{max}	W_j^{min}	W_j^{max}	W_j^{min}	W_j^{max}	W_j^{min}	W_j^{max}	W_j^{min}	W_j^{max}	W_j^{min}	W_j^{max}	W_j^{min}	W_j^{max}	W_j^{min}	W_j^{max}
E-NIII	0.23	0.23	0.23	0.23	0.23	0.23	0.40	0.40	0.40	0.40	0.40	0.40	0.56	0.56	0.56	0.56	0.56	0.56
NIV	0.44	0.59	0.47	0.62	0.49	0.63	0.90	1.15	0.95	1.21	0.98	1.23	1.94	2.70	2.07	2.86	2.17	2.96
NV	0.61	0.91	0.66	0.97	0.70	1.01	1.31	1.87	1.42	1.98	1.49	2.04	3.36	5.20	3.68	5.63	3.89	5.86
NVI	0.80	1.31	0.89	1.43	0.95	1.49	1.83	2.80	2.01	3.00	2.14	3.11	5.30	8.85	5.89	9.70	6.31	10.17
CI	1.08	1.94	1.23	2.15	1.34	2.26	2.62	4.30	2.92	4.65	3.14	4.87	8.55	15.29	9.62	16.95	10.42	17.90
CII	2.31	4.61	2.66	5.20	2.94	5.51	5.18	9.99	5.97	11.10	6.58	11.77	17.98	35.19	20.72	39.55	22.79	42.09
CIII	4.78	10.49	5.62	12.00	6.35	12.82	10.27	22.37	12.09	25.35	13.66	27.21	37.44	78.37	43.85	88.99	48.54	95.13
CIV	10.18	24.27	12.27	28.10	13.96	30.20	21.11	51.05	25.50	58.91	29.26	63.77	80.48	178.08	95.12	204.15	106.75	219.04
CV	25.68	66.59	31.63	77.94	36.58	84.36	51.54	138.08	63.56	162.21	74.01	177.05	204.98	480.14	244.94	554.96	277.95	597.47
Adult	96.16	271.08	120.91	321.26	141.84	349.34	184.57	553.57	233.59	662.52	276.49	728.69	766.22	1917.90	932.82	2237.50	1068.91	2416.83

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